# Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2017-2018 

Steven P. Richards<br>Washington Department of Fish \& Wildlife<br>And<br>Todd N. Pearsons<br>Public Utility District Number 2 of Grant County, Washington

October 22, 2018

## Disclaimer

This report is provided as an annual data update of the Public Utility District No. 2 of Grant County, Washington's (Grant PUD's) monitoring and evaluation plan for Priest Rapids Hatchery. All data are provisional and subject to change as new data and analyses become available. Readers are cautioned to use data at their own risk and should consult the most current report to obtain the most current and accurate information. Data sets will become final when they are published in peer reviewed scientific journals.

This report should be cited as:
Richards, S.P. and T.N. Pearsons. 2018. Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2017-2018. Public Utility District No. 2 of Grant County, Ephrata, Washington.

## Executive Summary

This report is the eighth annual report dedicated to monitoring and evaluating the Priest Rapids Hatchery (PRH) production of fall Chinook salmon. The PRH is located below Priest Rapids Dam adjacent to the Columbia River and has been in operation since 1963. The monitoring and evaluation program associated with PRH is intended to evaluate the performance of the program in meeting hatchery and natural production goals. This report is intended to be cumulative, but also focus attention on the most recent year of data collection and production (2017-2018).

The PRH was originally built to mitigate for the construction and operation of Priest Rapids and Wanapum dams. The hatchery is operated as an integrated program for the purpose of increasing harvest while limiting undesirable risks to the naturally spawning population. The hatchery produces 5.6 million subyearling fall Chinook salmon for Public Utility District No. 2 of Grant County, Washington’s (GPUD) mitigation requirement and 1.7 million subyearling fall Chinook salmon under contract with the United States Army Corps of Engineers for mitigation for the construction and operation of John Day Dam. These fish contribute significantly to a variety of fisheries, such as fisheries off the coasts of Alaska and Canada and fisheries in the Columbia River.

The estimated total escapement of fall Chinook salmon to the Hanford Reach in 2017 was 73,759 fish. This is similar to the mean abundances of the past few decades. The mean and median escapement for 1991 through 2017 was 75,146 and 57,710 fish, respectively.
The 2017 returns to PRH volunteer trap totaled 17,812 fall Chinook salmon. A total of 4,511 fish that returned to the volunteer trap at PRH were ponded at the hatchery for broodstock. An additional 348 fish from the Angler Broodstock Collection (ABC) fishery and 809 fish from Priest Rapids Dam Off-Ladder-Adult-Fish-Trap (OLAFT) were included in the broodstock in an effort to increase the number of natural-origin broodstock. In total, 5,668 fish were spawned to meet egg-take goals for multiple hatchery programs. The majority of the fish that were surplus to broodstock needs were provided to food-banks and tribes for consumption in recent years.

There were a number of similarities and differences of hatchery and natural origin fall Chinook salmon. The hatchery origin fish appeared to return at a younger age than natural origin fish. It appears that age- 2 and 3 hatchery origin fish tend to be larger than natural origin fish of the same age. Likewise, age-4 and 5 natural origin fish tend to be larger than their hatchery origin counterparts. The number of eggs, egg size, and egg mass produced by hatchery and natural origin females of similar length was similar. With the exception of one year (2013), egg retention in female carcasses in the Hanford Reach has been low.

Hatchery origin fish released from PRH spawned throughout the Hanford Reach. In addition, the hatchery origin proportions of spawners relative to total spawners in the different sections of the Hanford Reach were similar. Recent evidence suggested that adult carcasses drift downstream of their spawning location and bias the estimated spawning distribution downstream. Stray rates into other populations appeared to be low based upon coded-wire tag recoveries and PIT tag detections of PRH adults in the Snake River were also low. However, there have been notable numbers of PIT tag detections of PRH adults above Priest Rapids Dam.

The PRH continued to contribute substantially to ocean and Columbia River fisheries and tends to have higher adult recruitment rates than the natural spawning fall Chinook salmon to the

Hanford Reach of the Columbia River. Adult recruitment rate of brood year 2011 for PRH was 32.03 versus 4.93 for fish spawning in the Hanford Reach.

PRH origin fish were estimated to make up $6.5 \%$ of the natural spawning population in the Hanford Reach during 2017. All hatchery fish combined (including fish released from Ringold Hatchery and strays from outside the Hanford Reach) comprised 8.3\% of the fall Chinook salmon on the spawning grounds. Otolith recoveries at the PRH volunteer trap indicated that a very high percentage of fish returning to the PRH were of PRH origin. The proportion of natural influence (PNI) for Hanford Reach fall Chinook salmon including all hatcheries was 0.835 in 2017. This value was calculated using a gene flow model based on the Ford model and exceeded the PNI target of 0.67 for the fourth consecutive year. Adult management of fish at the PRH volunteer trap and alternative broodstock collection techniques to increase natural origin fish in the broodstock have contributed to improvements in PNI for the PRH program.

## Table of Contents

1.0 Introduction ..... 1
2.0 Objectives ..... 3
3.0 Project Coordination ..... 3
4.0 Life History - Hanford Reach Fall Chinook Salmon ..... 4
5.0 Sample Size Considerations ..... 4
6.0 Current Operation at Priest Rapids Hatchery. ..... 5
7.0 Origin of Adult Returns to Priest Rapids Hatchery ..... 8
7.1 Origin Based on Hatchery Marks ..... 9
7.2 Origin Based on Coded-Wire Tag Recoveries ..... 10
8.0 Broodstock Collection and Sampling ..... 11
8.1 Broodstock Age Composition ..... 11
8.2 Length by Age Class of Broodstock ..... 14
8.3 Gender Ratios ..... 15
8.4 Fecundity ..... 16
9.0 Hatchery Rearing ..... 21
9.1 Number of Eggs Taken ..... 21
9.2 Number of Acclimation Days ..... 21
9.3 Annual Releases, Tagging, and Marking ..... 22
9.4 Fish Size and Condition of Release ..... 25
9.5 Survival Estimates ..... 26
9.6 Juvenile PIT Tag Detections at the Priest Rapids Hatchery Array ..... 27
10.0 Adult Fish Pathogen Monitoring ..... 30
10.1 Juvenile Fish Health Inspections ..... 32
11.0 Redd Survey ..... 33
11.1 Hanford Reach Aerial Redd Counts ..... 34
11.2 Redd Distribution ..... 34
11.3 Spawn Timing ..... 36
11.4 Escapement ..... 36
11.5 Hatchery Discharge Channel Redd Counts. ..... 37
12.0 Carcass Surveys ..... 37
12.1 Hanford Reach Carcass Survey: Section 1 - 5 ..... 39
12.2 Proportion of Escapement Sampled: Section 1-5 ..... 39

[^0]12.3 Carcass Distribution and Origin. ..... 40
12.4 Priest Rapids Dam Pool Carcass Survey: Section 6 ..... 43
12.5 Number sampled: Section 6 ..... 43
12.5.1 Proportion of Escapement Sampled: Section 6. ..... 43
12.5.2 Carcass Origin: Section 6 ..... 44
12.6 Hatchery Discharge Channel: Section 7 and 8 Carcass Survey. ..... 45
12.7 Number sampled: Section 7 and 8 ..... 45
12.7.1 Proportion of Escapement Sampled: Section 7 and 8 ..... 45
12.7.2 Carcass Distribution and Origin: Section 7 and 8. ..... 46
13.0 Life History Monitoring ..... 47
13.1 Migration Timing ..... 47
13.2 Age at Maturity. ..... 48
13.3 Size at Maturity ..... 49
13.4 Gender Composition for Adult Escapement ..... 55
13.5 Egg Retention ..... 57
14.0 Contribution to Fisheries. ..... 59
15.0 Straying ..... 62
15.1 Genetics ..... 65
15.2 Proportion of Natural Influence ..... 65
15.3 Estimate of pNOB ..... 66
15.4 Estimates of pHOS ..... 67
15.5 Estimates of PNI ..... 68
16.0 Natural and Hatchery Replacement Rates ..... 70
17.0 Smolt-to-Adult Survivals ..... 71
18.0 ESA/HCP Compliance ..... 73
18.1 Broodstock Collection ..... 73
18.2 Hatchery Rearing and Release ..... 74
18.3 Distribution of Surpluses, Mortalities, and Spawned, Adult fall Chinook Salmon from Priest Rapids Hatchery ..... 74
18.4 Hatchery Effluent Monitoring. ..... 75
18.5 Ecological Risk Assessment ..... 75
19.0 Acknowledgments ..... 75
Literature Cited ..... 76
List of Figures
Figure 1 Location of Priest Rapids and Ringold Spring hatcheries and the Hanford Reach (indicated by stars) ..... 2
Figure 2 Priest Rapids Hatchery facility and Priest Rapids Dam Off-Ladder Adult Fish Trap (OLAFT) ..... 2
Figure 3 Weekly counts of fish adult Chinook salmon collected at the Priest Rapids Hatchery Volunteer Trap, 2017. ..... 7
Figure 4 Linear relationship between fecundity and fork length for combined samples of natural and hatchery origin fall Chinook salmon spawned at Priest Rapids Hatchery, Return Years 2010-2017. ..... 18
Figure $5 \quad$ Fecundity versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 2013-2017. ..... 19
Figure 6 Mean egg weight versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 2013-2017. ..... 20
Figure $7 \quad$ Total egg mass weight versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 2013-2017. ..... 20
Figure 8 Distribution of fall Chinook salmon redd counts by location for the 2017 aerial surveys in the Hanford Reach, Columbia River. (Data provided by Mission Support Alliance). ..... 35
Figure 9 Location of aerial redd index areas (green area numbers) and river boat carcass survey sections in the Hanford Reach. ..... 38

## List of Tables

Table 1 Source and disposition of Chinook salmon collected for broodstock at Priest Rapids Hatchery, Return Year 2017 ..... 6
Table 2 Total fish handled, numbers sampled, and estimates of hatchery and natural originChinook salmon collected at Priest Rapids Hatchery, Priest Rapids Dam Off-Ladder Adult Fish Trap, and Angler Broodstock Collection fishery. Origindetermined by otolith thermal marks, presence of coded-wire tags, and/or adiposeclips, Return Years 2013-2017.10
Table 3 Estimated proportion of hatchery and natural origin adult Chinook salmonreturning to the Priest Rapids Hatchery volunteer based on coded-wire tagexpansion. The entire collection was sampled for coded-wire tags, Return Years2005-2017.11
Table 4 Age composition for hatchery and natural origin fall Chinook salmon spawned at Priest Rapids Hatchery (includes all sources of broodstock), Return Years 2007- 2017. Proportions calculated from expanded age compositions by origin for each source of broodstock to account for differing sample rates. ..... 12
Table 5 Age composition for hatchery and natural origin fall Chinook broodstock collected from the Priest Rapids Hatchery volunteer trap, Return Years 2012- 2017 ..... 13
Table 6 Age composition for hatchery and natural origin fall Chinook salmon broodstock collected from the Off Ladder Adult Fish Trap at Priest Rapids Dam, Return Years 2012-2017. ..... 13
Table 7 Age composition for hatchery and natural origin fall Chinook salmon broodstock collected from Angler Broodstock Collection, Return Years 2012-2017. ..... 14
Table 8 Mean fork length (cm) at age (total age) of fall Chinook salmon sampled from each source of broodstock spawned at Priest Rapids Hatchery, Return Year 2017. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation. ..... 14
Table $9 \quad$ Mean fork length (cm) at age (total age) of hatchery and natural origin fall Chinook salmon collected from broodstock originating from the Priest Rapids Hatchery volunteer trap. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation, Return Years 2012-2017. ..... 15
Table 10 Number of male and female hatchery fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 2001-2017. Ratios of males to females are also provided. ..... 16
Table 11 Mean fecundity of fall Chinook salmon collected for broodstock at Priest Rapids Hatchery, Return Years 2001-2017. ..... 17
Table 12 Mean fecundity at age for fall Chinook salmon sampled at the Priest Rapids Hatchery, Return Years 2010-2017. N = sample size and SD = 1 standard deviation. ..... 19
Table 13 Number of eggs taken from fall Chinook salmon broodstock collected at Priest Rapids Hatchery, Return Years 1984-2017. ..... 21
Table 14 Number of days fall Chinook salmon fry were reared at Priest Rapids Hatchery prior to release, Brood Year 2017. ..... 22
Table 15 Number of marked, unmarked, and tagged fall Chinook salmon smolts released from Priest Rapids Hatchery, Brood Years 1977-2017. ..... 24
Table 16 Mean length (FL, mm), weight (g and fish/pound), and coefficient of variations (CV) of fall Chinook smolts released from Priest Rapids Hatchery, Brood Years 1991-2016. ..... 26
Table 17 Hatchery life stage survival ( $\mathrm{P}^{\wedge}$ ) for fall Chinook salmon at Priest Rapids Hatchery, Brood Years 1989-2017 ..... 27
Table 18 Number of sub-yearlings PIT tagged, mark, and release dates, and the number of unique tags detected at the array in the Priest Rapids discharge channel, Brood Years 2011-2017. ..... 29
Table 19 Viral inspections of fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 1991-2017. ..... 31
Table 20 ELISA test results to determine risk of bacterial kidney disease of adult female fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 2008- 2017 ..... 32
Table 21 Juvenile fish health inspections for Priest Rapids Hatchery fall Chinook salmon, Brood Years 2006-2017 ..... 33
Table 22 Summary of fall Chinook salmon peak redd counts for the 1948-2017 aerial surveys in the Hanford Reach, Columbia River. ..... 34
Table 23 Number of all Chinook salmon redds counted in difference reaches on the Hanford Reach area of the Columbia River during October 2017 through November 2017 aerial redd counts. (Data provided by Mission Support Alliance). ..... 35
Table 24 Calculation of escapement estimates for fall Chinook salmon in the Hanford Reach, Columbia River 2017. ..... 36
Table 25 Escapement for fall Chinook salmon in the Hanford Reach, Return Years 1991- 2017. ..... 37
Table 26 Numbers and proportions of total escapement of fall Chinook salmon carcasses surveyed for coded-wire tags within each survey section on the Hanford Reach, Return Years, 2010-2017. ..... 39
Table 27 Number of carcass surveys conducted by section in the Hanford Reach, Return Years 2010-2017. ..... 39
Table 28 Number of redds and carcasses, total spawning escapement, and proportion of escapement sampled for fall Chinook salmon in Section 1 through 5 of the Hanford Reach, Return Year 2017. ..... 40
Table 29 Numbers of natural and hatchery origin fall Chinook salmon carcasses sampled within Section 1 through 5 of Hanford Reach based on expansions of coded-wire tag recoveries, Return Years 2010-2017 ..... 41
Table 30 Origin of Chinook salmon carcasses recovered in the Hanford Reach by section based on recoveries of marked and unmarked carcasses within the biological sample, Return Years 2012-2017. ..... 42
Table $31 \quad$ Number of fall Chinook salmon carcasses sampled within Section 6 (Priest Rapids Dam Pool), Return Years 2010-2017. ..... 43
Table 32 Carcasses sampled, total spawning escapement and proportion of escapement for fall Chinook salmon in Section 6 (Priest Rapids Dam Pool), Return Years 2010- 2017. ..... 43
Table 33 Origin of fall Chinook salmon spawning in Section 6 (Priest Rapids Dam Pool), Return Years 2012-2017. ..... 44
Table 34 The number of fall Chinook salmon carcass surveys within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at the confluence of the hatchery discharge channel), Return Years 2010-2017. ..... 45
Table 35 Number of carcasses sampled, total spawning escapement and proportion of escapement sampled for fall Chinook salmon within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at confluence of the hatchery discharge channel), Return Year 2017. ..... 45
Table 36 The origin of Chinook salmon carcasses recovered within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at the confluence of the hatchery discharge channel), Return Years 2012-2017. ..... 46
Table 37 The week that $10 \%$, $50 \%$ (median), and $90 \%$ of the natural and hatchery origin fall Chinook salmon passed Bonneville Dam, 2010-2017. Migration timing is based on PIT tag passage of Hanford natural origin and Priest Rapids Hatchery in the adult fish ladder at Bonneville Dam. ..... 48
Table 38 Age compositions for fall Chinook salmon sampled in the Hanford Reach escapement compared to fall Chinook salmon sampled at Priest Rapids Hatchery (genders combined), Brood Years 1998-2012. ..... 50
Table 39 Age compositions for natural and hatchery origin fall Chinook salmon sampled in the Hanford Reach escapement, Brood Years 2007-2012. ..... 51
Table 40 Mean fork length (cm) at age (total age) of fall Chinook salmon sampled in the Hanford Reach escapement compared to fall Chinook salmon sampled at Priest Rapids Hatchery, Brood Years 1999-2012. N = sample size and SD = 1 standard deviation. ..... 53
Table 41 Mean fork length (cm) at age (total age) of natural and hatchery origin fall Chinook salmon that spawned naturally in the Hanford Reach, Brood Years 2007- 2012. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation. ..... 54
Table 42 Comparisons male to female ratio of fall Chinook salmon sampled at Priest Rapids Hatchery and in the Hanford Reach stream surveys, Brood Years 2007- 2012. ..... 56
Table 43 Comparison male to female ratio of fall Chinook salmon sampled in the Hanford Reach stream surveys, Brood Years 2007-2012. ..... 57
Table 44 Comparison of egg retention of natural and hatchery origin fall Chinook sampled in the Hanford Reach stream survey, Return Years 2015-2017. ..... 58
Table 45 Comparison of egg retention of natural and hatchery origin fall Chinook sampled in the Hanford Reach stream survey, Return Years 2010-2017. ..... 59
Table 46 Hatchery fall Chinook salmon contributions to harvest in the Hanford Reach fall Chinook salmon fishery. Coded-wire tag recoveries provided from RMIS database were expanded by sample rate and juvenile tag rate, Return Years 2003- 2017. ..... 61
Table $47 \quad$ Priest Rapids Hatchery coded-wire tag recoveries provided from RMIS by broodyear and harvest type expanded by sample rate and juvenile tag rate, Brood Years1997-2011. Data only includes coded-wire tag recoveries from adipose clippedfish expanded by the juvenile tag rate61
Table 48 Estimated number and proportions of Priest Rapids Hatchery fall Chinook salmon spawning escapement to Priest Rapids Hatchery and stream within and outside of the presumptive target stream by brood year (1992-2011). Coded-wire tag recoveries are expanded by juvenile mark rate and survey sample rate for each brood year. ..... 63
Table 49 Proportion of fall/summer Chinook spawning populations by return year (2000- 2016) comprised of Priest Rapids Hatchery fall Chinook from 1998-2011 brood releases based on coded-wire tag recoveries. ..... 64
Table 50 Last observations of unique PIT tagged adult fall Chinook from Priest Rapids Hatchery at detection sties upstream of McNary Dam, Brood Years 1999-2014. 65
Table 51 Origin of broodstock and pNOB apportioned to program for fall Chinook salmon spawned at Priest Rapids Hatchery, Return Years 20102-2017. ..... 66
Table 52 Origin of broodstock and pNOB apportioned to program for fall Chinook salmon spawned at Priest Rapids Hatchery, Brood Year 2017. ..... 67
Table 53 Proportion of hatchery Chinook salmon on the spawning grounds (pHOS) in the Hanford Reach, Brood Years 2012-2017. ..... 67
Table $54 \quad$ Origin of pHOS apportioned by program source for fall Chinook salmon spawning naturally in the Hanford Reach, Return Years 2012-2017. ..... 68
Table $55 \quad$ PNI of the Hanford Reach fall Chinook salmon supplementation program based on expanded coded-wire tag recoveries of all fish surveyed, Return Years 2001- 2017 ..... 69
Table 56 PNI estimates for the Hanford Reach fall Chinook salmon supplementation programs based on otoliths, Return Years 2012-2017. Calculated from multiple population gene flow model based on the Ford model which has been extended to three or more populations ..... 70
Table $57 \quad$ Broodstock spawned at Priest Rapids Hatchery, estimated escapement to the Hanford Reach, natural and hatchery origin recruits (NOR and HOR), and natural and hatchery replacement rates (NRR and HRR, with and without harvest) for natural origin fall Chinook salmon in the Hanford Reach, Brood Years 1996- 2011 ..... 71
Table 58 Smolt-to-adult Survial Ratios (SAR) for Priest Rapids Hatchery fall Chinook salmon, Brood Years 1992-2011. Data includes all coded-wire tag recoveries from adipose clipped fish ..... 72
Table 59 Smolt-to-adult Survival Ratios (SAR) for Hanford Reach natural origin fall Chinook salmon, Brood Years 1992-2011. Data includes all coded-wire tag recoveries from adipose clipped fish. ..... 73
Table 60 Recoveries and disposition of steelhead at the Priest Rapids Hatchery volunteer trap, Return Year 2017 ..... 74
Table 61 Disposition of Chinook salmon removed from Priest Rapids Hatchery volunteer trap, Return Year 2001-2017. ..... 74

## List of Appendices

Appendix A Evaluation of Coded-Wire Tag Bias ..... A-1
Appendix B Recovery of coded-wire tags collected from adult returns to the Priest Rapids Hatchery Volunteer Trap during Return Year 2017 ..... B-1
Appendix C Juvenile fish health inspections for Priest Rapids Hatchery fall Chinook salmon, Brood Years 1998-2017. The description in the Condition column indicates the presence of a certain condition within at least one of the fish examined. ..... C-1
Appendix D Number and percent of fall Chinook salmon redds counted in different reaches of the Columbia River, 2001-2017. Data for years 2001-2010 was collected by staff with Pacific Northwest National Laboratory. Data for years 2001-2017 was collected by staff with Environmental Assessment Services, LLC. ..... D-1
Appendix E Historical numbers of Chinook salmon carcasses recovered during the annual Hanford Reach fall Chinook salmon carcass survey, Return Years 1991-2017..E-1
Appendix F Estimated escapements for fall Chinook spawning in Hanford Reach and Priest Rapids Dam pool, Return Year 2017. ..... F-1
Appendix G Carcass drift assessment ..... G-1
Appendix H Carcass bias assessment results ..... H-1
Appendix I Demographic comparisons for double index tag groups released from Priest Rapids Hatchery, Brood Years 2009-2014. ..... I-1
Appendix J Alternative pNOB and PNI Estimates ..... J-1
Appendix K Explanation of methods for calculating adult-to-adult expansions based on coded- wire tag recoveries at Priest Rapids Hatchery ..... K-1

### 1.0 Introduction

The Public Utility District No. 2 of Grant County, Washington (Grant PUD) produces and releases 5.6 million subyearling fall Chinook salmon smolts from Priest Rapids Hatchery (PRH) as part of its mitigation for the construction and operation of Priest Rapids and Wanapum dams. The mitigation is the result of three components 1 ) inundation of historic spawning habitat (5 million), annual losses of fish that migrate through the project $(325,543)$, and flow fluctuation impacts in the Hanford Reach $(273,961)$. The PRH is located on the east bank of the Columbia River immediately downstream of Priest Rapids Dam (Figure 1 and Figure 2). The Washington Department of Fish \& Wildlife (WDFW) operates PRH which is owned, maintained, and funded by the Grant PUD. This report describes the monitoring and evaluation of the PRH M\&E program.
PRH also produces fish for other programs. PRH produces and releases 1.7 million subyearling smolts on-site for the U.S. Army Corps of Engineers (USACE) John Day Mitigation. An additional 4.1 million eyed eggs are targeted to provide fish for the USACE John Day Mitigation released at Ringold Springs Hatchery (RSH). The eggs for the RSH program are first transferred to Bonneville Hatchery for marking and ultimately $\sim 3.7$ million subyearlings are transported to, acclimated, and released as subyearling smolts from RSH. In recent years, PRH has accommodated egg-takes for fall Chinook salmon programs managed by either Yakama Nation (YN) or Umatilla Tribe as well as the WDFW's Salmon in the Classroom program and to support various research projects.

A Monitoring and Evaluation Plan for all Grant, Douglas, and Chelan County Public Utility Districts Hatchery Programs has been updated and approved by the committees that oversee the PUD hatchery programs (Hillman et al. 2017). This document provides guiding principles and approaches for the monitoring and evaluation (M\&E) of all PUD hatchery programs including PRH. Objectives, hypotheses, measured and derived variables, and field methods that were used to collect data are listed in this document.

This report of the PRH M\&E program is the eighth annual report (Hoffarth and Pearsons 2012a, 2012b, Richards et al. 2013, Richards and Pearsons 2014, 2015, 2016, and 2017) and encompasses data collected during the Washington State fiscal year (FY) 2017-2018 as well as earlier years where data were available. The data presented in this report are preliminary and subject to change as new data and analyses become available. Readers are encouraged to consult the most recent annual report in order to obtain the most current and accurate information.


Figure 1 Location of Priest Rapids and Ringold Spring hatcheries and the Hanford Reach (indicated by stars).


Figure $2 \quad$ Priest Rapids Hatchery facility and Priest Rapids Dam Off-Ladder Adult Fish Trap (OLAFT).

### 2.0 Objectives

The objective of the PRH M\&E plan is to evaluate the performance of the PRH program relative to the goals and objectives of the PRH program. The overarching goal of the PRH program is to meet Grant PUD's hatchery mitigation by producing fish for harvest while keeping genetic and ecological impacts within acceptable limits. The M\&E objectives of the PRH program are described below.

Objective 1: Determine if conservation programs have increased the number of naturally spawning and naturally produced adults of the target population and if the program has reduced the natural replacement rate (NRR) of the supplemented population.

Objective 2: Determine if the proportion of hatchery fish on the spawning ground affects the freshwater productivity of supplemented stocks.

Objective 3: Determine if the hatchery adult-to-adult survival (i.e., hatchery replacement rate, HRR) is greater than the natural adult-to-adult survival (i.e., natural replacement rate, NRR) and the target hatchery survival rate.

Objective 4: Determine if the proportion of hatchery origin spawners (pHOS or PNI) is meeting management targets.

Objective 5: Determine if the run timing, spawn timing, and spawning distribution of the hatchery component is similar to the natural component of the target population or is meeting programs-specific objectives.

Objective 6: Determine if stray rate of hatchery fish is below the acceptable levels to maintain genetic variation among stocks.

Objective 7: Determine if genetic diversity, population structure, and effective population size have changed in natural spawning populations as a result of the hatchery program.
Objective 8: Determine if hatchery programs have caused changes in phenotypic characteristics of natural populations.

Objective 9: Determine if hatchery fish were released at programmed size and number.
Objective 10: Determine if appropriate harvest rates have been applied to the conservation, safety-net, and segregated harvest programs to meet the HCP/SSSA goal of provided harvest opportunities while also contributing to population management and minimalizing risk to natural populations.

We also present information in this report about two regional objectives that relate to disease and ecological interactions.

### 3.0 Project Coordination

WDFW M\&E staff partially assigned to PRH also conducts similar work at RSH. The M\&E staff also works in conjunction with multiple WDFW groups that include PRH fish culture staff, the Columbia River Coded-Wire Tag Recovery Program (CRCWTRP), Region 3 Fish Management staff, the Supplementation Research Team in Wenatchee, and the Grant PUD biological science staff to complete many of the tasks included in the M\&E Plan. In addition, samples collected at the hatchery and in the field were transported to and analyzed by WDFW laboratories including the WFDW Scale Reading Lab and the WDFW Otolith Lab. Coded-wire tags (CWT) were processed by the M\&E staff either at the WDFW District 4 office or the PRH wet lab. Data and
analyses collected in association with the PRH M\&E and Hanford Reach population monitoring are incorporated into the WDFW Traps, Weirs, and Surveys (TWS) database which is administered by the WDFW staff stationed in the Region 5 Headquarters in Vancouver. Agency managers use these data for forecasting and managing fall Chinook salmon populations in the Columbia and Snake rivers and tributaries. WDFW and Grant PUD secured and held all environmental permits necessary for the work described in this report.

### 4.0 Life History - Hanford Reach Fall Chinook Salmon

The Hanford Reach is one of the last non-impounded reaches of the Columbia River and the location of the largest and most productive natural spawning fall Chinook salmon population in the United States (Harnish et al. 2012, Langshaw et al. 2015, Harnish 2017, Langshaw et al. 2017). The Hanford Reach extends 51 miles from the city of Richland to the base of Priest Rapids Dam. Natural origin fall Chinook salmon emerge from the substrate in the spring and rear in the Hanford Reach until outmigration in the summer. Egg-to-fry survival has been estimated to be about $71 \%$ in the Hanford Reach (Oldenburg et al. 2012) and egg-to-pre-smolt survival has been estimated to be about $40.2 \%$ (Harnish et al. 2012) or more recently at $63.4 \%$ (Harnish, 2017). Both of these estimates are high when compared to other Chinook salmon populations and flow management within the Hanford Reach has resulted in improvements in survival (Harnish et al. 2012, Harnish 2017, Langshaw et al. 2017). The age at maturity for naturally produced fish in the Hanford Reach varies between age-1 mini-jack and age-6 adults: albeit recoveries of age- 1 and 6 fish are generally rare. The age of fish reported in this document begins with the first birthday occurring the year after the parents spawned. The abundance of mini-jacks which mature as age-1 males is currently not known. Age-2 male fall Chinook salmon (a.k.a jacks) return to the Hanford Reach after spending roughly one year in the ocean. The majority of the natural origin adults return after spending three to four years in the ocean (age-4 and 5). A small portion, typically less than $2 \%$, will spend up to five years in the ocean and return as age- 6 . The ocean distribution of natural and hatchery origin Hanford Reach upriver brights are similar and range from the northern California coast to the Gulf of Alaska (Norris et al. 2000, Weitkamp 2010). The majority of the adults migrate north of the Columbia River with the harvest primarily occurring in non-selective ocean and freshwater fisheries (Norris et al, 2000). Adults return to the mouth of the Columbia River between August and October and spawn in large cobble substrate between October and December (Langshaw et al. 2017).

### 5.0 Sample Size Considerations

We attempted to strike an appropriate balance between objectives, statistical precision, logistics, and financial investment when setting sample size targets. A variety of approaches were used for setting sample sizes and this depended upon the objective. For example, a phased subsampling approach was used in some cases to determine age and origin and $100 \%$ sampling occurred in others (e.g., CWT, otoliths in fecundity samples). A phased approach was used to collect some biological samples with sufficient accuracy and precision. In the phased approach, we attempted to collect an excess number of raw samples such as carcasses and trap recoveries and then use post season analysis to determine sub-sampling strategies for otolith and scale reading where appropriate. The sample size target of systematic field sampling for later otolith reading is 2,500 of the carcasses in the Hanford Reach, 1,000 at the hatchery trap, and 1,000 of the hatchery volunteer broodstock, and 200 broodstock collected from each other source such as OLAFT and ABC fishery.

All adult fall Chinook salmon recovered at PRH, in the Hanford Reach sport fishery, and in the stream surveys were sampled for the presence of CWT to maximize the precision of estimates generated from these data. Representative otolith samples by survey type were randomly selected as a sub-sample for decoding to estimate origin by age class if numbers allowed. In some cases, all otolith samples for a survey type were processed if the sampling rate provided relatively low numbers of otoliths collected or if there was a need for higher precision or accuracy. During return year 2017, randomly selected sub-samples of otoliths collected from the PRH volunteer trap and volunteer broodstock were submitted for decoding. The methodologies for selecting otolith sub-samples have differed between return years as field methods changed and as new analyses facilitated improvements in approaches. In general, we randomly select otoliths from various survey types to obtain roughly 120 otoliths for each age and gender. In some cases, all otoliths were submitted for stratified groups (age/gender) when specific age classes contain less than 100 samples. For example, typically all samples of age- 5 and 6 fish were submitted because of the low number of fish represented in the field collected sample. The stratified sub-sample size refinement process is described in Richards and Pearsons (2014). The sub-sample groups often included fish possessing a CWT within the biological sample which increased the number of fish sampled for origin with no additional cost.

### 6.0 Current Operation at Priest Rapids Hatchery

The 2017 broodstock for PRH were collected at the hatchery volunteer trap, the OLAFT, and from the ABC fishery. The majority of the broodstock were collected from the PRH volunteer trap which was operated from September 11 through December 8, 2017. A total of 19,259 mature fall Chinook salmon were handled at during broodstock collection activities (Table 1). In attempt to increase pNOB for the Grant County Public Utility District No 2 (Grant PUD) program, the broodstock ponded excluded adipose intact fish with a fork length less than 74 cm and known hatchery fish (i.e., possessing an adipose clip and or CWT); hence, increasing the potential number of natural origin broodstock ponded. The USACE program included known hatchery fish due to the unavailability of adipose fin intact/non CWT fish. A portion of these known hatchery origin fish ponded were surplused as they were replaced by adipose fin intact/non CWT fish during subsequent trapping and ponding operations. In total, 1,566 known hatchery fish from the PRH volunteer trap were spawned for the Grant PUD and USACE program, most of which were used for the USACE program.
A portion of the fish intended for surplus from PRH were utilized for broodstock to support various fall Chinook salmon productions of Umatilla Tribe and Yakama Nation. These fish include 793 shipped to RSH to be shared between the two tribes and 931 fish spawned at PRH by the Yakama Nation staff. The PRH monitoring and evaluation (M\&E) staff categorized and sampled these fish as surplus from PRH. The carcasses were utilized for pet food since they were treated with formalin during the period in which they were held for broodstock.

Table 1 Source and disposition of Chinook salmon collected for broodstock at Priest Rapids Hatchery, Return Year 2017.

| Collection Location | Gender | Collected | Trap Surplused | Trap Mortalities | Ponded | Spawned ${ }^{1}$ | Pond Surplused | Pond Mortalities |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Volunteer Trap <br> (Sept 11 - Dec 8) | Males | 7,784 | 5,329 | 60 | 2,386 | 1,526 | 469 | 400 |
|  | Females | 8,587 | 3,748 | 103 | 4,651 | 2,985 | 873 | 865 |
|  | Jacks | 1,441 | 1,365 | 73 | 3 | 0 | 1 | 2 |
|  | Total | 17,812 | 10,442 | 236 | 7,040 | 4,511 | 1,343 | 1,267 |
| OLAFT <br> (Sept 8 - Oct 17) | Males | 251 | 0 | 0 | 251 | 177 | 15 | 59 |
|  | Females | 720 | 0 | 0 | 722 | 632 | 10 | 80 |
|  | Jacks | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Total | 973 | 0 | 0 | 973 | 809 | 25 | 139 |
| $\begin{gathered} \text { ABC } \\ \text { (Oct 27, } 28 \text { \& 29) } \end{gathered}$ | Males | 180 | 0 | 0 | 180 | 132 | 7 | 41 |
|  | Females | 296 | 0 | 0 | 296 | 216 | 4 | 76 |
|  | Jacks | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Total | 476 | 0 | 0 | 476 | 348 | 11 | 117 |
| Facility | Total | 19,261 | 10,442 | 236 | 8,489 | 5,668 | 1,379 | 1,523 |

${ }^{1}$ There were 9 males and 72 females taken directly from the trap and spawned. These fish are not included in the total fish ponded.

The pattern of arrival timing by week (Sunday through Saturday) for adult fall Chinook salmon to the PRH Volunteer Trap was determined to help schedule future sampling and broodstock activities. Trap operations during 2017 should have provided unimpeded access to the trap during most of each week. The trap was often closed Sunday afternoon through Tuesday from late October through mid-November due to spawning activities occurring Mondays and Tuesdays. The 2017 collection numbers suggest that peak arrival to the PRH Volunteer Trap occurred during late October and early November (Figure 3). This pattern is similar to that observed for years 2015 and 2016. The annual arrival timing prior to 2015 was typically a bimodal peak in which a large proportion of fish returning in late September and late October.


Figure $3 \quad$ Weekly counts of fish adult Chinook salmon collected at the Priest Rapids Hatchery Volunteer Trap, 2017.

PRH has four adult salmon holding ponds. Ponds 1 and 2 were used to hold broodstock collected at the PRH Volunteer Trap. Pond 3 was used on occasion to temporarily hold males collected from ABC and OLAFT or surplus broodstock from the PRH Volunteer Trap. Pond 4 was used to hold broodstock collected from the ABC and OLAFT along with 51 males collected from the PRH Volunteer Trap. The PRH staff generally transported fish from the volunteer trap seven days per week to collect broodstock and or to surplus the excess fish. Male fall Chinook salmon typically comprised the majority of the fish surplused from the trap. Spawning days generally occurred on Mondays and Tuesdays each week from October 23 through December 4 ( $\mathrm{N}=13$ ). The hatchery staff generally seined fish in each pond to sort and collect ripe fish in order to expedite the spawning operations and reduce stress on broodstock. The electro-anesthesia system was infrequently used for spawning fish in 2017 as it was deemed inefficient and too stressful on mature broodstock.

The egg-take goal from the 2017 PRH brood was 13,530,000 eggs. The actual egg-take for the Grant PUD and USACE programs was 13,738,916 ( $\sim 102 \%$ of the goal). During routine spawn days, the eggs from two females were stripped into a five-gallon bucket and then the milt from a single male was mixed with the eggs. Two buckets of fertilized eggs were then combined to help ensure fertilization. Fertilized eggs were then transferred to the incubation room, weighed to estimate numbers of eggs, and then placed in vertical incubation trays at roughly 7,000 eggs per tray.

Since 2014, a cooperative effort between WDFW and Grant PUD staff to perform real-time otolith reading (RTOR) coinciding with an alternative mating strategy has occurred during midNovember. In 2017, the RTOR occurred on November 13. This activity included examining 125 otoliths from unmarked males collected at OLAFT or the ABC fishery to identify 108 viable
natural origin fish. These 108 males were used for 1 x 4 matings with unmarked females collected either at the volunteer trap, OLAFT, or the ABC fishery. Milt from natural origin males was mixed with 4 females in a five-gallon bucket and then a pair of five-gallon buckets of eggs were combined before being transferred to the incubation room. An estimated 1,621,391green eggs were taken during the RTOR 1:4 crosses.

After shipping two large groups of eyed eggs to Bonneville Hatchery for hatching and early rearing, fry from the remaining eleven egg-takes were moved from the vertical trays in the incubation building to outdoor raceways between January 26 and March 8, 2017. The fry were reared in the raceways until they were of sufficient size that a portion of them could be marked in some manner (i.e., adipose clipped and or tagged). Fish receiving marks and or tags were collected directly from the raceways banks and then released into the corresponding concrete rearing ponds (e.g., fish moved from raceway bank E to channel pond E). Fish not selected for marking were transferred from the raceway banks into the corresponding rearing ponds. The growth of smolts from ponds E and D was accelerated for early releases that occurred on May 22 and 25, respectively. The remaining smolts were released between June 12 and June 21. All releases occurred at night. These fish migrate down the old one mile long spawning channel and then down the hatchery discharge channel to the Columbia River.

### 7.0 Origin of Adult Returns to Priest Rapids Hatchery

The origin of fish collected from the three locations was determined by examination of hatchery marks (i.e., otolith thermal marks, adipose clips, and CWTs) for the fish within the demographic sample groups. PRH origin fish were identified by their otolith mark or a CWT. The fish that did not possess an otolith mark or other hatchery marks and tags were classified as natural origin. Historically, the very low recovery ( $<1 \%$ ) of non-adipose clipped CWT strays at PRH suggests that a high percentage of the fish not possessing any type of hatchery mark may be of natural origin. In some sections of the report, we make a simplifying assumption that fish without hatchery marks are of natural origin.

Similar to that observed in previous years, there is a discrepancy between estimates of origin based on CWT and those based on otoliths marks. It's believed that estimates of origin based on otolith sampling may provide the most accurate data under the current marking regime at PRH due to discrepancies in the data associated with CWT results (Appendix A).
An examination of thermal mark accuracy was conducted for 2017. Where 360 known origin otoliths were blindly examined amongst the Hanford Reach spawning survey. An overall error rate of $3.8 \%$ was detected from the known origin samples. The majority of error (4.7\%) was found to be false negatives (no mark was detected when a mark should have been present $\mathrm{n}=13$ ), while only a single fish was falsely identified as thermally marked where it should have had no mark (1.1\%). Preliminary results suggest no directional bias. These results were similar to results found in Volk et al. (1999) for false negative error (1-5\%) but slightly better for false positives ( $\sim 6 \%$ error) (L. Campbell, WDFW, personal communication).

We present estimates of abundance based on CWTs (1:1 sample rate) and estimates based on sub-samples of hatchery marked fish collected from specific groups (varying sample rates) to illustrate differences in the estimates for the proportions of natural and hatchery origin fish recovered at PRH as well as the potential for creating a method to correct the historical database that was generated using CWT recoveries.

### 7.1 Origin Based on Hatchery Marks

For return year 2017, the proportion of broodstock obtained from the PRH volunteer trap that was natural origin is estimated at 0.092 . Overall, it is estimated that 0.101 of the fish surplused or removed as mortalities that originated from the PRH volunteer trap were natural origin. The proportion of natural origin fish used as broodstock from the OLAFT and ABC was estimated to be 0.864 and 0.907 , respectively. The estimated numbers of natural and hatchery origin broodstock spawned annually since return year 2013 are given in (Table 2).

For return years 2014 through 2017, a minimum fork-length threshold of ~73 cm was generally used to reduce the number of age-2 and 3 broodstock collected at OLAFT and the PRH volunteer trap along with the exclusion of hatchery marks and tags. Historical data suggests that age-2 and 3 fall Chinook salmon returning to the Hanford Reach comprise of a greater proportion of hatchery origin fish compared to age- 4 and 5 fall Chinook salmon returning to the Hanford Reach.

Table 2 Total fish handled, numbers sampled, and estimates of hatchery and natural origin Chinook salmon collected at Priest Rapids Hatchery, Priest Rapids Dam Off-Ladder Adult Fish Trap, and Angler Broodstock Collection fishery. Origin determined by otolith thermal marks, presence of coded-wire tags, and/or adipose clips, Return Years 2013-2017.

| Priest Rapids Hatchery Broodstock ${ }^{1}$ |  |  | Estimate (95\% CI) |  |
| :---: | :---: | :---: | :---: | :---: |
| Return Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| 2013 | 4,476 | 503 | 4,395 [4,319, 4,436] | 81 [40, 157] |
| 2014 | 4,427 | 574 | 4,228 [4,130, 4,294] | 199 [133, 297] |
| 2015 | 4,875 | 682 | 4,482 [4,368, 4,573] | 393 [302, 507] |
| 2016 | 4,324 | 827 | 4,067 [4,034, 4,095] | 257 [227, 290] |
| 2017 | 4,511 | 533 | 4,093 [3,967, 4,197] | 417 [414, 543] |
| Priest Rapids Hatchery Surplused from Trap |  |  | Estimate (95\% CI) |  |
| Return Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| $2013{ }^{\text {a }}$ | 37,355 | 608 | 36,085 [35,375, 36,533] | 1,270 [822, 1,980] |
| $2014{ }^{\text {b }}$ | 73,352 | 639 | 69,024 [67,484, 70,271] | 4,328 [3,081, 5,868] |
| $2015{ }^{\text {b }}$ | 57,625 | 619 | 54,646 [53,418, 55,551] | 2,979 [2,075, 4,207] |
| $2016{ }^{\text {a }}$ | 24,461 | 1,033 | 23,790 [23,737, 23,837] | 668 [619, 719] |
| 2017 | 13,301 | 1,426 | 11,954 [10,680, 10,803 ] | 1,348 [1218, 1492] |
| Off Ladder Fish Trap Broodstock ${ }^{\mathbf{1}}$ |  |  | Estimate (95\% CI) |  |
| Return Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| 2013 | 763 | 169 | 343 [242, 370] | 420 [392, 416] |
| 2014 | 825 | 225 | 143 [122, 166] | 682 [659, 703] |
| 2015 | 348 | 164 | 45 [29, 66] | 303 [282, 319] |
| 2016 | 366 | 211 | 99 [83, 117] | 267 [249, 283] |
| 2017 | 809 | 226 | 108 [78, 148] | 701 [661, 731] |
| Angler Broodstock Collection Broodstock |  |  | Estimate (95\% CI) |  |
| Return Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| 2013 | 308 | 293 | 59 [46, 75] | 249 [233, 262] |
| 2014 | 221 | 111 | 17 [9, 34] | 204 [187, 212] |
| 2015 | 301 | 141 | $11[4,26]$ | 290 [275, 297] |
| 2016 | 247 | 94 | $11[6,20]$ | 236 [227, 241] |
| 2017 | 348 | 171 | 33 [20, 52] | 315 [296, 328] |

${ }^{1}$ Includes only fish that were spawned.
${ }^{2}$ Origin based on the absence of otolith marks, coded-wire tags, or adipose clips.
${ }^{\text {a }}$ This data was collected from samples intermittently high-graded for broodstock and may not be representative of the entire return to the Priest Rapids Hatchery volunteer trap.
${ }^{\mathrm{b}}$ This data is representative of the entire volunteer return to the Priest Rapids Hatchery volunteer trap.

### 7.2 Origin Based on Coded-Wire Tag Recoveries

The expansions of CWT recoveries at PRH have until recent years frequently under estimated the returns of PRH origin fish by return year and brood year. This bias and steps taken to identify the source are provided in Appendix A.

All Chinook salmon returning to PRH and broodstock collected from the OLAFT and ABC were sampled for the presence of CWT. A total of 2,733 CWT fish were recovered from Chinook salmon sampled at PRH in 2017, of which 527 were obtained from the broodstock obtained from the PRH volunteer trap (Appendix B). The broodstock collected from the PRH volunteer trap were generally culled to exclude CWT fish for the purpose of increasing natural origin broodstock. Therefore, this CWT group is not representative of the volunteer broodstock. The ABC fish were not screened for a CWT during collection but were later scanned for CWT at the hatchery. There were nine CWT recovered from the ABC collection of which seven were
surplused. The staff collecting the OLAFT fish attempted to screen out CWT fish during the brood stock collection. There were four CWT recovered from the OLAFT collection and included in the spawn. The juvenile mark rate expansions of CWT recovered adults at PRH in 2017 suggest that $91.5 \%$ of the returns to the PRH volunteer trap were hatchery origin fish. If we were to make the assumption that these CWT expansions accurately reflected the proportion of hatchery origin fish, then the remaining $8.5 \%$ of the unaccounted fish could potentially be natural origin (Table 3).
During return year 2017, PRH origin CWT tags accounted for $86.9 \%$ of the total return to the PRH volunteer trap. There were 8 natural origin CWT Hanford Reach fall Chinook salmon recovered at the hatchery in 2017 of which 7 were excluded from the broodstock while sorting out adipose clipped fish to increase the proportion of natural origin broodstock. There is not an expansion factor for the natural origin CWT fish so there was no attempt to estimate the proportion of natural origin fish based on these CWT recoveries.

## Table 3 Estimated proportion of hatchery and natural origin adult Chinook salmon returning to the Priest Rapids Hatchery volunteer based on coded-wire tag expansion. The entire collection was sampled for coded-wire tags, Return Years 2005-2017.

| Return <br> Year | Returns to Priest <br> Rapids Hatchery <br> Volunteer Trap | Origin based on Coded-Wire Tag expansions |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Other Hatchery | Natural Origin $^{\mathbf{1}}$ |  |  |
|  | 10,616 | 0.622 | 0.006 | 0.329 |
| 2006 | 8,223 | 0.490 | 0.006 | 0.436 |
| 2007 | 6,000 | 0.671 | 0.004 | 0.525 |
| 2008 | 19,586 | 0.491 | 0.008 | 0.409 |
| 2009 | 12,778 | 0.428 | 0.003 | 0.540 |
| 2010 | 19,169 | 0.602 | 0.003 | 0.486 |
| 2011 | 20,823 | 0.613 | 0.006 | 0.381 |
| 2012 | 28,039 | 0.692 | 0.004 | 0.304 |
| 2013 | 41,831 | 0.713 | 0.034 | 0.252 |
| 2014 | 77,259 | 0.809 | 0.020 | 0.170 |
| 2015 | 63,978 | 0.914 | 0.015 | 0.071 |
| 2016 | 28,786 | 0.912 | 0.024 | 0.064 |
| 2017 | 17,812 | 0.868 | 0.046 | 0.086 |
| Mean | $\mathbf{2 7 , 3 0 0}$ | $\mathbf{0 . 6 7 9}$ | $\mathbf{0 . 0 1 4}$ | $\mathbf{0 . 3 1 2}$ |
| Median | $\mathbf{2 0 , 2 0 5}$ | $\mathbf{0 . 6 7 1}$ | $\mathbf{0 . 0 0 6}$ | $\mathbf{0 . 3 2 9}$ |

${ }^{1}$ The proportion not accounted for by coded-wire tag expansion is assumed to be of natural origin.

### 8.0 Broodstock Collection and Sampling

The broodstock collected at the PRH volunteer trap were systematically sampled at a rate of 1:5 for otoliths (origin), scales (age), gender, and length. The broodstock collected at the OLAFT and ABC were sampled at a 1:2 rate for otoliths (origin), scales (age), gender, and length.

### 8.1 Broodstock Age Composition

A combined total of 5,668 fish were spawned from the three sources of broodstock (i.e, PRH Volunteer Trap, ABC and OLAFT). The historical broodstock age compositions are not directly comparable to the 2012 through 2017 broodstock age compositions due to inconsistent methodology for assigning origin (Table 4). Prior to 2012, the origin of broodstock was estimated by adult CWT recoveries which in turn were expanded by the specific juvenile tag
rates. The broodstock age compositions for 2016 and 2017 are influenced by the selection of broodstock based on a 74 cm minimum fork length.

Table 4 Age composition for hatchery and natural origin fall Chinook salmon spawned at Priest Rapids Hatchery (includes all sources of broodstock), Return Years 2007-2017. Proportions calculated from expanded age compositions by origin for each source of broodstock to account for differing sample rates.

| Return Year | Origin | Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural ${ }^{1}$ | 0.000 | 1.000 | 0.000 | 0.000 | 0.000 |
|  | Hatchery ${ }^{1}$ | 0.081 | 0.274 | 0.486 | 0.138 | 0.020 |
| 2008 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery ${ }^{1}$ | 0.011 | 0.848 | 0.100 | 0.039 | 0.002 |
| 2009 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery ${ }^{1}$ | 0.012 | 0.086 | 0.883 | 0.019 | 0.000 |
| 2010 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery | 0.016 | 0.755 | 0.111 | 0.118 | 0.000 |
| 2011 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery ${ }^{1}$ | 0.010 | 0.229 | 0.753 | 0.008 | 0.000 |
| 2012 | Natural ${ }^{2}$ | 0.032 | 0.435 | 0.400 | 0.131 | 0.002 |
|  | Hatchery ${ }^{2}$ | 0.006 | 0.487 | 0.376 | 0.130 | 0.000 |
| 2013 | Natural ${ }^{2}$ | 0.000 | 0.446 | 0.517 | 0.037 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.001 | 0.658 | 0.339 | 0.002 | 0.000 |
| 2014 | Natural ${ }^{2}$ | 0.000 | 0.045 | 0.886 | 0.070 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.064 | 0.897 | 0.039 | 0.000 |
| 2015 | Natural ${ }^{2}$ | 0.000 | 0.183 | 0.506 | 0.305 | 0.006 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.210 | 0.680 | 0.110 | 0.000 |
| 2016 | Natural ${ }^{2}$ | 0.000 | 0.101 | 0.761 | 0.138 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.099 | 0.700 | 0.196 | 0.007 |
| 2017 | Natural ${ }^{2}$ | 0.000 | 0.130 | 0.618 | 0.252 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.074 | 0.663 | 0.258 | 0.005 |

${ }^{1}$ Origin determined from coded-wire tag expansions of juvenile mark rate.
${ }^{2}$ Origin determined from presence of hatchery marks (i.e., coded-wire tags, adipose clips, and otoliths)
In recent years, the broodstock selected from the PRH volunteer trap consisted primarily of age-4 fish (Table 5). The hatchery origin broodstock for return years 2012 and 2013 had higher proportions of age- 3 fish due to the scarcity of older fish returning to the trap. The hatchery and natural origin broodstock selected at the OLAFT were primarily age-4 (Table 6). Adipose clipped fish and jacks were generally excluded from the fish collected from the ABC fishery. In recent years, both the PRH origin and natural origin broodstock from the ABC were mostly age4 (Table 7).

Table 5 Age composition for hatchery and natural origin fall Chinook broodstock collected from the Priest Rapids Hatchery volunteer trap, Return Years 2012-2017.

| Return Year | Origin ${ }^{1}$ | Age Composition |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | N | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2012 | Natural | 39 | 0.000 | 0.295 | 0.585 | 0.121 | 0.000 |
|  | Hatchery | 646 | 0.000 | 0.477 | 0.389 | 0.134 | 0.000 |
| 2013 | Natural | 11 | 0.000 | 0.390 | 0.610 | 0.000 | 0.000 |
|  | Hatchery | 497 | 0.000 | 0.656 | 0.342 | 0.002 | 0.000 |
| 2014 | Natural | 26 | 0.000 | 0.115 | 0.885 | 0.000 | 0.000 |
|  | Hatchery | 548 | 0.000 | 0.065 | 0.899 | 0.036 | 0.000 |
| 2015 | Natural | 55 | 0.000 | 0.218 | 0.491 | 0.273 | 0.018 |
|  | Hatchery | 627 | 0.000 | 0.215 | 0.668 | 0.116 | 0.000 |
| 2016 | Natural | 49 | 0.000 | 0.102 | 0.776 | 0.122 | 0.000 |
|  | Hatchery | 778 | 0.000 | 0.100 | 0.763 | 0.136 | 0.000 |
| 2017 | Natural | 49 | 0.000 | 0.290 | 0.544 | 0.167 | 0.000 |
|  | Hatchery | 484 | 0.000 | 0.075 | 0.662 | 0.258 | 0.005 |
| Mean | Natural | 38 | 0.000 | 0.235 | 0.649 | 0.114 | 0.003 |
|  | Hatchery | 597 | 0.000 | 0.265 | 0.621 | 0.114 | 0.001 |

${ }^{1}$ Origin determined from "in-sample" otoliths, adipose clips and/or coded-wire tags.
Table 6 Age composition for hatchery and natural origin fall Chinook salmon broodstock collected from the Off Ladder Adult Fish Trap at Priest Rapids Dam, Return Years 2012-2017.

| Return Year | Origin ${ }^{1}$ | Age Composition |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | N | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2012 | Natural | 281 | 0.048 | 0.540 | 0.257 | 0.151 | 0.004 |
|  | Hatchery | 219 | 0.106 | 0.687 | 0.136 | 0.071 | 0.000 |
| 2013 | Natural | 116 | 0.000 | 0.353 | 0.595 | 0.052 | 0.000 |
|  | Hatchery | 85 | 0.000 | 0.588 | 0.400 | 0.012 | 0.000 |
| 2014 | Natural | 186 | 0.000 | 0.000 | 0.902 | 0.098 | 0.000 |
|  | Hatchery | 39 | 0.000 | 0.000 | 0.870 | 0.130 | 0.000 |
| 2015 | Natural | 143 | 0.000 | 0.132 | 0.514 | 0.347 | 0.007 |
|  | Hatchery | 21 | 0.000 | 0.211 | 0.563 | 0.226 | 0.000 |
| 2016 | Natural | 155 | 0.000 | 0.058 | 0.677 | 0.245 | 0.019 |
|  | Hatchery | 56 | 0.000 | 0.089 | 0.643 | 0.250 | 0.018 |
| 2017 | Natural | 226 | 0.000 | 0.036 | 0.668 | 0.277 | 0.000 |
|  | Hatchery | 35 | 0.000 | 0.028 | 0.723 | 0.249 | 0.000 |
| Mean | Natural | 185 | 0.008 | 0.187 | 0.602 | 0.195 | 0.005 |
|  | Hatchery | 76 | 0.018 | 0.267 | 0.556 | 0.156 | 0.003 |

[^1]Table 7 Age composition for hatchery and natural origin fall Chinook salmon broodstock collected from Angler Broodstock Collection, Return Years 20122017.

| Return Year | Origin ${ }^{1}$ | Age Composition |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | N | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2012 | Natural | 59 | 0.000 | 0.542 | 0.339 | 0.119 | 0.000 |
|  | Hatchery | 6 | 0.000 | 0.667 | 0.333 | 0.000 | 0.000 |
| 2013 | Natural | 237 | 0.000 | 0.511 | 0.468 | 0.021 | 0.000 |
|  | Hatchery | 56 | 0.000 | 0.839 | 0.161 | 0.000 | 0.000 |
| 2014 | Natural | 102 | 0.000 | 0.126 | 0.830 | 0.044 | 0.000 |
|  | Hatchery | 9 | 0.059 | 0.369 | 0.572 | 0.000 | 0.000 |
| 2015 | Natural | 136 | 0.000 | 0.196 | 0.499 | 0.305 | 0.000 |
|  | Hatchery | 5 | 0.000 | 0.397 | 0.603 | 0.000 | 0.000 |
| 2016 | Natural | 90 | 0.000 | 0.156 | 0.656 | 0.189 | 0.000 |
|  | Hatchery | 4 | 0.000 | 0.250 | 0.750 | 0.000 | 0.000 |
| 2017 | Natural | 16 | 0.000 | 0.127 | 0.561 | 0.312 | 0.000 |
|  | Hatchery | 155 | 0.000 | 0.055 | 0.649 | 0.296 | 0.000 |
| Mean | Natural | 107 | 0.000 | 0.276 | 0.559 | 0.165 | 0.000 |
|  | Hatchery | 39 | 0.010 | 0.430 | 0.511 | 0.049 | 0.000 |

${ }^{1}$ Origin determined from "in-sample" otoliths, adipose clips and/or coded-wire tags.

### 8.2 Length by Age Class of Broodstock

The mean fork length (cm) by age for each source of broodstock is provided in Table 8 and Table 9. Both the hatchery origin and natural origin age-3 fish collected at the OLAFT appear to be slightly larger than age-3 fish collected at other locations. This may be due to the size highgrading processes.
Table 8 Mean fork length (cm) at age (total age) of fall Chinook salmon sampled from each source of broodstock spawned at Priest Rapids Hatchery, Return Year 2017. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation.

| Source of Broodstock | Origin ${ }^{1}$ | Fall Chinook Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| Volunteer <br> Returns | Natural | 0 | -- | -- | 15 | 73 | 4 | 26 | 79 | 4 | 8 | 81 | 8 | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 39 | 72 | 4 | 315 | 77 | 4 | 127 | 82 | 6 | 3 | 84 | 3 |
| OLAFT | Natural | 0 | -- | -- | 8 | 75 | 3 | 154 | 79 | 4 | 63 | 85 | 5 | 1 | 96 | -- |
|  | Hatchery | 0 | -- | -- | 1 | 75 | -- | 25 | 80 | 4 | 9 | 82 | 3 | 0 | - | -- |
| ABC | Natural | 2 | 56 | 1 | 20 | 68 | 7 | 84 | 77 | 5 | 49 | 86 | 7 | 2 | 56 | 1 |
|  | Hatchery | 0 | -- | -- | 1 | 67 | -- | 10 | 78 | 4 | 5 | 79 | 4 | 0 | -- | -- |

${ }^{1}$ It is assumed for this analysis that all fish not possessing an otolith mark, ad-clipped or hatchery origin coded-wire tag were natural origin.

Table $9 \quad$ Mean fork length (cm) at age (total age) of hatchery and natural origin fall Chinook salmon collected from broodstock originating from the Priest Rapids Hatchery volunteer trap. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation, Return Years 2012-2017.

| Return Year | Origin ${ }^{1}$ | Fall Chinook Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2012 | Natural | 0 | -- | -- | 12 | 71 | 4 | 25 | 82 | 4 | 5 | 86 | 4 | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 298 | 70 | 4 | 253 | 81 | 5 | 91 | 88 | 7 | 0 | -- | -- |
| 2013 | Natural | 0 | -- | -- | 4 | 76 | 4 | 7 | 78 | 4 | 0 | -- | -- | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 288 | 71 | 4 | 200 | 80 | 5 | 2 | 85 | 4 | 0 | -- | -- |
| 2014 | Natural | 0 | -- | -- | 3 | 74 | 2 | 23 | 80 | 5 | 0 |  |  | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 36 | 70 | 3 | 491 | 78 | 5 | 21 | 87 | 6 | 0 | -- | -- |
| 2015 | Natural | 0 | -- | -- | 12 | 74 | 7 | 30 | 79 | 6 | 15 | 86 | 4 | 1 | 87 | 0 |
|  | Hatchery | 0 | -- | -- | 133 | 71 | 4 | 437 | 80 | 4 | 79 | 84 | 5 | 0 | -- | -- |
| 2016 | Natural | 0 | -- | -- | 78 | 73 | 3 | 594 | 79 | 4 | 106 | 85 | 6 | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 133 | 71 | 4 | 437 | 80 | 4 | 79 | 84 | 5 | 0 | -- | -- |
| 2017 | Natural | 0 | -- | -- | 15 | 73 | 4 | 26 | 79 | 4 | 8 | 81 | 8 | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 39 | 72 | 4 | 315 | 77 | 4 | 127 | 82 | 6 | 3 | 84 | 3 |

${ }^{1}$ It is assumed for this analysis that all fish not possessing an otolith mark, ad-clipped or hatchery origin codedwire tag were natural origin.

### 8.3 Gender Ratios

PRH staff sort and select broodstock from the trap to meet their egg-take goals and male-tofemale spawner ratio which is generally 1:2. Additional broodstock were collected from the OLAFT and ABC. The 2017 broodstock population was comprised of $67.6 \%$ females, resulting in an overall male to female ratio of $0.48: 1.00$, which is lower than the historic mean ratio of 0.53:1.00 (Table 10). This lower ratio of males to females resulted from the 111 matings of 1 male x 4 -females during the real-time otolith read/alternative mating strategy study.

Table 10 Number of male and female hatchery fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 2001-2017. Ratios of males to females are also provided.

| Return Year | Males (M) | Females (F) | M/F Ratio |
| :---: | :---: | :---: | :---: |
| 2001 | 1,697 | 3,289 | $0.52: 1.00$ |
| 2002 | 1,936 | 3,628 | $0.53: 1.00$ |
| 2003 | 1,667 | 3,176 | $0.52: 1.00$ |
| 2004 | 1,688 | 3,099 | $0.54: 1.00$ |
| 2005 | 1,962 | 3,326 | $0.59: 1.00$ |
| 2006 | 1,777 | 3,322 | $0.53: 1.00$ |
| 2007 | 850 | 1,301 | $0.65: 1.00$ |
| 2008 | 1,823 | 3,195 | $0.57: 1.00$ |
| 2009 | 1,531 | 3,000 | $0.51: 1.00$ |
| 2010 | 1,809 | 3,447 | $0.52: 1.00$ |
| 2011 | 1,858 | 3,000 | $0.62: 1.00$ |
| 2012 | 1,749 | 3,225 | $0.54: 1.00$ |
| 2013 | 1,865 | 3,578 | $0.52: 1.00$ |
| $2014^{\text {a }}$ | 1,805 | 3,688 | $0.49: 1: 00$ |
| $2015^{\text {a }}$ | 1,697 | 3,827 | $0.44: 1: 00$ |
| $2016^{\text {a }}$ | 1,537 | 3,401 | $0.45: 1.00$ |
| $2017^{\text {a }}$ | 1,835 | 3,835 | $0.48: 1.00$ |
| Mean | $\mathbf{1 , 7 1 1}$ | 3,255 | $\mathbf{0 . 5 3 : 1 . 0 0}$ |

${ }^{\text {a }}$ Includes broodstock used in the 1-male x 4-females alternative mating strategy.

### 8.4 Fecundity

The annual mean fecundity for PRH was calculated as the proportion of the total number of females spawned to the total estimated take of green eggs. The total number of green eggs is calculated after the first pick of dead eggs from the incubation trays. Fish culture staff weigh large lots of either dead or live eggs and then sub-sample the lots to calculate a mean individual egg weight. The number of eggs per lot is estimated by dividing the weight of the each egg lot by the calculated mean individual egg weight. The egg count for each lot is summed to estimate the facility egg-take. Each egg lot likely contained slightly varying amounts of interstitial water which might overestimate the egg count.

Fecundity for the 2017 broodstock averaged 3,651 eggs per female which is similar to that observed in 2015 and 2016 but less than the historical mean of 3,919 (Table 11). Pre-spawn egg loss was often observed during the electro-anesthetic and pneumatic fish euthanizing process (a physical strike to the head) and may contribute to the reduced fecundity of fish in recent years. In addition, the size and associated fecundity of Chinook salmon populations has been declining coast-wide and the reduction in fecundity at PRH may be the result of larger regional factors.

Table 11 Mean fecundity of fall Chinook salmon collected for broodstock at Priest Rapids Hatchery, Return Years 2001-2017.

| Return Year | Egg-Take | Viable Females | Fecundity/Female |
| :---: | :---: | :---: | :---: |
| 2001 | $10,750,000$ | 3,161 | 3,401 |
| 2002 | $12,180,000$ | 3,489 | 3,491 |
| 2003 | $12,814,000$ | 3,078 | 4,163 |
| 2004 | $12,753,500$ | 3,019 | 4,224 |
| 2005 | $14,085,000$ | 3,211 | 4,386 |
| 2006 | $13,511,200$ | 3,217 | 4,200 |
| 2007 a | $5,067,319$ | 1,249 | 4,057 |
| 2008 | $12,643,600$ | 3,074 | 4,113 |
| 2009 | $13,074,798$ | 2,858 | 4,575 |
| 2010 | $11,903,407$ | 3,342 | 3,562 |
| 2011 | $12,693,000$ | 3,038 | 4,178 |
| 2012 | $12,398,389$ | 3,053 | 4,061 |
| 2013 | $12,947,070$ | 3,473 | 3,728 |
| 2014 | $14,321,183$ | 3,563 | 4,019 |
| 2015 | $13,530,988$ | 3,706 | 3,651 |
| 2016 | $12,411,530$ | 3,401 | 3,649 |
| 2017 | $13,738,916$ | 3,763 | 3,651 |
| Mean | $\mathbf{1 2 , 9 6 6 , 2 8 8}$ | 3,327 | 3,919 |

Fecundities of individual females were taken from sub-samples at PRH during the spawn of 2010 through 2017 broodstock to estimate fecundity by length and age. For the 2013 through 2017 brood year data, we show comparisons between hatchery and natural origin fall Chinook salmon sampled at PRH that include fork length/fecundity, fork length/egg size (weight) and fork length, and gamete mass. For these years, we attempted to stratify the females sampled by fork length categories to obtain fecundity samples for all sizes of fish to better estimate the relationship between length and fecundity. However, the broodstock selection protocols in recent year have reduced the availability of females under 64 cm . Some fecundity data were obtained from females not used for broodstock (i.e., surplused) in order to bolster sample sizes. Therefore, comparisons between age classes are not representative of the females spawned from 2013 through 2017 broodstock populations.
M\&E staff performed the fecundity estimates on green eggs. The entire gamete mass was drained of most all ovarian fluid and weighed within 0.1 gram. Sub-sample sizes ranged between years from 60 or 100 green eggs which were counted out and weighed within 0.01 gram to estimate individual egg weight (g) for each female. Post brood year 2013, sample sizes were 100 eggs, which was determined to be sufficient based upon previous work that examined different samples sizes (Richards and Pearsons, 2014). The total fecundity of each female was estimated by dividing the weight of the total egg mass by the calculated mean individual egg weight. Each sample of the total egg mass likely contained slight varying amounts of ovarian fluid which might over estimate fecundity.

The fecundity data was pooled for return year 2010 through 2017 to provide a simple linear regression to predict fecundity based on fork-length (natural and hatchery females combined). This data shows a strong positive correlation between size and fecundity (Figure 4). The regression formula may be useful for coarse predictions of egg production for different size fish.


Figure 4 Linear relationship between fecundity and fork length for combined samples of natural and hatchery origin fall Chinook salmon spawned at Priest Rapids Hatchery, Return Years 2010-2017.
Fecundity samples collected in years 2010 through 2012 were not identified as to the origin of the females. For years 2013 through 2017, fecundity samples were taken from the broodstock at PRH to collect data associated with fecundity by size, age and origin (hatchery or natural).

Females were selected from both the PRH volunteer broodstock as well as from ponds which possessed broodstock primarily from the OLAFT and ABC. For the most part, the origin of fish during sampling was unknown. Therefore, we made a concerted effort to select females that were not adipose clipped so as to increase the chances of obtaining natural origin fish which were less common than hatchery origin fish. The origins of females sampled for fecundity were determined by hatchery marks (i.e., otoliths, adipose clips and CWTs). We make the assumption that fish not possessing any type of hatchery marks were of natural origin.
The mean fecundity by age is given in Table 12. This information is useful for forecasting potential egg-takes based on the numbers and age composition of the forecasted return.

Table 12 Mean fecundity at age for fall Chinook salmon sampled at the Priest Rapids Hatchery, Return Years 2010-2017. N = sample size and $\mathrm{SD}=1$ standard deviation.

| Return Year | Age-3 <br> Mean |  |  | $\mathbf{S D}$ | $\mathbf{N}$ | $\mathbf{M g e - 4}$ | $\mathbf{y y y}$ | Age-5 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{N}$ | Mean | SD |  |  |  |  |  |  |
| 2010 | 273 | 3,658 | 834 | 17 | 3,664 | 585 | 1 | 4,217 | -- |
| 2011 | 30 | 3,538 | 842 | 206 | 4,276 | 884 | 1 | 4,380 | -- |
| 2012 | 2 | 3,639 | 882 | 3 | 4,282 | 1089 | 0 | -- | -- |
| 2013 | 105 | 3,488 | 768 | 68 | 4,152 | 788 | 4 | 5,339 | 805 |
| 2014 | 1 | 3,358 | -- | 73 | 4,126 | 755 | 5 | 4,416 | 407 |
| 2015 | 5 | 3,169 | 382 | 53 | 3,662 | 606 | 25 | 4,746 | 691 |
| 2016 | 14 | 3,192 | 559 | 101 | 3,676 | 639 | 36 | 4,173 | 693 |
| 2017 | 0 | -- | -- | 65 | 3,754 | 689 | 31 | 4,163 | 712 |
| Mean | $\mathbf{5 4}$ | $\mathbf{3 , 4 3 5}$ | $\mathbf{7 1 1}$ | $\mathbf{7 3}$ | $\mathbf{3 , 9 4 9}$ | $\mathbf{7 5 4}$ | $\mathbf{1 3}$ | $\mathbf{4 , 4 9 1}$ | $\mathbf{6 6 2}$ |

The data collected from return years 2013 through 2017 were pooled to increase the number of samples for a given fork length. The linear relationships between fork length and variables including fecundity, mean egg weight, and total egg mass weight for natural and hatchery origin females subsampled are plotted Figure 5, Figure 6 and Figure 7. All relationships show a positive correlation with fork length. In addition, the relationships between fish size and egg data were similar for hatchery and natural origin fish.


Figure $5 \quad$ Fecundity versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 2013-2017.


Figure 6 Mean egg weight versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 20132017.


Figure $7 \quad$ Total egg mass weight versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 20132017.

### 9.0 Hatchery Rearing

### 9.1 Number of Eggs Taken

In 2017, an estimated 13,738,916 eggs were collected at PRH (Table 13). The egg-take goal for return year 2017 was $13,530,000$. The egg-take goal is calculated annually based on current program needs. This goal is established to meet the fall Chinook salmon production goals at both PRH and RSH as well as provide eggs for the Salmon in the Classroom Program.

PRH incubates approximately 8.4 million eyed eggs to produce the 7.3 million smolt release at the hatchery. Roughly an additional 4.1 million eyed eggs are needed to meet the program goal of eyed eggs delivered to Bonneville Hatchery for the 3.5 million subyearling releases from RSH.

Table 13 Number of eggs taken from fall Chinook salmon broodstock collected at Priest Rapids Hatchery, Return Years 1984-2017.

| Return Year | Number of Eggs Taken | Return Year | Number of Eggs Taken |
| :---: | :---: | :---: | :---: |
| 1984 | 10,342,000 | 2002 | 12,180,000 |
| 1985 | 10,632,000 | 2003 | 12,814,000 |
| 1986 | 22,126,100 | 2004 | 12,753,500 |
| 1987 | 24,123,000 | 2005 | 14,085,000 |
| 1988 | 16,682,000 | 2006 | 13,511,200 |
| 1989 | 13,856,500 | 2007 | 5,067,319 |
| 1990 | 9,605,000 | 2008 | 12,643,600 |
| 1991 | 6,338,000 | 2009 | 13,074,798 |
| 1992 | 11,156,400 | 2010 | 11,903,407 |
| 1993 | 14,785,000 | 2011 | 12,693,000 |
| 1994 | 16,074,600 | 2012 | 12,398,389 |
| 1995 | 17,345,900 | 2013 | 13,276,000 |
| 1996 | 14,533,500 | 2014 | 14,321,818 |
| 1997 | 17,007,000 | 2015 | 13,530,988 |
| 1998 | 13,981,300 | 2016 | 12,411,530 |
| 1999 | 16,089,600 | 2017 | 13,738,916 |
| 2000 | 15,359,500 |  |  |
| 2001 | 10,750,000 | 10 year (08-17) Mean ${ }^{1}$ | 12,999,245 |

${ }^{1}$ Began additional annual egg-takes starting in return year 2008 for the 3.5 million Ringold Springs Hatchery Program

### 9.2 Number of Acclimation Days

The 2017 brood were incubated on a combination of well and river water before being transferred to intermediate concrete raceways and then transferred to the concrete holding ponds for final acclimation before release into the Columbia River in late May and June 2018. The eggtakes for the 2017 brood were distributed into thirteen batches associated with the dates in which fish were spawned. The number of acclimation days ranged from 99 for the later egg-takes to 119 for the earlier egg-takes (Table 14).

Table 14 Number of days fall Chinook salmon fry were reared at Priest Rapids Hatchery prior to release, Brood Year 2017.

| Batch | Egg Tray to Raceway Transfer Date | Release Date | Number of Days |
| :---: | :---: | :---: | :---: |
| 1 | January 26 into Bank E | May 22 | 116 |
| 2 | January 26 into Bank E | May 22 | 116 |
| 3 | February 13 into Bank D | May 25 | 101 |
|  | February 13 into Bank C | June 12 | 119 |
| 4 | All eggs shipped to Bonneville |  |  |
| 5 | February 13 into Bank C | June 12 | 119 |
| 6 | Nearly all shipped to Bonneville |  | 100 |
| 7 | March 7 into Bank B | June 15 | 100 |
| 8 | March 7 into Bank A | June 21 | 100 |
| 9 | March 7 into Bank A | June 21 | 99 |
| 10 | March 8 into Bank A | June 21 | 99 |
| 12 | March 8 into Bank A | June 21 | 99 |
| 13 | March 8 into Bank A | June 21 | 99 |

### 9.3 Annual Releases, Tagging, and Marking

The annual release of fall Chinook salmon smolts from PRH has ranged considerably since the initial release of roughly 2.38 million smolts from the 1979 brood year to over roughly 10.30 million from the 1982 brood year (Table 15). The 2017 release goal for PRH was 7,299,504 smolts. This goal includes a recent increase in the Grant PUD mitigation from 5,000,000 to 5,599,504 combined with the ongoing USACE’s John Day mitigation of 1,700,000 smolts.

In 2018, PRH released an estimated 7,987,222 subyearling fall Chinook salmon from the 2017 broodstock (Table 16). Fish were released between May 22 and June 21.

Various mark types and rates have occurred at PRH over the years for both the Grant PUD and USACE mitigation fish. In 1976, PRH began adipose fin clipping and coded-wire tagging a portion of the juvenile fall Chinook released to determine PRH contributions to ocean and river fisheries. The smolt production at PRH associated with the USACE mitigation increased the number of adipose clipped smolts released by $\sim 1.7$ million starting with brood year 2006. The number of coded-wire tagged fish released from PRH increased to $>1.2$ million fish starting with brood year 2009 of which $\sim 600,000$ were adipose clipped. An additional 1 million adipose clipped smolts were included in the release since brood year 2011.
All PRH releases for both mitigation programs were $100 \%$ otolith marked beginning with the 2008 release. All intra-annual releases from PRH have the same annual otolith pattern, but the pattern differs between years. The eyed eggs produced for the RSH program have received an otolith mark for brood years 2010 through 2016. Otolith marking of the RSH production was discontinued beginning with the 2017 brood.

Since 1987, the U.S. Section of the Pacific Salmon Commission (PSC) has supported a coordinated project which seeks to capture and CWT 200,000 naturally produced juvenile fall Chinook salmon in the Hanford Reach (Fryer 2017). Fish are collected with seines over a ten day period between late May and early June. Fish are approximately $40-80 \mathrm{~mm}$ long at the time of capture. Recoveries from these tagged fish are used to estimate harvest exploitation rates and
interception rates for Hanford Reach natural origin fall Chinook salmon. These data have also more recently been used to estimate the number of natural origin juveniles produced in the Hanford Reach (Harnish et al. 2012, Harnish 2017).

Table 15 Number of marked, unmarked, and tagged fall Chinook salmon smolts released from Priest Rapids Hatchery, Brood Years 1977-2017.

| Brood Year | Total Released | Non Ad-Clip Released | AD/CWT | CWT Only | AD Only | PIT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 150,625 | 0 | 147,338 | 0 | 3,287 |  |
| 1978 | 153,840 | 0 | 152,532 | 0 | 1,308 |  |
| 1979 | 3,005,654 | 2,858,509 | 147,145 | 0 |  |  |
| 1980 | 4,832,591 | 4,581,054 | 251,537 | 0 |  |  |
| 1981 | 5,509,241 | 5,198,365 | 310,876 | 0 |  |  |
| 1982 | 10,296,700 | 9,888,989 | 407,711 | 0 |  |  |
| 1983 | 9,742,700 | 9,517,263 | 222,055 | 0 | 3,382 |  |
| 1984 | 6,363,000 | 6,253,240 | 106,960 | 0 | 2,800 |  |
| 1985 | 6,048,000 | 5,843,176 | 203,534 | 0 | 1,290 |  |
| 1986 | 7,709,000 | 7,506,142 | 201,843 | 0 | 1,015 |  |
| 1987 | 7,709,000 | 7,501,578 | 196,221 | 0 | 11,201 |  |
| 1988 | 5,404,550 | 5,200,080 | 201,608 | 0 | 2,862 |  |
| 1989 | 6,431,100 | 6,224,770 | 194,530 | 0 | 11,800 |  |
| 1990 | 5,333,500 | 5,134,031 | 199,469 | 0 |  |  |
| 1991 | 7,000,100 | 6,798,453 | 201,647 | 0 |  |  |
| 1992 | 7,134,159 | 6,939,537 | 194,622 | 0 |  |  |
| 1993 | 6,705,836 | 6,520,153 | 185,683 | 0 |  |  |
| 1994 | 6,702,000 | 6,526,120 | 175,880 | 0 |  | 1,500 ${ }^{\text {c }}$ |
| 1995 | 6,700,000 | 6,503,811 | 196,189 | 0 |  | $3,000^{\text {c }}$ |
| 1996 | 6,644,100 | 6,450,885 | 193,215 | 0 |  | $3,000{ }^{\text {c }}$ |
| 1997 | 6,737,600 | 6,541,351 | 196,249 | 0 |  | $3,000{ }^{\text {c }}$ |
| 1998 | 6,504,800 | 6,311,140 | 193,660 | 0 |  | $3,000{ }^{\text {c }}$ |
| 1999 | 6,856,000 | 6,651,664 | 204,336 | 0 |  | $3,000^{\text {c }}$ |
| 2000 | 6,862,550 | 6,661,771 | 200,779 | 0 |  | $3,000{ }^{\text {c }}$ |
| 2001 | 6,779,035 | 6,559,109 | 219,926 | 0 |  | $3,000{ }^{\text {c }}$ |
| 2002 | 6,777,605 | 6,422,232 | 355,373 | 0 |  | $3,000^{\text {c }}$ |
| 2003 | 6,814,560 | 6,415,444 | 399,116 | 0 |  | $3,000^{\text {c }}$ |
| 2004 | 6,599,838 | 6,399,766 | 200,072 | 0 |  | $3,000{ }^{\text {c }}$ |
| 2005 | 6,876,290 | 6,676,845 | 199,445 | 0 |  | $3,000^{\text {c }}$ |
| 2006 | 6,743,101 | 4,912,487 | 202,000 | 0 | 1,628,614 | $3,000^{\text {c }}$ |
| $2007{ }^{\text {a }}$ | 4,548,307 | 4,344,926 | 202,568 | 0 | $813^{\text {b }}$ | $3,000{ }^{\text {c }}$ |
| $2008{ }^{\text {a }}$ | 6,788,314 | 4,850,844 | 218,082 | 0 | 1,719,388 | 2,994 ${ }^{\text {c }}$ |
| $2009{ }^{\text {a }}$ | 6,776,651 | 3,413,334 | 619,568 | 1,026,561 | 1,717,188 | 1,995 ${ }^{\text {c }}$ |
| $2010^{\text {a }}$ | 6,798,390 | 3,383,859 | 602,580 | 1,108,990 | 1,702,961 | 3,000 ${ }^{\text {c }}$ |
| $2011^{\text {a }}$ | 7,056,948 | 3,094,666 | 595,608 | 598,031 | 2,768,643 | 42,844 ${ }^{\text {c }}$ |
| $2012{ }^{\text {a }}$ | 6,822,861 | 2,905,694 | 603,930 | 601,009 | 2,712,228 | 42,908 ${ }^{\text {c }}$ |
| $2013{ }^{\text {a }}$ | 7,267,248 | 3,347,417 | 603,417 | 603,439 | 2,712,975 | $42,908^{\text {c }}$ |
| $2014{ }^{\text {a }}$ | 7,039,544 | 3,125,734 | 600,688 | 600,730 | 2,712,392 | $42,621^{\text {c }}$ |
| $2015{ }^{\text {a }}$ | 7,242,054 | 3,317,992 | 602,116 | 601,770 | 2,720,176 | $42,999{ }^{\text {d }}$ |
| $2016{ }^{\text {a }}$ | 7,006,252 | 3,045,689 | 603,539 | 603,864 | 2,710,302 | 42,858 ${ }^{\text {d }}$ |
| $2017{ }^{\text {a }}$ | 7,987,222 | 4,067,088 | 602,725 | 607,287 | 2,710,121 | 42,978 ${ }^{\text {c }}$ |

${ }^{\text {a }}$ Entire release was otolith marked
${ }^{\mathrm{b}}$ Low returns to PRH precluded the production of the USACE adipose clipped release.
${ }^{\text {c }}$ PIT tagged fish were included within the other mark group totals
${ }^{\text {d }}$ PIT tagged fish were not adipose clipped and reported as a unique group.

### 9.4 Fish Size and Condition of Release

The data associated with fish size and condition at release from PRH prior to brood year 2013 was obtained from the hatchery staff. The mean fish weight was obtained by weighing groups of roughly 300 fish sampled from each pond to the nearest gram and then dividing the group weight by the total number of fish weighed. The fork length of each fish from the group weight was measured to the nearest millimeter to calculate mean length and coefficient of variation. Samples from each of the rearing ponds were taken the day of release. The results were pooled to provide mean estimates for the facility as a whole. The size and condition data for the 2013 through 2017 broods were collected by M\&E staff the day prior to or day of release for each pond. We attempted to collect representative samples by capturing multiple groups of fish with a cast net from the lower, middle, and upper third of the rearing pond. Each fish sampled was individually weighed to the nearest 0.1 gram and measured for fork length to the nearest millimeter. The results were pooled to provide mean estimates for the facility as a whole.

The goal for PRH is to release fall Chinook salmon smolts at 50 fish per pound. At release, the smolts from the 2017 brood averaged 49 fish per pound with a mean fork length of 89 mm , and a mean CV of 6.1 (Table 16). For brood years 1991 through 2016, smolts released from PRH have averaged 48 fish per pound with a mean fork length of 95 and a mean CV of 7.3.

Table 16 Mean length ( $\mathrm{FL}, \mathrm{mm}$ ), weight ( g and fish/pound), and coefficient of variations (CV) of fall Chinook smolts released from Priest Rapids Hatchery, Brood Years 1991-2016.

| Brood year | Release Year | Fork Length (mm) |  | Mean Weight |  | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean | CV | Grams (g) | Fish/pound |  |
| 1991 | 1992 | 93 | 8.7 | 8.3 | 55 | 1,500 |
| 1992 | 1993 | 92 | 8.6 | 8.3 | 54 | 1,500 |
| 1993 | 1994 | 95 | 6.9 | 9.3 | 49 | 1,500 |
| 1994 | 1995 | 96 | 6.7 | 9.7 | 47 | 1,500 |
| 1995 | 1996 | 97 | 6.6 | 10 | 45 | 1,500 |
| 1996 | 1997 | 95 | 11 | 8.7 | 52 | 1,500 |
| 1997 | 1998 | 103 | 8.9 | 10.1 | 45 | 1,500 |
| 1998 | 1999 | 95 | 6.5 | 9.6 | 48 | 1,500 |
| 1999 | 2000 | 93 | 6.6 | 8.9 | 51 | 1,500 |
| 2000 | 2001 | 97 | 6.3 | 10.2 | 45 | 1,500 |
| 2001 | 2002 | 96 | 6.9 | 10.1 | 45 | 1,500 |
| 2002 | 2003 | 95 | 6.9 | 9.5 | 48 | 1,500 |
| 2003 | 2004 | 96 | 6.8 | 9.6 | 48 | 1,500 |
| 2004 | 2005 | 95 | 5.9 | 9.4 | 48 | 1,500 |
| 2005 | 2006 | 98 | 6.3 | 10.1 | 45 | 1,500 |
| 2006 | 2007 | 98 | 7.0 | 9.9 | 46 | 1,500 |
| 2007 | 2008 | 101 | 8.3 | 10.2 | 45 | 1,200 |
| 2008 | 2009 | 94 | 6.7 | 9.3 | 49 | 1,500 |
| 2009 | 2010 | 94 | 7.3 | 9.2 | 49 | 1,500 |
| 2010 | 2011 | 92 | 9.1 | 9.7 | 47 | 1,500 |
| 2011 | 2012 | 94 | 7.1 | 9.2 | 49 | 1,500 |
| 2012 | 2013 | 95 | 7.6 | 9.7 | 47 | 1,500 |
| 2013 | 2014 | 92 | 8.4 | 9.0 | 50 | 648 |
| 2014 | 2015 | 91 | 6.6 | 8.7 | 52 | 1,728 |
| 2015 | 2016 | 92 | 6.1 | 9.3 | 49 | 1,595 |
| 2016 | 2017 | 89 | 6.1 | 9.3 | 49 | 1,788 |
| 2017 | 2018 | 91 | 6.6 | 9.2 | 50 | 1,633 |
| Mean |  | 95 | 7.3 | 9.4 | 48 | 1,485 |

### 9.5 Survival Estimates

The survival proportion ( $\mathrm{P} \wedge$ ) for fertilized egg to juvenile release for brood year 2017 was 0.928 which is higher than the historic mean of 0.867 (Table 17). The green egg to eyed egg stage is the most critical life stage at PRH during incubation/juvenile rearing because the greatest level of loss annually occurs at this stage. The green egg to eyed egg survival P^ for brood year 2017 was 0.917 which is similar to the historical mean of 0.902 .

In 2017, survival $\mathrm{P}^{\wedge}$ of fish ponded for broodstock was 0.821 which is slightly lower than the historic mean of 0.846 . The trapping operations in 2014 through 2017 were carried out in a manner which generally reduced fish densities in the trap; possibly resulting in reduced ponding mortality.

Table 17 Hatchery life stage survival ( $\mathrm{P}^{\wedge}$ ) for fall Chinook salmon at Priest Rapids Hatchery, Brood Years 1989-2017.

| Brood year | PRH Volunteers Ponded to Spawned |  |  |  | Unfertilized to Eyed Egg | Eyed egg to Ponding | Ponding to Release | Fertilized Egg to Release |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Female | Male | Jack | Total |  |  |  |  |
| 1989 |  |  |  | 0.919 | 0.866 | 0.976 | 0.950 | 0.821 |
| 1990 |  |  |  | 0.947 | 0.869 | 0.996 | 0.984 | 0.852 |
| 1991 |  |  |  | 0.973 | 0.948 | 0.993 | 0.998 | 0.922 |
| 1992 |  |  |  | 0.952 | 0.945 | 0.991 | 0.965 | 0.901 |
| 1993 |  |  |  | 0.917 | 0.941 | 0.984 | 0.974 | 0.902 |
| 1994 |  |  |  | 0.710 | 0.935 | 0.985 | 0.953 | 0.878 |
| 1995 |  |  |  | 0.897 | 0.914 | 0.980 | 0.962 | 0.862 |
| 1996 |  |  |  | 0.908 | 0.924 | 0.997 | 0.983 | 0.897 |
| 1997 |  |  |  | 0.900 | 0.915 | 0.996 | 0.970 | 0.790 |
| 1998 |  |  |  | 0.834 | 0.914 | 0.998 | 0.970 | 0.884 |
| 1999 |  |  |  | 0.759 | 0.897 | 0.997 | 0.995 | 0.888 |
| 2000 |  |  |  | 0.868 | 0.898 | 0.995 | 0.985 | 0.884 |
| 2001 | 0.776 | 0.732 | 0.665 | 0.757 | 0.886 | 0.994 | 0.975 | 0.859 |
| 2002 | 0.835 | 0.829 | 0.705 | 0.828 | 0.880 | 0.995 | 0.979 | 0.858 |
| 2003 | 0.893 | 0.817 | 0.698 | 0.858 | 0.882 | 0.989 | 0.989 | 0.868 |
| 2004 | 0.958 | 0.915 | 0.646 | 0.845 | 0.881 | 0.975 | 0.985 | 0.846 |
| 2005 | 0.890 | 0.890 | 0.782 | 0.886 | 0.914 | 0.976 | 0.991 | 0.884 |
| 2006 | 0.918 | 0.924 | 0.695 | 0.913 | 0.897 | 0.975 | 0.981 | 0.859 |
| 2007 | 0.967 | 0.748 | 0.642 | 0.861 | 0.858 | 0.996 | 0.981 | 0.898 |
| 2008 | 0.943 | 0.896 | 0.877 | 0.924 | 0.902 | 0.973 | 0.877 | 0.877 |
| 2009 | 0.848 | 0.901 | 0.916 | 0.864 | 0.912 | 0.977 | 0.891 | 0.891 |
| 2010 | 0.803 | 0.831 | 0.803 | 0.809 | 0.913 | 0.985 | 0.977 | 0.841 |
| 2011 | 0.611 | 0.847 | 0.737 | 0.679 | 0.903 | 0.985 | 0.985 | 0.875 |
| 2012 | 0.643 | 0.786 | 0.630 | 0.688 | 0.873 | 0.970 | 0.962 | 0.787 |
| 2013 | 0.698 | 0.660 | 0.333 | 0.684 | 0.884 | 0.983 | 0.951 | 0.806 |
| 2014 | 0.830 | 0.880 | N/A | 0.847 | 0.865 | 0.933 | 0.978 | 0.913 |
| 2015 | 0.841 | 0.810 | N/A | 0.830 | 0.917 | 0.934 | 0.985 | 0.919 |
| 2016 | 0.873 | 0.782 | N/A | 0.843 | 0.899 | 0.825 | 0.989 | 0.816 |
| 2017 | 0.820 | 0.824 | N/A | 0.821 | 0.917 | 0.942 | 0.985 | 0.928 |
| Mean | 0.832 | 0.828 | 0.702 | 0.846 | 0.902 | 0.976 | 0.971 | 0.867 |

### 9.6 Juvenile PIT Tag Detections at the Priest Rapids Hatchery Array

Roughly 3,000 sub-yearlings at PRH were annually PIT tagged and released from PRH for brood years 1995 through 2010 to assess timing, migration speed, and juvenile survival from PRH to McNary Dam. The analysis for these measures is reported annually by the Fish Passage Center and can be found at www.fpc.org/documents/FPC_memos.html.
Beginning with the 2011 brood, approximately 40,000 additional juveniles were annually PIT tagged and released to bolster the data collected for estimation of juvenile abundance at release and adult straying. These tags can also be used to estimate adult migration timing, conversion rates from Bonneville Dam to McNary Dam to PRH, smolt to adult survival rates, as well as fallback and re-ascension estimates at McNary, Ice Harbor, and Priest Rapids dams. The annual
detection rates are given in Table 18. Prior to the 2012 release (brood year 2011), a PIT tag array consisting of six antennas was installed in the hatchery discharge channel to detect both juvenile out-migrants and adult returns. The detection rates reported below account for the relatively few shed PIT tags found in the rearing raceways. Prior to the release of the 2016 brood, the mortalities routinely recovered from the rearing ponds were not scanned for PIT tags. This prohibits us from knowing the actual total number of PIT tagged fish released. Hence, the overall proportion of released PIT tagged fish detected would likely be higher than reported if we knew the actual number of live PIT tagged fish that left the ponds.

The overall detection rate for the releases of the 2011 brood year was $70.4 \%$. The releases occurred over an eight day period, with only two days of consecutive releases. Detection rates for the 2011 brood year release may have been reduced as a result of the array being inundated by high river elevations during portions of releases. The overall detection rate for the 2012 brood year was $3.4 \%$. The low detection rates were likely due to force releasing all of the smolts in four consecutive days which appears to have overwhelmed the PIT tag detection equipment. The restricted release period was necessitated by the construction schedule of the new hatchery.
A concerted effort was made during both the 2013 and 2014 brood year releases to improve the PIT tag detection efficiency at the PRH array. First, the automatic upload function of the array was discontinued to reduce the usage demand on the system's processor. Secondly, the five releases from the hatchery were conducted over a fourteen day period beginning on June 12 to spread out over time the number of PIT tags passing the array. This was managed by pulling the individual weir boards for each pond over a two day period. The percentage of PIT tagged subyearlings detected for the 2013 and 2014 brood years were $92.9 \%$ and $94.5 \%$, respectively.
The releases of the 2015 brood occurred every two days between June 16 and June 24, 2016 to accommodate a day versus night release evaluation. During the evaluation, all weir boards for a given pond where incrementally pulled over an eight hour period on the date of release. Overall, $84.3 \%$ of the PIT tagged subyearlings were detected. The detected rate between release groups varied from $33.6 \%$ to $97.0 \%$. These values are lower than the previous two years. It's possible that forced releases over an eight hour period may have resulted in high rates of tag collision at the array resulting in poor detection efficiency.

The releases of the 2016 and 2017 broods were initiated at 9PM for each pond. All weir boards were pulled by 3AM. Releases occurred irregularly between May 23 and June 20. For both release years, we anticipated river flows during May and June to exceed 340kcfs which results in the inundation of the PRH array. A temporary two antenna array was installed at a higher elevation near the upper end of the discharge channel to complement the PRH array. The overall detection rate for the 2016 brood was $95.4 \%$ for the combined release of all ponds, ranging from $89.7 \%$ to $97.5 \%$. The overall detection rate for the 2017 brood was $86.8 \%$, ranging from $83.5 \%$ to $93.0 \%$.

Table 18 Number of sub-yearlings PIT tagged, mark, and release dates, and the number of unique tags detected at the array in the Priest Rapids discharge channel, Brood Years 2011-2017.

| Brood <br> Year | Tag File | Tagging Date | Release Date | \# Tagged | \# of Tags <br> Recovered from Facility Mortalities | \# of Unique Detections | \% <br> Detected |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | CSM12114.A01 | 4/23/2012 | 6/20/2012 | 9937 | No Data | 6,277 | 63.2 |
| 2011 | CSM12114.A04 | 4/23/2012 | 6/14/2012 | 9948 | No Data | 6,674 | 67.1 |
| 2011 | CSM12114.A03 | 4/24/2012 | 6/15/2012 | 9997 | No Data | 6,963 | 69.7 |
| 2011 | CSM12115.A02 | 4/24/2012 | 6/16/2012 | 9967 | No Data | 8,115 | 81.4 |
| 2011 | CSM12115.A02 | 5/30/2012 | 6/20/2012 | 1000 | No Data | 499 | 49.9 |
| 2011 | SMP12151.PR2 | 5/30/2012 | 6/16/2012 | 998 | No Data | 806 | 80.8 |
| 2011 | SMP12152.PR3 | 5/31/2012 | 6/12/2012 | 996 | No Data | 810 | 81.3 |
|  |  |  | Totals | 42,844 | N/A | 30,144 | 70.4 |
| 2012 | CSM13143.A06 | 5/23/2013 | 6/14/2013 | 9,982 | No Data | 317 | 3.2 |
| 2012 | CSM13143.A07 | 5/23/2013 | 6/13/2013 | 9,983 | No Data | 267 | 2.7 |
| 2012 | CSM13144.A08 | 5/24/2013 | 6/12/2013 | 9,974 | No Data | 335 | 3.4 |
| 2012 | CSM13144.A09 | 5/24/2013 | 6/15/2013 | 9,977 | No Data | 325 | 3.3 |
| 2012 | SMP13149.PR1 | 5/29/2013 | 6/15/2013 | 997 | No Data | 131 | 13.1 |
| 2012 | SMP13149.PR2 | 5/29/2013 | 6/14/2013 | 996 | No Data | 33 | 3.3 |
| 2012 | SMP13150.PR3 | 5/30/2013 | 6/12/2013 | 999 | No Data | 48 | 4.9 |
|  |  |  | Totals | 42,908 | N/A/ | 1,456 | 3.4 |
| 2013 | CSM14148.PRA | 5/28/2014 | 6/25/2014 | 7,994 | 21 | 7,215 | 90.5 |
| 2013 | CSM14148.PRB | 5/28/2014 | 6/23/2014 | 7,998 | 14 | 7,215 | 92.5 |
| 2013 | CSM14149.PRC | 5/29/2014 | 6/18/2014 | 7,996 | 11 | 7,443 | 93.2 |
| 2013 | CSM14149.PRD | 5/29/2014 | 6/16/2014 | 7,993 | 6 | 7,662 | 95.9 |
| 2013 | CSM14149.PRE | 5/29/2014 | 6/12/2014 | 7,998 | 7 | 7,407 | 92.7 |
| 2013 | SMP14148.PR1 | 5/29/2014 | 6/25/2014 | 996 | 0 | 914 | 91.8 |
| 2013 | SMP14148.PR2 | 5/29/2014 | 6/18/2014 | 994 | 0 | 927 | 93.3 |
| 2013 | SMP14149.PR3 | 5/30/2014 | 6/12/2014 | 998 | 0 | 951 | 95.3 |
|  |  |  | Total | 42,967 | 59 | 39,908 | 92.9 |
| 2014 | CSM15147.PRE | 5/27/2015 | 6/12/2015 | 7,999 | 169 | 7,438 | 95.0 |
| 2014 | CSM15147.PRD | 5/27/2015 | 6/15/2015 | 7,996 | 39 | 7,685 | 96.6 |
| 2014 | CSM15147.PRC | 5/27/2015 | 6/18/2015 | 7,996 | 63 | 7,524 | 94.8 |
| 2014 | CSM15147.PRB | 5/28/2015 | 6/22/2015 | 7,998 | 50 | 7,696 | 96.8 |
| 2014 | CSM15147.PRA | 5/28/2015 | 6/25/2015 | 7,994 | 31 | 7,447 | 93.5 |
| 2014 | SMP15140.PR1 | 5/20/2015 | 6/25/2015 | 993 | 0 | 940 | 94.7 |
| 2014 | SMP15140.PR2 | 5/20/2015 | 6/18/2015 | 998 | 0 | 946 | 94.8 |
| 2014 | SMP15141.PR3 | 5/21/2015 | 6/12/2015 | 999 | 0 | 935 | 93.6 |
|  |  |  | Total | 42,973 | 352 | 40,611 | 95.3 |
| 2015 | CSM16153.PRE | 6/01/2016 | 6/16/2016 | 7,996 | 13 | 6,032 | 75.6 |
| 2015 | CSM16153.PRD | 6/01/2016 | 6/18/2016 | 7,998 | 224 | 7,537 | 97.0 |
| 2015 | CSM16153.PRC | 6/01/2016 | 6/20/2016 | 7,985 | 137 | 6,777 | 86.4 |
| 2015 | CSM16154.PRB | 6/02/2016 | 6/22/2016 | 7,993 | 13 | 7,136 | 89.4 |
| 2015 | CSM16154.PRA | 6/02/2016 | 6/24/2016 | 7,990 | 26 | 6,590 | 82.7 |
| 2015 | SMP16153.PR1 | 6/01/2016 | 6/24/2016 | 995 | 88 | 513 | 56.6 |
| 2015 | SMP16153.PR2 | 6/01/2016 | 6/20/2016 | 998 | 5 | 795 | 80.1 |
| 2015 | SMP16154.PR3 | 6/02/2016 | 6/16/2016 | 1001 | 109 | 300 | 33.6 |
|  |  |  | Totals | 42,956 | 615 | 35,680 | 84.3 |
| 2016 | BMI17129.PRE | 5/09/2017 | 5/23/2017 | 7,996 | 18 | 7,279 | 91.2 |
| 2016 | BMI17129.PRD | 5/09/2017 | 5/25/2017 | 7,998 | 7 | 7,790 | 97.5 |


| Brood <br> Year | Tag File | Tagging Date | Release <br> Date | \# Tagged | \# of Tags Recovered from Facility Mortalities | \# of Unique Detections | \% <br> Detected |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | BMI17143.PRC | 5/23/2017 | 6/09/2017 | 7,981 | 32 | 7,714 | 97.0 |
| 2016 | BMI17143.PRB | 5/23/2017 | 6/12/2017 | 7,995 | 24 | 7,633 | 95.8 |
| 2016 | BMI17144.PRA | 5/24/2017 | 6/19/2017 | 7,995 | 46 | 7,633 | 96.0 |
| 2016 | SMP17128.PR1 | 5/08/2017 | 5/23/2017 | 600 | 0 | 538 | 89.7 |
| 2016 | SMP17129.PR2 | 5/09/2017 | 5/25/2017 | 600 | 0 | 579 | 96.5 |
| 2016 | SMP17144.PR3 | 5/24/2017 | 6/09/2017 | 598 | 0 | 568 | 95.0 |
| 2016 | SMP17144.PR4 | 5/24/2017 | 6/12/2017 | 601 | 0 | 581 | 96.7 |
| 2016 | SMP17144.PR5 | 5/24/2017 | 6/19/2017 | 600 | 2 | 570 | 95.3 |
|  |  |  | Totals | 42,964 | 129 | 40,885 | 95.4 |
| 2017 | BMI2018128PRE | 5/08/2018 | 5/23/2018 | 7,999 | 24 | 6,681 | 83.5 |
| 2017 | BMI2018128PRD | 5/08/2018 | 5/25/2018 | 7,997 | 11 | 6,957 | 87.0 |
| 2017 | BMI2018149PRC | 5/29/2018 | 6/11/2018 | 7,997 | 6 | 7,435 | 93.0 |
| 2017 | BMI2018150PRB | 5/30/2018 | 6/14/2018 | 7,997 | 15 | 6,916 | 86.5 |
| 2017 | BMI2018151PRA | 5/31/2018 | 6/20/2018 | 7,994 | 16 | 6,725 | 84.1 |
| 2017 | SMP2018129002 | 5/09/2018 | 5/23/2018 | 599 | 4 | 508 | 84.8 |
| 2017 | SMP2018129001 | 5/09/2018 | 5/25/2018 | 597 | 1 | 524 | 87.8 |
| 2017 | SMP2018149PR3 | 5/29/2018 | 6/11/2018 | 599 | 1 | 556 | 92.8 |
| 2017 | SMP2018149PR4 | 5/29/2018 | 6/14/2018 | 597 | 0 | 510 | 85.4 |
| 2017 | SMP2018150PR5 | 5/30/2018 | 6/20/2018 | 597 | 0 | 505 | 84.6 |
|  |  |  | Totals | 42,973 | 78 | 37,317 | 86.8 |

### 10.0 Adult Fish Pathogen Monitoring

At spawning, a portion of the adult fall Chinook broodstock are sampled for infectious hemotopoietic necrosis virus (IHNV), infectious pancreatic necrosis virus (IPNV), viral hemorrhagic septicemia virus (VHSV), paramyxovirus, aquaroviruses, as well as Renibacterium salmoninarum, the causative agent for bacterial kidney disease (BKD). Viral and bacterial screening included sampling the ovarian fluid and kidney/spleen for pathogens. All results of viral testing in since 1991 were classified as negative (Table 19).

Table 19 Viral inspections of fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 1991-2017.

| Year | Date(s) | Stock | Life stage | Ovarian Fluid | Kidney/Spleen | Results |
| :---: | ---: | ---: | :---: | :---: | :---: | :---: |
| 1991 | 28-Oct, 4, 13-Nov | Priest Rapids | Adult | 150 | 60 | Negative |
| 1992 | 2,9-Nov | Priest Rapids | Adult | 150 | 60 | Negative |
| 1993 | 25-Oct, 1-Nov | Priest Rapids | Adult | 150 | 60 | Negative |
| 1994 | 7-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1995 | 9,13,19,21-Nov | Priest Rapids | Adult | 160 | 160 | Negative |
| 1996 | 17-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1997 | 17-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1998 | 16-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1999 | 8-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2000 | 13-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2001 | 13-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2002 | 13-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2003 | 17-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2004 | 8-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2005 | 14-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2006 | 6-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2007 | 5-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2008 | 3-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2009 | 2-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2010 | 15-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2011 | 7,14, 21-Nov | Priest Rapids | Adult | 180 | 180 | Negative |
| 2012 | 5-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2013 | 18-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2014 | 18-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2015 | 11-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2016 | 8-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2017 | 1,3,8-Nov | Priest Rapids | Adult | 268 | Negative |  |

Annual testing for BKD was initiated with the 2008 broodstock to address concerns associated with shipping eyed-eggs to Bonneville Hatchery for the USACE RSH production. The risk of BKD was assayed using the enzyme linked immunosorbent assay (ELISA) for R. salmoninarum antigen (Elliot 2012). Differences in normal screening for BKD occurred at PRH during 2017. The fish health staff from Oregon Department of Fish and Wildlife tested 268 adults originating from PRH and incorporated into the Umatilla-John Day Mitigation Program for BKD. These fish were trapped at PRH and then transported and spawned at RSH in early November. Adult broodstock BKD monitoring in 2017 indicated that 267 of the 268 ( $99.6 \%$ ) females tested had ELISA values less than an optical density of 0.10 (Below Low); 1 of the 268 samples was in the Low category between 0.10 and 0.199 (Table 20). Since 2008, tests have shown very low percentages of fish with values greater and 0.10.

Table 20 ELISA test results to determine risk of bacterial kidney disease of adult female fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 2008-2017.

| Year | Stock | $\mathbf{N}$ | \%Below-Low | \% Low | \% Mod | \% High |
| :--- | :--- | ---: | :---: | :---: | :---: | :---: |
| 2008 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2009 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2010 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2011 | Priest Rapids | 135 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2012 | Priest Rapids | 60 | $98.3 \%$ | $0.0 \%$ | $1.7 \%$ | $0.0 \%$ |
| 2013 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2014 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2015 | Priest Rapids | 60 | $98.3 \%$ | $1.7 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2016 | Priest Rapids | 60 | $98.3 \%$ | $1.7 \%$ | $0.0 \%$ | $0.0 \%$ |
| $2017^{\text {a }}$ | Priest Rapids | 268 | $99.6 \%$ | $0.4 \%$ | $0.0 \%$ | $0.0 \%$ |

### 10.1 Juvenile Fish Health Inspections

Juvenile fish are inspected for the presence of pathogens and other conditions on a monthly basis following ponding (AFS-FHS 2014). The 2017 brood year juveniles were generally healthy throughout the rearing period with the exception of fish reared in Raceway Bank C (Table 21). The presence of bacterial gill disease appeared in several raceway ponds in Bank C. This resulted in periods of elevated mortalities which prompted treatments with minerally balanced granulated solar salts. Inspection results for brood years 1995 through 2009 are provided in Appendix C.

Table 21 Juvenile fish health inspections for Priest Rapids Hatchery fall Chinook salmon, Brood Years 2006-2017.

| Date | Stock | Brood <br> Year | Condition |
| :--- | :--- | :--- | :--- |
| 18-Feb-10 | Priest Rapids | 2009 | Coagulated Yolk Syndrome observed in some fish sampled |
| 1-Apr-10 | Priest Rapids | 2009 | Healthy |
| 19-May-10 | Priest Rapids | 2009 | Healthy |
| 25-Mar-11 | Priest Rapids | 2010 | Healthy |
| 18-Apr-11 | Priest Rapids | 2010 | Healthy |
| 06-Jun-11 | Priest Rapids | 2010 | Healthy |
| 01-Mar-12 | Priest Rapids | 2011 | Healthy |
| 26-Apr-12 | Priest Rapids | 2011 | Healthy |
| 24-May-12 | Priest Rapids | 2011 | Healthy |
| 11-Feb-13 | Priest Rapids | 2012 | Healthy |
| 3-Mar-13 | Priest Rapids | 2012 | Healthy |
| 29-Apr-13 | Priest Rapids | 2012 | Healthy |
| 28-May-13 | Priest Rapids | 2012 | Healthy |
| 27-Mar-14 | Priest Rapids | 2013 | Dropout Syndrome present |
| 23-Apr-14 | Priest Rapids | 2013 | Dropout Syndrome present |
| 29-May-14 | Priest Rapids | 2013 | Healthy |
| 26-Feb-15 | Priest Rapids | 2014 | Coagulated Yolk Syndrome observed in some fish sampled |
| 26-Mar-15 | Priest Rapids | 2014 | Healthy |
| 21-Apr-15 | Priest Rapids | 2014 | Healthy |
| 28-May-15 | Priest Rapids | 2014 | Healthy |
| 22-June-15 | Priest Rapids | 2014 | Columnaris present in some fish sampled from Channel Pond B. |
| 24-Feb-16 | Priest Rapids | 2015 | Healthy |
| 15-Mar-16 | Priest Rapids | 2015 | Coagulated Yolk Syndrome observed in some fish sampled |
| 15-June-16 | Priest Rapids | 2015 | Mild Ich infection but healthy and ready for release |
| 24-Feb-17 | Priest Rapids | 2016 | Presence of bacterial gill disease in Raceway Bank D and E |
| 21-Mar-17 | Priest Rapids | 2016 | Presence of bacterial gill disease in Raceway Pond B2 |
| 6-June-17 | Priest Rapids | 2016 | Mild Ich infection in Channel Ponds A, B, C |
| 21-Mar-18 | Priest Rapids | 2017 | Healthy |
| 19-Apr-18 | Priest Rapids | 2017 | Bacterial gill dieses present in Raceway Pond C4 |
| 7-May-18 | Priest Rapids | 2017 | Bacterial gill dieses present in Raceway Ponds C2 and C3 |
| 17-May-18 | Priest Rapids | 2017 | Re-examine Raceway Ponds C2 and C3 found fish healthy |
| 17-May-18 | Priest Rapids | 2017 | Pre-release examine Raceway Banks D and E found fish healthy <br> 6-June-18 |
| Priest Rapids | 2017 | Pre-release examine of Raceway Banks A and B found fish healthy |  |

### 11.0 Redd Survey

Fall Chinook salmon redd surveys were performed in the Hanford Reach during 2017 by staff with Mission Support Alliance under contract with the United States Department of Energy. WDFW M\&E staff performed fall Chinook salmon redd surveys in the PRH discharge channel during 2017.

### 11.1 Hanford Reach Aerial Redd Counts

Aerial redd counts in the Hanford Reach were performed by Mission Support Alliance on October 23, November 6, and 19 during 2017 (Nugent 2017). Redd counts should be considered an index of the total number of redds in the Hanford Reach. Redds may not be visible during flights due to wind, turbidity, ambient light, and depth. The first two surveys occurred on a Monday. The last survey occurred on Sunday when outflows at Priest Rapids Dam were lowered to nearly 47 kcfs in conjunction with the Vernita Bar Settlement Agreement surveys performed by Grant PUD and WDFW. It is reported that viewing conditions during the surveys were generally fair on the last flight; high clouds and light wind (USDOE In Press). The peak fall Chinook Salmon redd count for the Hanford Reach in 2017 was 8,646 (Table 22).

Table 22 Summary of fall Chinook salmon peak redd counts for the 1948-2017 aerial surveys in the Hanford Reach, Columbia River.

| Year | Redds | Year | Redds | Year | Redds | Year | Redds | Year | Redds |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1948 | 787 | 1963 | 1,254 | 1978 | 3,028 | 1993 | 2,863 | 2008 | 5,588 |
| 1949 | 313 | 1964 | 1,477 | 1979 | 2,983 | 1994 | 5,619 | 2009 | 4,996 |
| 1950 | 265 | 1965 | 1,789 | 1980 | 1,487 | 1995 | 3,136 | 2010 | 8,817 |
| 1951 | 297 | 1966 | 3,101 | 1981 | 4,866 | 1996 | 7,618 | 2011 | 8,915 |
| 1952 | 528 | 1967 | 3,267 | 1982 | 4,988 | 1997 | 7,600 | 2012 | 8,368 |
| 1953 | 139 | 1968 | 3,560 | 1983 | 5,290 | 1998 | 5,368 | 2013 | 17,398 |
| 1954 | 160 | 1969 | 4,508 | 1984 | 7,310 | 1999 | 6,068 | 2014 | 15,951 |
| 1955 | 60 | 1970 | 3,813 | 1985 | 7,645 | 2000 | 5,507 | 2015 | 20,678 |
| 1956 | 75 | 1971 | 3,600 | 1986 | 8,291 | 2001 | 6,248 | 2016 | 13,268 |
| 1957 | 525 | 1972 | 876 | 1987 | 8,616 | 2002 | 8,083 | 2017 | 8,646 |
| 1958 | 798 | 1973 | 2,965 | 1988 | 8,475 | 2003 | 9,465 |  |  |
| 1959 | 281 | 1974 | 728 | 1989 | 8,834 | 2004 | 8,468 |  |  |
| 1960 | 258 | 1975 | 2,683 | 1990 | 6,506 | 2005 | 7,891 |  |  |
| 1961 | 828 | 1976 | 1,951 | 1991 | 4,939 | 2006 | 6,508 |  |  |
| 1962 | 1,051 | 1977 | 3,240 | 1992 | 4,926 | 2007 | 4,023 |  |  |
| Mean (2008-2017) |  |  |  |  |  |  |  |  | 10,742 |

### 11.2 Redd Distribution

The main spawning areas observed during the 2017 counts were located near Vernita Bar and among Islands 4-6 (Table 23 \& Figure 8). Historical redd counts by location from 2001 through 2017 are included in Appendix D of this report.

Table 23 Number of all Chinook salmon redds counted in difference reaches on the Hanford Reach area of the Columbia River during October 2017 through November 2017 aerial redd counts. (Data provided by Mission Support Alliance).

| General Location | Start <br> KM | End <br> KM | Total <br> Length | 23-Oct | 6-Nov | 19-Nov | Max <br> Count | Avg. Redd <br> Per River <br> KM |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Islands 17-21 | 545 | 558 | 13 | 0 | 2 | 0 | 2 | 0 |
| Islands 11-16 | 558 | 573 | 15 | 11 | 120 | 280 | 280 | 19 |
| Islands 8-10 | 587 | 593 | 6 | 19 | 864 | 900 | 900 | 150 |
| Near Island 7 | 593 | 594 | 1 | 0 | 22 | 670 | 670 | 670 |
| Island 6 (lower half) | 594 | 599 | 5 | 5 | 680 | 900 | 900 | 180 |
| Island 4, 5 and upper 6 | 599 | 602 | 3 | 11 | 418 | 911 | 911 | 304 |
| Near Island 3 | 602 | 604 | 2 | 0 | 40 | 500 | 500 | 250 |
| Near Island 2 | 604 | 606 | 2 | 0 | 281 | 790 | 790 | 395 |
| Near Island 1 | 606 | 608 | 2 | 2 | 145 | 330 | 330 | 165 |
| Near Coyote Rapids | 608 | 619 | 11 | 0 | 0 | 80 | 80 | 16 |
| Midway (China Bar) | 620 | 630 | 10 | 4 | 14 | 75 | 75 | 38 |
| Near Vernita Bar | 630 | 635 | 5 | 85 | 1,310 | 3,200 | 3,200 | 640 |
| Near Priest Rapids Dam | 635 | 638 | 3 | 0 | 0 | 10 | 10 | 3 |
| Total | -- | -- | -- | $\mathbf{1 3 7}$ | $\mathbf{3 , 8 9 6}$ | $\mathbf{8 , 6 4 6}$ | $\mathbf{8 , 6 4 6}$ | -- |



Figure 8 Distribution of fall Chinook salmon redd counts by location for the 2017 aerial surveys in the Hanford Reach, Columbia River. (Data provided by Mission Support Alliance).

### 11.3 Spawn Timing

Based on aerial redd counts and Vernita Bar spawning ground surveys, fall Chinook salmon spawning in the Hanford Reach during 2017 began in late October and ended in late November. River temperatures below Priest Rapids Dam varied from $10.4^{\circ} \mathrm{C}$ (October 23) to $11.5^{\circ} \mathrm{C}$ (November 19) during the spawning period which is typical to that of previous years.

### 11.4 Escapement

The estimated total escapement of fall Chinook salmon to the Hanford Reach for the 2017 return year was 73,759 fishError! Reference source not found.). The historical mean and median escapement for 1991 through 2017 is 75,145 and 57,145 fish, respectively (Table 25). The estimated adult Chinook salmon per redd is calculated by dividing the adult escapement to the Hanford Reach by peak number of redds reported in the redd survey. The estimated annual escapements to the Hanford Reach were not adjusted for pre-spawn mortality. For 2017, the estimated nine fish per redd was the same as the historical mean.

Table 24 Calculation of escapement estimates for fall Chinook salmon in the Hanford Reach, Columbia River 2017.

| Count Source | Return Year 2017 |  |  |
| :--- | ---: | ---: | ---: |
| Jack | Total |  |  |
| McNary Ladder Counts | Adult | 12,014 | 164,199 |
| Adjusted Priest Rapids Adult Passage ${ }^{1}$ | 22,748 | 1,694 | 24,442 |
| Ice Harbor Adult Passage | 26,393 | 5,057 | 31,450 |
| Prosser Adult Passage | 1,947 | 356 | 2,303 |
| Priest Rapids Hatchery | 15,571 | 1,441 | 17,012 |
| Wanapum Tribal Fishery | 0 | 0 | 0 |
| Ringold Springs Hatchery | 1,244 | 47 | 1,291 |
| Yakima River Escapement (Below Prosser) | 520 | 75 | 595 |
| Yakima River Sport Harvest | 470 | 16 | 486 |
| Hanford Sport Harvest | 11,496 | 872 | 12,368 |
| Angler Broodstock Collection | 492 | 0 | 492 |
| Total Non-Hanford Reach Escapement | $\mathbf{8 0 , 8 8 1}$ | $\mathbf{9 , 5 5 8}$ | $\mathbf{9 0 , 4 3 9}$ |
| Hanford Reach Escapement | $\mathbf{7 1 , 3 0 3}$ | $\mathbf{2 , 4 5 6}$ | $\mathbf{7 3 , 7 5 9}$ |

${ }^{1}$ Gross passage count reduced $26.6 \%$ to correct for estimated over counts resulting from fallbacks and re-ascension. The adjustments to adult fish passage were estimated by analysis of the PIT tag detections at PIT tag arrays located in the adult fish ways of the Priest Rapids Dam adult fishway and the discharge channel for Priest Rapids Hatchery.

Table 25 Escapement for fall Chinook salmon in the Hanford Reach, Return Years 1991-2017.

| Return Year | \# Fish per Redd | Redds | Total Escapement ${ }^{1}$ |
| :---: | :---: | :---: | :---: |
| 1991 | 11 | 4,939 | 52,196 |
| 1992 | 9 | 4,926 | 41,952 |
| 1993 | 13 | 2,863 | 37,347 |
| 1994 | 11 | 5,619 | 63,103 |
| 1995 | 18 | 3,136 | 55,208 |
| 1996 | 6 | 7,618 | 43,249 |
| 1997 | 6 | 7,600 | 43,493 |
| 1998 | 7 | 5,368 | 35,393 |
| 1999 | 5 | 6,068 | 29,812 |
| 2000 | 9 | 5,507 | 48,020 |
| 2001 | 10 | 6,248 | 59,848 |
| 2002 | 10 | 8,083 | 84,509 |
| 2003 | 9 | 9,465 | 100,508 |
| 2004 | 10 | 8,468 | 87,696 |
| 2005 | 9 | 7,891 | 71,967 |
| 2006 | 8 | 6,508 | 51,701 |
| 2007 | 6 | 4,018 | 22,272 |
| 2008 | 5 | 5,618 | 29,058 |
| 2009 | 7 | 4,996 | 36,720 |
| 2010 | 10 | 8,817 | 87,016 |
| 2011 | 8 | 8,915 | 75,256 |
| 2012 | 7 | 8,368 | 57,710 |
| 2013 | 10 | 17,398 | 174,651 |
| 2014 | 12 | 15,951 | 183,749 |
| 2015 | 13 | 20,678 | 266,327 |
| 2016 | 9 | 13,268 | 116,388 |
| 2017 | 9 | 8,646 | 73,759 |
| Mean | 9 | 8,036 | 75,146 |
| Median | 9 | 7,600 | 57,710 |

${ }^{1}$ Escapement includes adults and jacks

### 11.5 Hatchery Discharge Channel Redd Counts

The M\&E staff conducted redd counts in the PRH discharge channel on November 20, November 27, and December 5, 2017. Similar to historical observations, the majority of spawning activity was located in a 200 meter section of the discharge channel adjacent to the volunteer trap. A peak count of 17 redds occurred on the December 5 survey. We observed superimposition occurring during multiple surveys; thus making it difficult to determine the total number of redds in a given survey. Viewing conditions during each survey were good to excellent.

### 12.0 Carcass Surveys

Prior to 2010, the carcass surveys in the Hanford Reach were generally performed by two boat crews of two staff operating seven days a week. Beginning in 2010, with support of the PRH M\&E Program, the effort was increased to three boats with a three-person crew operating seven days per week. The extra staffing was necessary to maintain the overall sampling efficiency given the additional effort required to pull otoliths from fish sampled and achieve hatchery M\&E
objectives. The sampling goal for obtaining sufficient number of CWTs is $10 \%$ of the escapement.
Carcass surveys were performed from November 2 through December 13, 2017. All recovered carcasses were sampled for the presence of a CWT. Of those, ~33\% were sampled (i.e., random systematic 1:3 rate) for scales (age), otoliths, gender, length, and egg retention. All carcasses recovered were chopped in half after sampling to prevent the chance of double sampling during subsequent surveys.

Similar to methods used since 2010, the carcass survey crews recorded the sections in which carcasses were recovered in the Hanford Reach and adjacent areas. The Hanford Reach survey is divided into Sections 1 through 5 (Figure 9). The Priest Rapids Pool is designated as Section 6. The PRH discharge channel and the area of the Columbia River immediately below the discharge channel are designated as Sections 7 and 8, respectively. The fall Chinook salmon carcasses recovered in Section 8 were likely wash outs from the hatchery discharge channel.

- Section 1. Priest Rapids Dam to Vernita Bridge (14 km)
- Section 2. Vernita Bridge to Island 2 (19 km)
- Section 3. Island 2 to Powerline Towers at Hanford town site (21 km)
- Section 4. Power line Towers to Wooded Island (21 km)
- Section 5. Wooded Island to Interstate 182 Bridge (19 km)
- Section 6. Priest Rapids Pool (34 km)
- Section 7. Priest Rapids Hatchery discharge channel ( 0.5 km )
- Section 8. Columbia River at the mouth of the Hatchery discharge channel ( 0.5 km )


Figure 9 Location of aerial redd index areas (green area numbers) and river boat carcass survey sections in the Hanford Reach.

### 12.1 Hanford Reach Carcass Survey: Section 1 - 5

Staff recovered 5,591 fall Chinook salmon carcasses in the Hanford Reach in 2017; equating to $7.6 \%$ of the estimated fall Chinook salmon escapement (Table 26). The annual number of fall Chinook salmon carcass recovered in the Hanford Reach for the period of 1991 through 2017 is provided in Appendix E.
Table 26 Numbers and proportions of total escapement of fall Chinook salmon carcasses surveyed for coded-wire tags within each survey section on the Hanford Reach, Return Years, 2010-2017.

| Return <br> Year | $\# \mathbf{N}$ |  | $\# \mathbf{N}$ |  | $\# \mathbf{3}$ |  | $\# \mathbf{4}$ |  | $\# \mathbf{5}$ |  | Total Sampled |  |  |
| :---: | :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | 1,832 | 0.021 | 519 | 0.006 | 3,129 | 0.036 | 3,362 | 0.039 | 937 | 0.011 | 9,779 | 0.112 | 87,016 |
| 2011 | 1,581 | 0.021 | 160 | 0.002 | 2,606 | 0.035 | 2,622 | 0.035 | 1,422 | 0.019 | 8,391 | 0.111 | 75,256 |
| 2012 | 1,091 | 0.019 | 149 | 0.003 | 1,685 | 0.029 | 2,213 | 0.038 | 1,676 | 0.029 | 6,814 | 0.118 | 57,715 |
| 2013 | 2,182 | 0.012 | 1,973 | 0.011 | 2,844 | 0.016 | 3,774 | 0.022 | 2,298 | 0.013 | 13,071 | 0.075 | 174,651 |
| 2014 | 2,682 | 0.015 | 1,142 | 0.006 | 5,544 | 0.030 | 4,573 | 0.025 | 2,815 | 0.015 | 16,756 | 0.091 | 183,680 |
| 2015 | 2,913 | 0.011 | 823 | 0.003 | 6,187 | 0.023 | 5,868 | 0.022 | 1,947 | 0.007 | 17,738 | 0.067 | 266,346 |
| 2016 | 1,141 | 0.010 | 513 | 0.004 | 2,796 | 0.024 | 2,977 | 0.026 | 1,459 | 0.013 | 8,886 | 0.076 | 116,388 |
| 2017 | 1,098 | 0.015 | 346 | 0.005 | 1,275 | 0.17 | 1850 | 0.025 | 1,022 | 0.014 | 5,591 | 0.076 | 73,759 |
| Mean | $\mathbf{1 , 8 1 5}$ | $\mathbf{0 . 0 1 6}$ | $\mathbf{7 0 3}$ | $\mathbf{0 . 0 0 5}$ | $\mathbf{3 , 2 5 8}$ | $\mathbf{0 . 0 4 5}$ | $\mathbf{3 , 4 0 5}$ | $\mathbf{0 . 0 2 9}$ | $\mathbf{1 , 6 9 7}$ | $\mathbf{0 . 0 1 5}$ | $\mathbf{1 0 , 8 7 8}$ | $\mathbf{0 . 0 9 1}$ | $\mathbf{1 2 9 , 3 5 1}$ |

The survey effort was not equal for each section. Sections 3 and 4 were surveyed the most because these sections generally contain the largest number of carcasses (Table 27). As each season progresses, crews focused their effort in sections that provided greater chances to recover carcasses.
Table 27 Number of carcass surveys conducted by section in the Hanford Reach, Return Years 2010-2017.

| Return Year | $\# \mathbf{1}$ | $\# \mathbf{2}$ | $\# \mathbf{3}$ | $\# \mathbf{4}$ | $\boldsymbol{\#} \mathbf{5}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | 21 | 6 | 26 | 26 | 11 | 90 |
| 2011 | 33 | 5 | 38 | 29 | 13 | 118 |
| 2012 | 19 | 4 | 26 | 28 | 24 | 101 |
| 2013 | 18 | 15 | 16 | 17 | 13 | 79 |
| 2014 | 23 | 17 | 30 | 31 | 24 | 125 |
| 2015 | 23 | 8 | 35 | 37 | 13 | 116 |
| 2016 | 18 | 11 | 29 | 27 | 15 | 100 |
| 2017 | 19 | 14 | 30 | 31 | 17 | 111 |
| Mean | $\mathbf{2 2}$ | $\mathbf{1 0}$ | $\mathbf{2 9}$ | $\mathbf{2 8}$ | $\mathbf{1 6}$ | $\mathbf{1 0 5}$ |

### 12.2 Proportion of Escapement Sampled: Section 1-5

The spawning escapement for sections 1 through 5 was estimated by the proportion of redds counted in aerial surveys to the estimated escapement of natural spawners to the Hanford Reach (see Section 14 - Redd Surveys). The calculations for estimating the escapement to the Hanford Reach are given in Appendix F.
We have identified through the carcass bias assessment that an unknown number of carcasses drift into downstream sections after spawning. The recovery of these carcasses may confound the estimate of the spawning escapement sampled by section as shown in Table 28. For example,
there were no redds identified in Section 5 but 1,459 carcasses were recovered in that section. It is likely that sections 1 and 3 , that have the greatest number of redds and largest spawning escapement, end up with a net loss of carcasses to downstream sections. In 2017, we continued a pilot study to evaluate the magnitude and distribution of post spawn carcass drift. The preliminary results of this study are included in the Appendix G.
Table 28 Number of redds and carcasses, total spawning escapement, and proportion of escapement sampled for fall Chinook salmon in Section 1 through 5 of the Hanford Reach, Return Year 2017.

| Survey <br> Section | Total Number of <br> Redds | Total Number of <br> Carcasses | Spawning Escapement ${ }^{\mathbf{1}}$ | Proportion of <br> Escapement Sampled |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 3,285 | 1,098 | 28,024 | 0.039 |
| 2 | 410 | 346 | 3,498 | 0.138 |
| 3 | 4,671 | 1,275 | 39,848 | 0.025 |
| 4 | 280 | 1,850 | 2,389 | 0.471 |
| 5 | 0 | 1,022 | 0 | -- |
| Total | $\mathbf{8 , 6 4 6}$ | $\mathbf{5 , 5 9 1}$ | $\mathbf{7 3 , 7 5 9}$ | $\mathbf{0 . 0 7 6}$ |

${ }^{1}$ Calculated based on proportion of redds by section

### 12.3 Carcass Distribution and Origin

Two methods were used to estimate the origin of carcasses recovered in the sections 1 through 5. The first method includes the expansion of pooled CWT recoveries using juvenile tag rates and survey sample rate. The second method includes calculating the proportion of combined hatchery marks (i.e., otolith mark, adipose clips, and CWTs) to non-marked carcasses. Estimates for both methods are given for the 2012-2017 adult returns: these years include otolith marks for all ages of PRH origin fish.

The assumption was made that all Chinook salmon not accounted by hatchery origin CWT expansions were of natural origin. This assumption may underestimate the number of hatchery carcasses recovered in the annual surveys. We have compelling evidence to suggest this is the case with annual returns to PRH prior to return year 2014. The expansion of CWT recoveries suggest that $11.6 \%$ of the fall Chinook salmon carcasses recovered in the 2017 Hanford Reach stream surveys were hatchery origin (Table 29). This estimate is slightly greater than the mean pHOS value generated from CWT recoveries for years 2010 through 2017. The expanded CWT recovery data suggest the hatchery origin component of the escapement included $7.8 \%$ from PRH, $3.5 \%$ from RSH and $0.3 \%$ from other hatcheries. The highest proportions of hatchery origin carcasses recovered based on CWT recoveries were in Sections 2, and 4, respectively.

The second estimate of origin of carcasses recovered is based on the proportion of hatchery marked to non-marked fish. For this method, we assume that all hatchery origin carcasses recovered are marked in some manner (e.g., otolith marks, CWT, and adipose clips) and that we are able to accurately detect these marks and tags.
PRH has marked their entire juvenile releases with annual marks on the otoliths beginning with progeny of brood year 2007. For the 2013-2017 returns, age-2 through 6 PRH origin carcasses recovered were otolith marked. The age-6 PRH origin fish were not otolith marked during return year 2012. However, since there were no age-6 fish recovered in the carcass surveys or at PRH, it is assumed that few, if any PRH origin age-6 fish spawned in the Hanford Reach. Adipose clipped Chinook salmon without a CWT and without a thermal otolith mark were classified as
strays from other hatcheries. The natural origin fish were identified by either a Hanford Reach origin CWT or by the presence of an adipose fin and the absence of an otolith mark. The demographic sample data suggests that $8.3 \%$ of fall Chinook salmon carcasses recovered in the 2017 Hanford Reach stream survey were hatchery origin (Table 30). For recent years, the hatchery proportions were generally higher in the upstream survey sections.
Table 29 Numbers of natural and hatchery origin fall Chinook salmon carcasses sampled within Section 1 through 5 of Hanford Reach based on expansions of coded-wire tag recoveries, Return Years 2010-2017.

| Return <br> Year | Origin | Hanford Reach Sections |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Natural | 1,751 | 473 | 3,020 | 3,242 | 909 | 9,395 | 0.960 |
|  | Proportion |  |  |  |  |  |  |  |
| of Sample |  |  |  |  |  |  |  |  |$|$

Table 30 Origin of Chinook salmon carcasses recovered in the Hanford Reach by section based on recoveries of marked and unmarked carcasses within the biological sample, Return Years 2012-2017.

| Year | Origin | \# 1 | \# 2 | \# 3 | \# 4 | \# 5 | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2012$ <br> Biological sample Rate 1:4 $\mathrm{N}=1,609$ | PRH ${ }^{1}$ | 23 | 2 | 26 | 18 | 38 | 107 | 0.067 |
|  | Other Hatchery ${ }^{2}$ | 10 | 2 | 25 | 45 | 22 | 104 | 0.065 |
|  | Total Hatchery | 33 | 4 | 51 | 63 | 60 | 211 | 0.131 |
|  | Natural ${ }^{3}$ | 228 | 30 | 347 | 460 | 333 | 1,398 | 0.869 |
|  | Proportion Hatchery | 0.126 | 0.118 | 0.128 | 0.120 | 0.153 | 0.131 |  |
| $2013^{a}$ <br> Biological sample rate $=1: 5$ and then randomly sub-sampled, $\mathrm{N}=$ 712 | PRH ${ }^{1}$ | 32 | 19 | 34 | 30 | 32 | 147 | 0.206 |
|  | Other Hatchery ${ }^{2}$ | 6 | 3 | 16 | 21 | 6 | 52 | 0.073 |
|  | Total Hatchery | 38 | 22 | 50 | 51 | 38 | 199 | 0.279 |
|  | Natural ${ }^{3}$ | 76 | 84 | 113 | 155 | 85 | 513 | 0.721 |
|  | Proportion Hatchery | 0.333 | 0.208 | 0.307 | 0.248 | 0.309 | 0.279 |  |
| $2014^{\text {a }}$ <br> Biological sample rate $=1: 5$ and then randomly sub-sampled, $\mathrm{N}=$ 2,426 | PRH ${ }^{1}$ | 37 | 7 | 45 | 35 | 11 | 135 | 0.056 |
|  | Other Hatchery ${ }^{2}$ | 12 | 5 | 16 | 32 | 18 | 83 | 0.034 |
|  | Total Hatchery | 49 | 12 | 61 | 67 | 29 | 218 | 0.090 |
|  | Natural ${ }^{3}$ | 347 | 142 | 711 | 612 | 396 | 2208 | 0.910 |
|  | Proportion Hatchery | 0.124 | 0.078 | 0.079 | 0.099 | 0.068 | 0.090 |  |
| 2015 <br> Biological sample $\begin{aligned} & \text { rate }=1: 7 \\ & \mathrm{~N}=2,485 \end{aligned}$ | PRH ${ }^{1}$ | 47 | 12 | 61 | 55 | 13 | 188 | 0.076 |
|  | Other Hatchery ${ }^{2}$ | 6 | 2 | 17 | 20 | 7 | 52 | 0.021 |
|  | Total Hatchery | 53 | 14 | 78 | 75 | 20 | 240 | 0.097 |
|  | Natural ${ }^{3}$ | 346 | 101 | 792 | 752 | 254 | 2,245 | 0.903 |
|  | Proportion Hatchery | 0.133 | 0.122 | 0.090 | 0.091 | 0.073 | 0.097 |  |
| 2016 <br> Biological sample $\begin{aligned} & \text { rate }=1: 5 \\ & \mathrm{~N}=1,743 \end{aligned}$ | PRH ${ }^{1}$ | 27 | 12 | 42 | 22 | 10 | 113 | 0.066 |
|  | Other Hatchery ${ }^{2}$ | 9 | 6 | 31 | 23 | 13 | 82 | 0.048 |
|  | Total Hatchery | 36 | 18 | 73 | 45 | 23 | 195 | 0.114 |
|  | Natural ${ }^{3}$ | 182 | 80 | 465 | 534 | 257 | 1,518 | 0.886 |
|  | Proportion Hatchery | 0.165 | 0.184 | 0.136 | 0.078 | 0.082 | 0.114 |  |
| $2017$ <br> Biological sample $\begin{aligned} & \text { rate }=1: 3 \\ & \mathrm{~N}=1,813 \end{aligned}$ | PRH ${ }^{1}$ | 42 | 19 | 21 | 19 | 16 | 117 | 0.065 |
|  | Other Hatchery ${ }^{2}$ | 7 | 2 | 4 | 14 | 6 | 33 | 0.018 |
|  | Total Hatchery | 49 | 21 | 25 | 33 | 22 | 150 | 0.083 |
|  | Natural ${ }^{3}$ | 311 | 86 | 391 | 564 | 311 | 1,663 | 0.917 |
|  | Proportion Hatchery | 0.136 | 0.196 | 0.060 | 0.055 | 0.066 | 0.083 |  |
| Mean Proportion | PRH ${ }^{1}$ | 0.137 | 0.115 | 0.088 | 0.061 | 0.086 | 0.089 |  |
|  | Other Hatchery ${ }^{2}$ | 0.033 | 0.036 | 0.045 | 0.054 | 0.040 | 0.043 |  |
|  | Total Hatchery | 0.170 | 0.151 | 0.133 | 0.115 | 0.125 | 0.132 |  |
|  | Natural ${ }^{3}$ | 0.830 | 0.849 | 0.867 | 0.885 | 0.875 | 0.868 |  |

${ }^{\text {a }}$ Estimate of origin based on random sub-sample of biological sample.
${ }^{1}$ Priest Rapids Hatchery fish were identified by either the presence Priest Rapids Hatchery otolith mark or codedwire tag
${ }^{2}$ Other hatchery strays were identified as adipose clipped Chinook salmon without a Priest Rapids Hatchery coded-wire tag and without a thermal otolith mark or by the presence of other hatchery coded-wire tags.
${ }^{3}$ Natural origin fish were identified by either a Hanford Reach origin coded-wire tag or by the presence of an adipose fin and the absence of an otolith mark.

### 12.4 Priest Rapids Dam Pool Carcass Survey: Section 6

In total, five carcass surveys were performed in Section 6 during return year 2017 (Table 31). Surveys were scheduled once or twice a week between early November and mid-December.
Table 31 Number of fall Chinook salmon carcasses sampled within Section 6 (Priest Rapids Dam Pool), Return Years 2010-2017.

|  | Section 6 |  |
| :---: | :---: | :---: |
| Year | \# of Carcasses | \# of Surveys |
| 2010 | 123 | 8 |
| 2011 | 69 | 7 |
| 2012 | 72 | 4 |
| 2013 | 407 | 7 |
| 2014 | 237 | 7 |
| 2015 | 155 | 6 |
| 2016 | 139 | 8 |
| 2017 | 40 | 5 |
| Mean | $\mathbf{1 5 5}$ | $\mathbf{7}$ |

### 12.5 Number sampled: Section 6

Survey crews recovered 40 Chinook salmon in Section 6 during return year 2017 (Table 31). All fish recovered were scanned for the presence of a CWT. Carcass recoveries in the lower portion of the pool suggest that carcasses drift downstream of the spawning areas below Wanapum Dam into deeper water where they are difficult to locate and recover.

### 12.5.1 Proportion of Escapement Sampled: Section 6

The spawning escapement for Section 6 was calculated by subtracting from the Priest Rapids Dam fall Chinook salmon passage count, the fall Chinook salmon passage at Wanapum Dam, tribal and sport harvest of fall Chinook salmon in the Priest Rapids Dam pool, and the estimated fallback of fall Chinook salmon at Priest Rapids Dam (Appendix F).
The 2017 fall Chinook salmon spawning escapement estimate for Section 6 is 1,788 fish. Overall, roughly 2\% of the total estimated spawning escapement in Section 6 was sampled in 2017 (Table 32).
Table 32 Carcasses sampled, total spawning escapement and proportion of escapement for fall Chinook salmon in Section 6 (Priest Rapids Dam Pool), Return Years 2010-2017.

| Return Year | \# of Surveys | \# of Carcasses | Spawning Escapement | Escapement Sampled |
| :---: | :---: | :---: | :---: | :---: |
| 2010 | 8 | 123 | 11,121 | 0.011 |
| 2011 | 7 | 69 | 11,362 | 0.006 |
| 2012 | 4 | 72 | 21,919 | 0.003 |
| 2013 | 7 | 407 | 62,237 | 0.007 |
| 2014 | 7 | 237 | 25,179 | 0.009 |
| 2015 | 6 | 155 | 38,313 | 0.004 |
| 2016 | 8 | 139 | 13,162 | 0.011 |
| 2017 | 5 | 40 | 1,788 | 0.022 |
| Mean | $\mathbf{7}$ | $\mathbf{1 5 5}$ | $\mathbf{2 3 , 1 3 5}$ | $\mathbf{0 . 0 0 9}$ |

### 12.5.2 Carcass Origin: Section 6

Similar to those methods described in detail in the previous section, the carcasses included in the 1:1 demographic sample were identified as hatchery origin based on a combination of hatchery marks and tags (i.e., otoliths marks, adipose clips, and CWTs). Natural origin carcasses were identified by the absence of any hatchery mark or the presence of a natural origin CWT.
An estimated $46.6 \%$ of the fall Chinook salmon spawning in Section 6 were hatchery origin of which $93.8 \%$ were PRH origin (Table 33).
Table 33 Origin of fall Chinook salmon spawning in Section 6 (Priest Rapids Dam Pool), Return Years 2012-2017.

| Year | Origin | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: |
| $\begin{gathered} 2012 \\ \mathrm{~N}=70 \end{gathered}$ | $\mathrm{PRH}^{1}$ | 18 | 0.257 |
|  | Other Hatchery ${ }^{2}$ | 2 | 0.029 |
|  | Total Hatchery | 20 | 0.286 |
|  | Natural ${ }^{3}$ | 50 | 0.714 |
| $\begin{gathered} 2013 \\ \mathrm{~N}=98 \end{gathered}$ | PRH ${ }^{1}$ | 62 | 0.633 |
|  | Other Hatchery ${ }^{2}$ | 5 | 0.051 |
|  | Total Hatchery | 67 | 0.684 |
|  | Natural ${ }^{3}$ | 31 | 0.316 |
| $\begin{gathered} 2014 \\ \mathrm{~N}=229 \end{gathered}$ | $\mathrm{PRH}^{1}$ | 81 | 0.354 |
|  | Other Hatchery ${ }^{2}$ | 5 | 0.022 |
|  | Total Hatchery | 86 | 0.376 |
|  | Natural ${ }^{3}$ | 143 | 0.624 |
| $\begin{gathered} 2015 \\ \mathrm{~N}=244 \end{gathered}$ | PRH ${ }^{1}$ | 83 | 0.535 |
|  | Other Hatchery ${ }^{2}$ | 3 | 0.019 |
|  | Total Hatchery | 155 | 0.555 |
|  | Natural ${ }^{3}$ | 69 | 0.445 |
| $\begin{gathered} 2016 \\ \mathrm{~N}=134 \end{gathered}$ | PRH ${ }^{1}$ | 66 | 0.475 |
|  | Other Hatchery ${ }^{2}$ | 3 | 0.022 |
|  | Total Hatchery | 69 | 0.496 |
|  | Natural ${ }^{3}$ | 65 | 0.468 |
| $\begin{gathered} 2017 \\ \mathrm{~N}=40 \end{gathered}$ | PRH ${ }^{1}$ | 15 | 0.375 |
|  | Other Hatchery ${ }^{2}$ | 1 | 0.025 |
|  | Total Hatchery | 16 | 0.400 |
|  | Natural ${ }^{3}$ | 24 | 0.600 |
| Mean Proportions | PRH ${ }^{1}$ |  | 0.411 |
|  | Other Hatchery ${ }^{2}$ |  | 0.026 |
|  | Total Hatchery |  | 0.523 |
|  | Natural ${ }^{3}$ |  | 0.477 |

${ }^{1}$ Priest Rapids Hatchery fish were identified by either the presence of thermal otolith mark or by the presence of a PRH origin coded-wire tag
${ }^{2}$ Other hatchery strays were identified as adipose clipped Chinook salmon without a Priest Rapids Hatchery coded-wire tag and without a thermal otolith mark.
${ }^{3}$ Natural origin fish were identified by either a Hanford Reach origin coded-wire tag or by the presence of an adipose fin and the absence of an otolith mark.

### 12.6 Hatchery Discharge Channel: Section 7 and 8 Carcass Survey

During return year 2017, crews performed one carcass surveys in Section 8 by boat and one carcass survey in Section 7 by foot. It has been observed that many carcasses drift out of the discharge channel under full flow conditions. Performing carcass surveys in the discharge channel when it is at full flow is difficult and dangerous due to poor footing and high velocities. Staff performed the one survey in Section 7 on December 6 when discharge levels in the channel were still high. It's likely a portion of the carcasses may have drifted out of the discharge channel by the date that it was surveyed.

### 12.7 Number sampled: Section 7 and 8

Survey crews recovered 9 carcasses in Section 7 and 16 in Section 8 (Table 34). All fish recovered were scanned for the presence of a CWT.

Table 34 The number of fall Chinook salmon carcass surveys within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at the confluence of the hatchery discharge channel), Return Years 2010-2017.

|  | Section 7 |  | Section 8 |  | Total |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Return Year | \# of <br> Carcasses | \# of <br> Surveys | \# of <br> Carcasses | \# of <br> Surveys | \# of <br> Carcasses | \# of <br> Surveys |
| 2010 | 87 | 1 | 123 | 9 | 210 | 10 |
| 2011 | 123 | 2 | 80 | 8 | 203 | 10 |
| 2012 | 99 | 3 | 108 | 10 | 207 | 13 |
| 2013 | 105 | 3 | 159 | 4 | 264 | 7 |
| 2014 | 9 | 1 | 52 | 7 | 61 | 8 |
| 2015 | 33 | 1 | 26 | 2 | 59 | 3 |
| 2016 | 3 | 1 | 7 | 1 | 10 | 2 |
| 2017 | 9 | 1 | 16 | 1 | 25 | 2 |
| Mean | $\mathbf{5 9}$ | $\mathbf{2}$ | $\mathbf{7 1}$ | $\mathbf{5}$ | $\mathbf{1 3 0}$ | $\mathbf{7}$ |

12.7.1 Proportion of Escapement Sampled: Section 7 and 8

The 2017 fall Chinook salmon spawning escapement index for Sections 7 and 8 is 34 fish (Table 35). The spawning escapement for these Sections was calculated using the expansion factor of two fish/redd, based on a 0.5 male/female sex ratio including jacks, as estimated from the Hanford Reach 2017 escapement. Therefore, the assumption is made that each of the 17 redds represents one female and one male. We assume that most of the carcasses recovered in Section 8 drifted downstream from Section 7. In addition, it is likely a portion of carcasses from Sections 7 and 8 drift downstream into Sections 1 and 2.

Table 35 Number of carcasses sampled, total spawning escapement and proportion of escapement sampled for fall Chinook salmon within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at confluence of the hatchery discharge channel), Return Year 2017.

| Section | Total Number of Carcasses | Spawning Escapement | Escapement Sampled |
| :---: | :---: | :---: | :---: |
| $\# 7$ | 9 | 34 | 0.735 |
| $\# 8$ | 16 | 0 |  |
| Total | $\mathbf{2 5}$ | $\mathbf{3 4}$ | $\mathbf{0 . 7 3 5}$ |

### 12.7.2 Carcass Distribution and Origin: Section 7 and 8

The demographic sample rate was set at 1:1 to account for the low numbers of carcasses recovered. As described in detail previously, the carcasses included the demographic sample were identified as hatchery origin based on a combination of hatchery marks and tags (i.e., otoliths marks, adipose clips, and CWTs). Natural origin carcasses were identified by the absence of any hatchery mark or the presence of a natural origin CWT.

It is estimated that $75.0 \%$ of fall Chinook salmon recovered in Sections 7 and 8 were hatchery origin of which all were PRH origin (Table 36).
Table 36 The origin of Chinook salmon carcasses recovered within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at the confluence of the hatchery discharge channel), Return Years 2012-2017.

| Return Year | Origin | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: |
| $\begin{gathered} 2012 \\ \mathrm{~N}=70 \end{gathered}$ | PRH | 18 | 0.257 |
|  | Other Hatchery | 2 | 0.029 |
|  | Total Hatchery | 20 | 0.286 |
|  | Natural | 50 | 0.714 |
| $\begin{gathered} 2013 \\ \mathrm{~N}=33 \end{gathered}$ | PRH | 28 | 0.848 |
|  | Other Hatchery | 2 | 0.061 |
|  | Total Hatchery | 30 | 0.909 |
|  | Natural | 3 | 0.091 |
| $\begin{aligned} & 2014 \\ & \mathrm{~N}=5 \end{aligned}$ | PRH | 3 | 0.600 |
|  | Other Hatchery | 0 | 0.000 |
|  | Total Hatchery | 3 | 0.600 |
|  | Natural | 2 | 0.400 |
| $\begin{gathered} 2015 \\ \mathrm{~N}=59 \end{gathered}$ | PRH | 19 | 0.322 |
|  | Other Hatchery | 2 | 0.034 |
|  | Total Hatchery | 21 | 0.356 |
|  | Natural | 38 | 0.644 |
| $\begin{gathered} 2016 \\ \mathrm{~N}=6 \end{gathered}$ | PRH | 4 | 0.667 |
|  | Other Hatchery | 1 | 0.167 |
|  | Total Hatchery | 5 | 0.833 |
|  | Natural | 1 | 0.167 |
| $\begin{gathered} 2017 \\ \mathrm{~N}=6 \end{gathered}$ | PRH | 6 | 0.750 |
|  | Other Hatchery | 0 | 0.000 |
|  | Total Hatchery | 6 | 0.750 |
|  | Natural | 2 | 0.250 |
| Means Proportions | PRH |  | 0.574 |
|  | Other Hatchery |  | 0.023 |
|  | Total Hatchery |  | 0.622 |
|  | Natural |  | 0.378 |

### 13.0 Life History Monitoring

Migration timing of hatchery and natural origin Hanford Reach fall Chinook salmon is estimated from arrival timing at Bonneville Dam based on PIT tag observations at the adult fish ladder for both PRH and Hanford Reach origin fall Chinook salmon.

Life history characteristics of Hanford Reach fall Chinook salmon were assessed by examining carcasses on spawning grounds, fish collected or examined at broodstock collection sites, and by reviewing tagging data and fisheries statistics.

For the 2012-2017 returns, the origin of fall Chinook salmon for the comparison of age and length at maturity is based on a combination of hatchery marks and tags (i.e., otolith, adipose clips, and CWT). PRH origin fall Chinook Salmon were identified by either the presence of an otolith mark specific to PRH or by the presence of a PRH origin CWT. Adipose clipped Chinook salmon without a CWT and without an otolith mark were classified as fish from other hatcheries. The natural origin fish were identified by either a Hanford Reach origin CWT or by the presence of an adipose fin combined with the absence of any hatchery marks. The age composition for both the natural and hatchery origin fall Chinook salmon recovered in return years 2012-2017 were assembled from the carcass recoveries in sections 1-8 of the Hanford Reach.

In order to make coarse comparisons between hatchery and natural origin fish prior to return year 2012, the designation of origin required the assumption that all fish collected in the Hanford Reach, except for those that were of known hatchery origin (e.g., adipose clipped or possessed a CWT), were natural origin. We know this was not the case, but we were not able to identify all of the hatchery origin fish in the demographic samples and it was assumed that the majority of the fish sampled in the stream surveys were natural origin.

### 13.1 Migration Timing

PIT tag observations for both PRH and Hanford Reach natural origin adult fall Chinook salmon at the PIT tag arrays in the Bonneville Dam adult fish ladders were used to assess arrival timing. The PIT tag observation data was obtained from the PTAGIS website. Arrival date for each unique tagged adult was based on its first observation date and time at Bonneville Dam. Annually, the sample sizes have been relatively small due to the low numbers of both hatchery and natural origin fall Chinook salmon PIT tagged. Beginning with the 2011 brood, the number of juveniles PIT tagged at PRH increased from 3,000 to roughly 43,000 annually

The adult PIT tag detections at Bonneville Dam are useful to compare migration timing between Hanford Reach natural origin and PRH origin fall Chinook salmon because harvest and other losses upstream of Bonneville Dam reduce the number of potential detections at upstream sites.

The $10^{\text {th }}, 50^{\text {th }}$, and $90^{\text {th }}$ percentiles of the annual migration timing to Bonneville Dam are given in (Table 37). The observation sample size of both groups of PIT tagged fish at Bonneville Dam can be small and therefore, may not be representative of the populations. However this may be the best migration information currently available.

Table 37 The week that $10 \%, 50 \%$ (median), and $90 \%$ of the natural and hatchery origin fall Chinook salmon passed Bonneville Dam, 2010-2017. Migration timing is based on PIT tag passage of Hanford natural origin and Priest Rapids Hatchery in the adult fish ladder at Bonneville Dam.

| Return Year | Origin | Hanford Reach Fall Chinook Migration Time (Date) |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Priest Rapids Origin |  |  |  |  | Hanford Reach Natural Origin |  |  |  |  |
|  |  | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 |
| 2010 | $10^{\text {th }}$ Percentile | 28-Aug | 26-Aug | -- | 24-Aug | -- | 31-Aug | 5-Sep | 25-Aug | -- | -- |
|  | $50^{\text {th }}$ Percentile | 9-Sep | 17-Sep | -- | 4-Sep | -- | 21-Sep | 17-Sep | 9-Sep | -- | -- |
|  | $90^{\text {th }}$ Percentile | 15-Sep | 24-Sep | -- | 6-Sep | -- | 4-Oct | 6-Oct | 15-Sep | -- | -- |
|  | N | 5 | 20 | 0 | 3 | 0 | 8 | 22 | 18 | 0 | 0 |
| 2011 | $10^{\text {th }}$ Percentile | 8-Aug | 3-Sep | 23-Aug | -- | -- | -- | 4-Sep | 24-Aug | 4-Aug | 4-Aug |
|  | $50^{\text {th }}$ Percentile | 8-Sep | 20-Sep | 8-Sep | -- | -- | -- | 4-Sep | 10-Sep | 30-Aug | 30-Aug |
|  | $90^{\text {th }}$ Percentile | 21-Sep | 25-Sep | 21-Sep | -- | -- | -- | 10-Sep | 2-Oct | 1-Sep | 1-Sep |
|  | N | 6 | 7 | 10 | 0 | 0 | 0 | 2 | 65 | 3 | 3 |
| 2012 | $10^{\text {th }}$ Percentile | 31-Aug | 6-Sep | 13-Sep | 7-Sep | -- | 14-Sep | 4-Sep | 28-Aug | 27-Aug | 27-Aug |
|  | $50^{\text {th }}$ Percentile | 16-Sep | 11-Sep | 13-Sep | 7-Sep | -- | 23-Sep | 16-Sep | 5-Sep | 8-Sep | 8-Sep |
|  | $90^{\text {th }}$ Percentile | 27-Sep | 21-Sep | 19-Sep | 7-Sep | -- | 10-Oct | 26-Sep | 21-Sep | 19-Sep | 19-Sep |
|  | N | 7 | 13 | 2 | 1 | 0 | 10 | 11 | 19 | 26 | 26 |
| 2013 | $10^{\text {th }}$ Percentile | 24-Aug | 28-Aug | 25-Aug | -- | -- | 11-Sep | 2-Sep | 2-Sep | 9-Aug | 9-Aug |
|  | $50^{\text {th }}$ Percentile | 8-Sep | 9-Sep | 3-Sep | -- | -- | 11-Sep | 22-Sep | 9-Sep | 27-Aug | 27-Aug |
|  | $90^{\text {th }}$ Percentile | 18-Sep | 22-Sep | 15-Sep | -- | -- | 11-Sep | 10-Oct | 19-Sep | 2-Oct | 2-Oct |
|  | N | 40 | 55 | 16 | 0 | 0 | 1 | 29 | 22 | 10 | 10 |
| 2014 | $10^{\text {th }}$ Percentile | 6-Sep | 4-Sep | 5-Sep | -- | -- | 24-Sep | 10-Sep | 3-Sep | 29-Aug | 29-Aug |
|  | $50^{\text {th }}$ Percentile | 16-Sep | 13-Sep | 12-Sep | -- | -- | 25-Sep | 11-Sep | 12-Sep | 1-Sep | 1-Sep |
|  | $90^{\text {th }}$ Percentile | 28-Sep | 25-Sep | 23-Sep | -- | -- | 1-Oct | 28-Sep | 26-Sep | 15-Sep | 15-Sep |
|  | N | 175 | 228 | 50 | 0 | 0 | 3 | 4 | 62 | 5 | 5 |
| 2015 | $10^{\text {th }}$ Percentile | 16-Oct | 8-Sep | 25-Aug | 14-Sep | -- | -- | 10-Sep | 30-Aug | 29-Aug | 29-Aug |
|  | $50^{\text {th }}$ Percentile | $16-$ Oct | 21-Sep | 6-Sep | 26-Sep | -- | -- | 20-Sep | 10-Sep | 9-Sep | 9-Sep |
|  | $90^{\text {th }}$ Percentile | 16-Oct | 9-Oct | 18-Sep | 26-Sep | -- | -- | 1-Oct | 25-Sep | 25-Sep | 25-Sep |
|  | N | 1 | 345 | 323 | 2 | 0 | 0 | 5 | 13 | 32 | 32 |
| 2016 | $10^{\text {th }}$ Percentile | -- | 30-Aug | 8-Aug | 14-Aug | -- | -- | 21-Sep | 28-Aug | 31-Aug | 31-Aug |
|  | $50^{\text {th }}$ Percentile | -- | 13-Sep | 7-Sep | 1-Sep | -- | -- | 21-Sep | 10-Sep | 7-Sep | 7-Sep |
|  | $90^{\text {th }}$ Percentile | -- | 6-Oct | 19-Sep | 15-Sep | -- | -- | 14-Oct | 19-Sep | 14-Sep | 14-Sep |
|  | N | 0 | 41 | 182 | 41 | 0 | 0 | 2 | 10 | 5 | 5 |
| 2017 | $10^{\text {th }}$ Percentile | 10-Sep | 5-Sep | 5-Sep | 31-Aug | 27-Sep | 24-Sep | 12-Sep | 26-Aug | 5-Sep | -- |
|  | $50^{\text {th }}$ Percentile | 20-Sep | 18-Sep | 14-Sep | 12-Sep | 27-Sep | 24-Sep | 12-Sep | 12-Sep | 15-Sep | -- |
|  | $90^{\text {th }}$ Percentile | 31-Oct | 9-Oct | 24-Sep | 18-Sep | 27-Sep | 24-Sep | 12-Sep | 3-Nov | 11-Oct | -- |
|  | N | 8 | 19 | 63 | 48 | 1 | 1 | 1 | 19 | 13 | 0 |

### 13.2 Age at Maturity

Prior to return year 2012, the fish origin was assigned by location of survey due to the lack of identifiable hatchery marks and low CWT recoveries that were not representative of natural origin fish. Hence, the age composition for natural origin returns was generated from all the samples collected within the carcass survey regardless of true origin. Likewise, the age composition for hatchery origin fish was generated from all samples collected at PRH regardless of true origin.

The age compositions of the Hanford Reach escapement and the PRH returns are not directly comparable between locations without some adjustment. There is likely a recovery bias against
smaller/younger fish in the stream surveys (Zhou 2002; Murdoch et al. 2010; Richards and Pearsons, 2013). Hence, the age composition for the Hanford Reach escapement is likely biased towards larger/older fish. Results and brief discussion for the pilot carcass bias assessments are given in Appendix H. All fish recovered from the PRH volunteer trap are available for systematic sampling; reducing the potential bias of the age composition data. Although this dataset is imperfect, the dataset is maintained for future reference should a method be established to correct the data for associated age and origin bias (Table 38).
The availability of otolith data combined with other hatchery mark data from the Hanford Reach carcass recoveries for the 2012 through 2017 return years provide the ability to estimate age compositions for both hatchery and natural origin fish within the Hanford Reach escapement (Table 39). However, the hatchery origin age composition may be influenced by the low number of hatchery origin fish present in the demographic samples which is further reduced by subsampling the demographic origin. In addition, the age composition for both groups may be biased towards larger fish due to potential size recovery biases in the carcass surveys. Larger demographic samples per return year are required to better represent the age composition data before conclusions can be made. Beginning with return year 2014, the sub-sample size to determine origin was increased substantially to roughly 2,500 fish in order to capture more hatchery origin fish in the sub-sample. Regardless of the methodology, it appears that natural origin fish return at older ages than hatchery origin fish. More specifically, the proportion of hatchery origin fish was higher than natural origin fish at ages 2 and 3, and the opposite was true for ages 4, 5, and 6 during brood years 2007-2011.

### 13.3 Size at Maturity

Prior to return year 2012, the size (fork length) at maturity comparisons between fall Chinook salmon recovered at PRH and the Hanford Reach stream survey were calculated in a similar manner as the age composition data for the same time period. Likewise, the assignment of origin was based on the survey (i.e., stream or hatchery). The estimates based on this method may not be representative of natural and hatchery origin fish due to possible size bias during recovery of carcasses.

The availability of otolith marks in addition to other hatchery marks (i.e., otoliths, adipose clips, and CWTs) for the 2012 through 2017 return years provide the ability to estimate size at maturity for both hatchery and natural origin fish within the Hanford Reach escapement.
The size at maturity data is essentially complete for brood years 2007 through 2012. It appears that age- 2 and 3 hatchery origin fish tend to be larger than natural origin fish. Likewise, age- 4 and 5 natural origin fish tend to be larger than their hatchery origin counterparts (Table 40). This pattern is also seen in Table 41 with the exception that age- 3 fish are similar size in the PRH and stream sample.

Table 38 Age compositions for fall Chinook salmon sampled in the Hanford Reach escapement compared to fall Chinook salmon sampled at Priest Rapids Hatchery (genders combined), Brood Years 1998-2012.

| Brood Year | Source ${ }^{1}$ | Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 1998 | Escapement | 0.119 | 0.097 | 0.420 | 0.346 | 0.018 |
|  | PRH Returns | 0.034 | 0.575 | 0.353 | 0.038 | 0.000 |
| 1999 | Escapement | 0.123 | 0.089 | 0.390 | 0.392 | 0.005 |
|  | PRH Returns | 0.061 | 0.366 | 0.432 | 0.140 | 0.001 |
| 2000 | Escapement | 0.262 | 0.081 | 0.290 | 0.359 | 0.009 |
|  | PRH Returns | 0.070 | 0.303 | 0.467 | 0.152 | 0.007 |
| 2001 | Escapement | 0.152 | 0.149 | 0.488 | 0.206 | 0.005 |
|  | PRH Returns | 0.061 | 0.506 | 0.309 | 0.122 | 0.002 |
| 2002 | Escapement | 0.178 | 0.154 | 0.568 | 0.099 | 0.001 |
|  | PRH Returns | 0.103 | 0.386 | 0.466 | 0.043 | 0.001 |
| 2003 | Escapement | 0.249 | 0.170 | 0.248 | 0.331 | 0.000 |
|  | PRH Returns | 0.041 | 0.443 | 0.355 | 0.160 | 0.000 |
| 2004 | Escapement | 0.216 | 0.064 | 0.406 | 0.311 | 0.003 |
|  | PRH Returns | 0.133 | 0.398 | 0.406 | 0.063 | 0.000 |
| 2005 | Escapement | 0.151 | 0.082 | 0.306 | 0.458 | 0.003 |
|  | PRH Returns | 0.116 | 0.572 | 0.284 | 0.028 | 0.000 |
| 2006 | Escapement | 0.109 | 0.052 | 0.632 | 0.206 | 0.000 |
|  | PRH Returns | 0.331 | 0.325 | 0.314 | 0.030 | 0.000 |
| 2007 | Escapement | 0.109 | 0.230 | 0.490 | 0.171 | 0.001 |
|  | PRH Returns | 0.103 | 0.483 | 0.381 | 0.033 | 0.000 |
| 2008 | Escapement | 0.159 | 0.193 | 0.511 | 0.137 | 0.000 |
|  | PRH Returns | 0.221 | 0.497 | 0.279 | 0.002 | 0.000 |
| 2009 | Escapement | 0.091 | 0.136 | 0.688 | 0.083 | 0.001 |
|  | PRH Returns | 0.125 | 0.564 | 0.2410 | 0.071 | 0.000 |
| 2010 | Escapement | 0.020 | 0.269 | 0.441 | 0.265 | 0.006 |
|  | PRH Returns | 0.108 | 0.386 | 0.468 | 0.038 | 0.000 |
| 2011 | Escapement | 0.100 | 0.086 | 0.634 | 0.178 | 0.002 |
|  | PRH Returns | 0.065 | 0.430 | 0.448 | 0.056 | 0.001 |
| $2012{ }^{\text {a }}$ | Escapement | 0.185 | 0.280 | 0.363 | 0.172 | -- |
|  | PRH Returns | 0.178 | 0.539 | 0.210 | 0.072 | -- |
| Mean 1998-2012 | Escapement | 0.148 | 0.141 | 0.459 | 0.248 | 0.003 |
|  | PRH Returns | 0.117 | 0.452 | 0.361 | 0.070 | 0.001 |
| Mean 2007-2012 | Escapement | 0.111 | 0.197 | 0.523 | 0.168 | 0.001 |
|  | PRH Returns | 0.133 | 0.483 | 0.338 | 0.045 | 0.000 |

[^2]Table 39 Age compositions for natural and hatchery origin fall Chinook salmon sampled in the Hanford Reach escapement, Brood Years 2007-2012.

| Brood Year | Origin ${ }^{1}$ | $\mathrm{N}^{2}$ | Male Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural | 1,093 | No otolith data | 0.377 | 0.483 | 0.139 | 0.002 |
|  | Hatchery | 121 |  | 0.801 | 0.116 | 0.083 | 0.000 |
| 2008 | Natural | 1,234 | 0.044 | 0.336 | 0.502 | 0.118 | 0.000 |
|  | Hatchery | 49 | 0.255 | 0.299 | 0.353 | 0.092 | 0.000 |
| 2009 | Natural | 816 | 0.034 | 0.231 | 0.660 | 0.076 | 0.000 |
|  | Hatchery | 139 | 0.033 | 0.270 | 0.678 | 0.019 | 0.000 |
| 2010 | Natural | 2,097 | 0.008 | 0.361 | 0.454 | 0.176 | 0.000 |
|  | Hatchery | 333 | 0.043 | 0.814 | 0.108 | 0.034 | 0.000 |
| 2011 | Natural | 838 | 0.182 | 0.157 | 0.547 | 0.112 | 0.002 |
|  | Hatchery | 72 | 0.113 | 0.232 | 0.577 | 0.078 | 0.000 |
| $2012^{\text {a }}$ | Natural | 857 | 0.058 | 0.528 | 0.319 | 0.094 | -- |
|  | Hatchery | 86 | 0.077 | 0.683 | 0.223 | 0.017 | -- |
| Mean | Natural | 1,156 | 0.065 | 0.332 | 0.494 | 0.119 | 0.001 |
|  | Hatchery | 133 | 0.104 | 0.517 | 0.343 | 0.054 | 0.000 |
|  |  |  | Female Age Composition |  |  |  |  |
| Brood Year | Origin ${ }^{1}$ | $\mathbf{N}^{2}$ | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural | 1,299 | No otolith data | 0.047 | 0.706 | 0.247 | 0.000 |
|  | Hatchery | 167 |  | 0.532 | 0.317 | 0.151 | 0.000 |
| 2008 | Natural | 426 | 0.000 | 0.117 | 0.679 | 0.204 | 0.000 |
|  | Hatchery | 74 | 0.000 | 0.176 | 0.651 | 0.172 | 0.000 |
| 2009 | Natural | 486 | 0.000 | 0.033 | 0.789 | 0.175 | 0.003 |
|  | Hatchery | 188 | 0.000 | 0.060 | 0.918 | 0.021 | 0.000 |
| 2010 | Natural | 1,934 | 0.000 | 0.026 | 0.542 | 0.432 | 0.000 |
|  | Hatchery | 353 | 0.000 | 0.418 | 0.448 | 0.133 | 0.000 |
| 2011 | Natural | 926 | 0.000 | 0.005 | 0.775 | 0.217 | 0.002 |
|  | Hatchery | 118 | 0.000 | 0.022 | 0.782 | 0.195 | 0.000 |
| $2012^{\text {a }}$ | Natural | 1,064 | 0.000 | 0.133 | 0.538 | 0.329 | -- |
|  | Hatchery | 127 | 0.000 | 0.382 | 0.479 | 0.138 | -- |
| Mean | Natural | 1,023 | 0.000 | 0.060 | 0.672 | 0.267 | 0.001 |
|  | Hatchery | 171 | 0.000 | 0.265 | 0.599 | 0.135 | 0.000 |
|  |  |  | Gender Combined Age Composition |  |  |  |  |
| Brood Year | Origin ${ }^{1}$ | $\mathbf{N}^{2}$ | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural | 2,392 | No Otolith Date | 0.201 | 0.602 | 0.196 | 0.001 |
|  | Hatchery | 288 |  | 0.656 | 0.225 | 0.119 | 0.000 |
| 2008 | Natural | 1,660 | 0.022 | 0.230 | 0.587 | 0.160 | 0.002 |
|  | Hatchery | 123 | 0.100 | 0.224 | 0.535 | 0.141 | 0.000 |
| 2009 | Natural | 1,302 | 0.019 | 0.147 | 0.715 | 0.118 | 0.001 |
|  | Hatchery | 327 | 0.012 | 0.136 | 0.831 | 0.021 | 0.000 |
| 2010 | Natural | 4,052 | 0.004 | 0.185 | 0.501 | 0.304 | 0.006 |
|  | Hatchery | 686 | 0.022 | 0.617 | 0.278 | 0.084 | 0.000 |
|  |  |  |  |  |  |  |  |


| Brood Year | Origin $^{\mathbf{1}}$ | $\mathbf{N}^{\mathbf{2}}$ |  | Gender Combined Age Compositions |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2011 | Natural | 1,764 | 0.088 | 0.079 | 0.665 | 0.166 | 0.002 |
|  | Hatchery | 190 | 0.038 | 0.093 | 0.713 | 0.156 | 0.000 |
| $2012^{\mathrm{a}}$ | Natural | 1,921 | 0.030 | 0.336 | 0.426 | 0.208 | -- |
|  | Hatchery | 213 | 0.030 | 0.500 | 0.379 | 0.091 | -- |
| Mean | Natural | 2,182 | 0.033 | 0.196 | 0.583 | 0.192 | 0.002 |
|  | Hatchery | 304.5 | 0.040 | 0.371 | 0.493 | 0.102 | 0.000 |

${ }^{1}$ Origin based on the presence of otoliths marks, hatchery coded-wire tags, and adipose clips present in the sub-sample.
${ }^{2} \mathrm{~N}$ equals the number fish included in the demographic sample for a specific brood year. Sample rates varied between return years; therefore the age composition is based on pooled sample data expanded for total returns by year.
${ }^{\text {a }}$ Does not include age-6 returns

Table 40 Mean fork length (cm) at age (total age) of fall Chinook salmon sampled in the Hanford Reach escapement compared to fall Chinook salmon sampled at Priest Rapids Hatchery, Brood Years 1999-2012. N = sample size and SD = 1 standard deviation.

| Brood Year | Origin | Fall Chinook fork length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | $\left\lvert\, \begin{aligned} & \mathbf{S} \\ & \mathbf{D} \end{aligned}\right.$ | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 1999 | Escapement | 83 | 44 | 4 | 227 | 70 | 6 | 1,423 | 86 | 7 | 1,085 | 93 | 7 | 22 | 103 | 10 |
|  | PRH Returns | 85 | 46 | 5 | 488 | 70 | 5 | 762 | 84 | 6 | 170 | 92 | 6 | 2 | 94 | 11 |
| 2000 | Escapement | 17 | 44 | 4 | 118 | 65 | 7 | 428 | 82 | 6 | 669 | 94 | 8 | 6 | 96 | 9 |
|  | PRH Returns | 25 | 44 | 5 | 136 | 69 | 6 | 196 | 82 | 6 | 58 | 93 | 7 | 2 | 103 | 10 |
| 2001 | Escapement | 32 | 44 | 5 | 251 | 69 | 6 | 1,157 | 84 | 6 | 288 | 93 | 7 | 18 | 97 | 5 |
|  | PRH Returns | 121 | 48 | 4 | 1,040 | 69 | 5 | 628 | 81 | 6 | 183 | 91 | 6 | 9 | 94 | 9 |
| 2002 | Escapement | 31 | 46 | 4 | 229 | 70 | 6 | 194 | 86 | 8 | 239 | 95 | 8 | 2 | 99 | 6 |
|  | PRH Returns | 80 | 52 | 4 | 281 | 70 | 5 | 246 | 84 | 6 | 61 | 91 | 6 | 1 | 73 | -- |
| 2003 | Escapement | 19 | 48 | 5 | 42 | 69 | 7 | 395 | 85 | 6 | 450 | 96 | 8 | 0 | -- | -- |
|  | PRH Returns | 12 | 49 | 6 | 93 | 70 | 6 | 215 | 83 | 6 | 20 | 91 | 4 | 0 | -- | -- |
| 2004 | Escapement | 34 | 47 | 4 | 71 | 68 | 6 | 386 | 84 | 6 | 208 | 94 | 8 | 2 | 91 | 1 |
|  | PRH Returns | 19 | 55 | 4 | 115 | 69 | 5 | 51 | 84 | 5 | 9 | 95 | 7 | 0 | -- | - |
| 2005 | Escapement | 25 | 50 | 5 | 202 | 70 | 6 | 532 | 84 | 7 | 744 | 96 | 8 | 5 | 96 | 6 |
|  | PRH Returns | 31 | 49 | 4 | 429 | 73 | 4 | 428 | 84 | 6 | 180 | 91 | 6 | 0 | -- | -- |
| 2006 | Escapement | 20 | 48 | 4 | 85 | 69 | 6 | 962 | 86 | 6 | 340 | 92 | 7 | 0 | -- | -- |
|  | PRH Returns | 3 | 45 | 3 | 42 | 71 | 4 | 170 | 84 | 6 | 13 | 92 | 7 | 0 | -- | -- |
| 2007 | Escapement | 24 | 46 | 5 | 642 | 72 | 6 | 1,468 | 84 | 7 | 482 | 92 | 7 | 1 | 105 | - |
|  | PRH Returns | 5 | 50 | 4 | 1,149 | 71 | 4 | 1,419 | 80 | 5 | 179 | 87 | 6 | 0 | -- | -- |
| 2008 | Escapement | 34 | 50 | 4 | 243 | 70 | 5 | 620 | 84 | 7 | 72 | 92 | 8 | 1 | 84 | -- |
|  | PRH Returns | 22 | 52 | 5 | 652 | 69 | 4 | 573 | 81 | 6 | 1 | 84 | 0 | 0 | -- | -- |
| 2009 | Escapement | 50 | 48 | 4 | 421 | 69 | 6 | 931 | 81 | 6 | 183 | 92 | 10 | 1 | 73 | -- |
|  | PRH Returns | 308 | 48 | 4 | 1,690 | 68 | 5 | 218 | 77 | 5 | 66 | 86 | 7 | 0 | -- | -- |
| 2010 | Escapement | 63 | 47 | 7 | 1,040 | 68 | 5 | 2,754 | 82 | 7 | 826 | 88 | 7 | 25 | 90 | 6 |
|  | PRH Returns | 883 | 48 | 4 | 1,375 | 69 | 4 | 1,413 | 78 | 5 | 55 | 84 | 4 | 1 | 65 | -- |
| 2011 | Escapement | 58 | 46 | 4 | 266 | 67 | 5 | 1,151 | 80 | 6 | 465 | 88 | 7 | 8 | 91 | 12 |
|  | PRH Returns | 111 | 47 | 3 | 694 | 67 | 4 | 355 | 77 | 5 | 109 | 84 | 6 | 1 | 87 | -- |
| 2012 ${ }^{\text {a }}$ | Escapement | 79 | 47 | 4 | 489 | 67 | 5 | 936 | 80 | 6 | 670 | 85 | 7 | -- | -- | -- |
|  | PRH Returns | 335 | 48 | 5 | 607 | 67 | 5 | 568 | 78 | 5 | 484 | 81 | 6 | -- | -- | -- |
| Mean 99-12 | Escapement | 41 | 47 | 5 | 309 | 69 | 6 | 953 | 83 | 7 | 480 | 92 | 8 | 8 | 93 | 6 |
|  | PRH Returns | 146 | 49 | 4 | 628 | 69 | 5 | 517 | 81 | 6 | 113 | 89 | 6 | 3 | 86 | 8 |
| $\begin{aligned} & \text { Mean } \\ & \text { 07-12 } \end{aligned}$ | Escapement | 51 | 47 | 5 | 517 | 69 | 5 | 1,310 | 82 | 7 | 450 | 90 | 8 | 7 | 89 | 6 |
|  | PRH Returns | 277 | 49 | 4 | 1,028 | 69 | 4 | 758 | 79 | 5 | 149 | 84 | 6 | 1 | 76 | -- |

[^3]Table 41 Mean fork length (cm) at age (total age) of natural and hatchery origin fall Chinook salmon that spawned naturally in the Hanford Reach, Brood Years 2007-2012. N = sample size and SD = 1 standard deviation.

| Brood Year | Origin | Male Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2007 | Natural | No otolith Data |  |  | 364 | 70 | 5 | 205 | 84 | 8 | 143 | 98 | 9 | 0 | -- | -- |
|  | Hatchery |  |  |  | 44 | 72 | 4 | 16 | 82 | 5 | 6 | 94 | 7 | 0 | -- | -- |
| 2008 | Natural | 22 | 49 | 4 | 134 | 69 | 5 | 260 | 85 | 8 | 25 | 99 | 7 | 0 |  |  |
|  | Hatchery | 8 | 52 | 3 | 20 | 69 | 5 | 7 | 86 | 4 | 2 | 91 | 15 | 0 |  |  |
| 2009 | Natural | 3 | 48 | 3 | 325 | 68 | 6 | 123 | 82 | 6 | 40 | 99 | 7 | 0 |  |  |
|  | Hatchery | 2 | 55 | 5 | 34 | 71 | 6 | 21 | 79 | 10 | 2 | 96 | 6 | 0 |  |  |
| 2010 | Natural | 31 | 45 | 4 | 291 | 68 | 7 | 855 | 83 | 8 | 135 | 94 | 8 | 4 | 97 | 8 |
|  | Hatchery | 28 | 49 | 5 | 58 | 69 | 6 | 35 | 79 | 8 | 7 | 92 | 7 | 0 | -- | -- |
| 2011 | Natural | 31 | 45 | 4 | 176 | 66 | 5 | 403 | 81 | 8 | 137 | 94 | 7 | 3 | 104 | 3 |
|  | Hatchery | 27 | 49 | 5 | 19 | 68 | 4 | 31 | 80 | 6 | 7 | 88 | 7 | 0 | -- | -- |
| 2012 ${ }^{\text {a }}$ | Natural | 46 | 47 | 4 | 321 | 67 | 6 | 311 | 80 | 8 | 144 | 92 | 8 | 0 | -- | -- |
|  | Hatchery | 7 | 49 | 5 | 49 | 69 | 5 | 25 | 83 | 6 | 3 | 88 | 10 | 0 | -- | -- |
| Mean | Natural | 27 | 47 | 4 | 268 | 68 | 6 | 360 | 83 | 8 | 106 | 96 | 8 | 1 | 102 | 4 |
|  | Hatchery | 14 | 51 | 5 | 37 | 70 | 5 | 23 | 82 | 7 | 5 | 91 | 8 | 0 | -- | -- |
|  |  | Female Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brood |  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |
| Year | Origin | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2007 | Natural | No otolith Data |  |  | 83 | 72 | 5 | 375 | 83 | 5 | 314 | 89 | 4 | 0 | -- | -- |
|  | Hatchery |  |  |  | 48 | 72 | 4 | 48 | 80 | 4 | 8 | 85 | 5 | 0 | -- | -- |
| 2008 | Natural | 0 |  |  | 36 | 70 | 3 | 344 | 83 | 5 | 49 | 88 | 5 | 1 | 84 | 0 |
|  | Hatchery | 0 |  |  | 23 | 70 | 5 | 21 | 82 | 4 | 7 | 85 | 6 | 0 |  |  |
| 2009 | Natural | 0 |  |  | 44 | 71 | 5 | 105 | 80 | 4 | 82 | 87 | 11 | 1 | 73 | 0 |
|  | Hatchery | 0 |  |  | 12 | 68 | 4 | 49 | 78 | 6 | 4 | 85 | 4 | 0 |  |  |
| 2010 | Natural | 0 | -- | -- | 33 | 71 | 5 | 999 | 87 | 5 | 528 | 85 | 4 | 20 | 89 | 5 |
|  | Hatchery | 0 | -- | -- | 22 | 69 | 4 | 144 | 79 | 5 | 29 | 82 | 4 | 0 | -- | -- |
| 2011 ${ }^{\text {a }}$ | Natural | 0 | -- | -- | 7 | 67 | 5 | 597 | 80 | 5 | 283 | 85 | 5 | 5 | 84 | 7 |
|  | Hatchery | 0 | -- | -- | 4 | 65 | 2 | 72 | 77 | 4 | 34 | 84 | 4 | 0 | -- | -- |
| 2012 ${ }^{\text {a }}$ | Natural | 0 | -- | -- | 77 | 68 | 3 | 449 | 80 | 4 | 480 | 83 | 6 | 0 | -- | -- |
|  | Hatchery | 0 | -- | -- | 42 | 68 | 3 | 83 | 78 | 6 | 38 | 81 | 5 | 0 | -- | -- |
| Mean | Natural | 0 | -- | -- | 47 | 70 | 4 | 478 | 82 | 5 | 291 | 86 | 6 | 5 | 83 | 3 |
|  | Hatchery | 0 | -- | -- | 26 | 69 | 3 | 75 | 80 | 5 | 24 | 83 | 5 | 0 | 0 | 0 |

## Table 41 continues onto next page

Table 41 Continued

| Brood Year | Origin | Gender Combined Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2007 | Natural | No otolith Data |  |  | 447 | 70 | 5 | 580 | 83 | 6 | 457 | 92 | 6 | 0 | -- | -- |
|  | Hatchery |  |  |  | 92 | 72 | 4 | 64 | 81 | 4 | 28 | 87 | 6 | 0 | -- | -- |
| 2008 | Natural | 22 | 49 | 4 | 170 | 69 | 5 | 604 | 84 | 6 | 74 | 92 | 6 | 1 | 84 | 0 |
|  | Hatchery | 8 | 52 | 3 | 43 | 70 | 5 | 28 | 83 | 4 | 9 | 86 | 8 | 0 |  |  |
| 2009 | Natural | 3 | 48 | 3 | 369 | 68 | 6 | 228 | 81 | 5 | 122 | 91 | 10 | 1 | 73 | 0 |
|  | Hatchery | 2 | 55 | 5 | 46 | 70 | 5 | 70 | 78 | 7 | 6 | 89 | 5 | 0 |  |  |
| 2010 | Natural | 31 | 45 | 4 | 324 | 69 | 6 | 1854 | 82 | 8 | 763 | 88 | 5 | 24 | 90 | 7 |
|  | Hatchery | 29 | 50 | 6 | 80 | 69 | 6 | 179 | 79 | 7 | 36 | 84 | 5 | 0 | -- | -- |
| 2011 ${ }^{\text {a }}$ | Natural | 31 | 45 | 4 | 183 | 66 | 5 | 1000 | 80 | 6 | 420 | 88 | 7 | 8 | 91 | 12 |
|  | Hatchery | 28 | 50 | 6 | 23 | 67 | 4 | 103 | 78 | 5 | 41 | 84 | 5 | 0 | -- | -- |
| 2012 ${ }^{\text {a }}$ | Natural | 46 | 47 | 4 | 389 | 67 | 5 | 760 | 80 | 6 | 624 | 85 | 7 | -- | -- | -- |
|  | Hatchery | 7 | 49 | 5 | 91 | 68 | 4 | 108 | 79 | 6 | 41 | 81 | 6 | -- | -- | -- |
| Mean | Natural | 27 | 47 | 4 | 313 | 68 | 5 | 838 | 82 | 6 | 414 | 89 | 7 | 7 | 89 | 10 |
|  | Hatchery | 15 | 51 | 5 | 63 | 69 | 5 | 92 | 80 | 6 | 25 | 86 | 6 | 0 | 0 | 0 |

${ }^{\text {a }}$ Brood year does not include age- 6 returns

### 13.4 Gender Composition for Adult Escapement

Prior to return year 2012, the gender ratio comparisons between fall Chinook salmon recovered at PRH and the Hanford Reach stream survey were based on the survey type (i.e., stream or hatchery). Although the estimates based on this method are imperfect, we continue to present this information to maintain the longest data set available (Table 42).
Gender ratios (male/females) by brood year and origin of adult fall Chinook salmon sampled in the Hanford Reach carcass survey are given in (Table 43). Annually, higher male to female ratios have been observed in the natural origin fish than that of the hatchery origin fish. This may be the result of earlier age of maturity of hatchery origin fish and a size related bias of recovering carcasses in the Hanford Reach.

Table 42 Comparisons male to female ratio of fall Chinook salmon sampled at Priest Rapids Hatchery and in the Hanford Reach stream surveys, Brood Years 2007-2012.

| Brood Year | Origin | Male ${ }^{1}$ : Female Ratio |
| :---: | :---: | :---: |
| 1996 | Stream | 0.94:1 |
|  | Hatchery | 1.98:1 |
| 1997 | Stream | 0.48:1 |
|  | Hatchery | 1.88:1 |
| 1998 | Stream | 0.66:1 |
|  | Hatchery | 1.38:1 |
| 1999 | Stream | 0.71:1 |
|  | Hatchery | 2.15:1 |
| 2000 | Stream | 1.51:1 |
|  | Hatchery | 2.40:1 |
| 2001 | Stream | 0.67:1 |
|  | Hatchery | 2.31:1 |
| 2002 | Stream | 1.40:1 |
|  | Hatchery | 1.94:1 |
| 2003 | Stream | 1.25:1 |
|  | Hatchery | 1.64:1 |
| 2004 | Stream | 1.17:1 |
|  | Hatchery | 1.63:1 |
| 2005 | Stream | 0.87:1 |
|  | Hatchery | 2.15:1 |
| 2006 | Stream | 0.75:1 |
|  | Hatchery | 2.57:1 |
| 2007 | Stream | 0.78:1 |
|  | Hatchery | 1.60:1 |
| 2008 | Stream | 0.82:1 |
|  | Hatchery | 1.89:1 |
| 2009 | Stream | 1.07:1 |
|  | Hatchery | 2.57:1 |
| 2010 | Stream | 0.70:1 |
|  | Hatchery | 1.47:1 |
| 2011 | Stream | 0.71:1 |
|  | Hatchery | 2.00:1 |
| 2012 ${ }^{\text {a }}$ | Stream | 1:1.14 |
|  | Hatchery | 1.91:1 |
| Mean | Stream | 0.92:1 |
|  | Hatchery | 1.97:1 |

[^4]Table 43 Comparison male to female ratio of fall Chinook salmon sampled in the Hanford Reach stream surveys, Brood Years 2007-2012.

| Brood Year | Origin | Male ${ }^{1}$ : Female Ratio |
| :---: | :---: | :---: |
| $2007{ }^{\text {a }}$ | Natural | 0.86:1.00 |
|  | Hatchery | 0.74:1.00 |
| 2008 | Natural | 1.07:1.00 |
|  | Hatchery | 0.64:1.00 |
| 2009 | Natural | 1.37:1.00 |
|  | Hatchery | 0.56:1.00 |
| 2010 | Natural | 1.02:1.00 |
|  | Hatchery | 1.01:1.00 |
| 2011 | Natural | 0.94:1.00 |
|  | Hatchery | 0.51:1.00 |
| $2012{ }^{\text {b }}$ | Natural | 1.06:1.00 |
|  | Hatchery | 0.65:1.00 |
| Mean | Natural | 1.05:1.00 |
|  | Hatchery | 0.69:1:00 |

${ }^{1}$ Includes both adult males and jacks. ${ }^{\text {a }}$ Does not include age-2. ${ }^{\mathrm{b}}$ Includes age-2 through 5.

### 13.5 Egg Retention

All female Chinook included in the demographic sample for the Hanford Reach stream surveys were examined for egg retention to assess spawn success. The females sampled were partitioned into the egg retention categories of $0 \%, 25 \%, 50 \%, 75 \%$ and $100 \%$. The assignment of origin for each female for years 2010 and 2011 were based on the presence or absence of an adipose fin. The adipose intact group may include non-adipose clipped fish from PRH. A combination of hatchery marks (i.e., adipose clips, CWTs, and otolith marks were used to identify hatchery origin fish in years 2013-2017. For all years, we assume that fish not possessing any hatchery marks are natural origin fish.

The assessment of egg retention is compromised by the loss of eggs during the collection and transport of carcasses prior to sampling. Therefore, our estimates of egg retention are likely to be underestimates and our estimates of egg loss are likely to be overestimates. In addition, the methods for quantifying egg retention and assignment of origin for each female have varied between years. The amount of egg retention for years 2010 through 2013 were determined by visual estimates; whereas, during 2014 through 2017, the amount of retention was based on egg counts when the gametes were not completely intact. For these recent data sets, the percent of egg retention was calculated by dividing the amount of egg retained by an estimated fecundity based on length versus fecundity regressions by origin (Hatchery or Natural). An explanation of these regressions is provided in the fecundity section of this report.

Starting in return year 2015, staff recorded visual observations of egg retention based on the standard egg retention categories to make comparisons with egg retention based on egg counts. The data from the egg counts were categorized into the standard egg retention categories based on the following ranges: $1=100-88 \%, 2=87-63 \%, 3=62-38 \%, 4=37-11 \%$, and $5=10-0 \%$. This comparison may allow us to assess the egg retention estimates based on methods used prior to 2015. The difference between two methods was less than 1 percentage point by category for each year (Table 44), which provides some confidence that the visual methods of the past may provide reasonable indices of spawning success.

Table 44 Comparison of egg retention of natural and hatchery origin fall Chinook sampled in the Hanford Reach stream survey, Return Years 2015-2017.

| Survey Year | Egg Retention Categories | \% by Category based on Egg Counts | \% by Category based on Visual Observations | Difference between Actual and Observed (\%) |
| :---: | :---: | :---: | :---: | :---: |
| 2015 | 1-100\% | 0.0 | 0.4 | -0.4 |
|  | 2-75\% | 0.3 | 0.5 | -0.2 |
|  | 3-50\% | 0.6 | 0.6 | 0.0 |
|  | 4-25\% | 1.5 | 1.6 | -0.1 |
|  | 5-0\% | 97.7 | 96.9 | 0.8 |
|  | N = 1,405 |  |  |  |
| 2016 | 1-100\% | 0.0 | 0.2 | -0.2 |
|  | 2-75\% | 0.2 | 0.3 | -0.1 |
|  | 3-50\% | 0.6 | 1.1 | -0.5 |
|  | 4-25\% | 1.8 | 1.2 | 0.6 |
|  | 5-0\% | 97.4 | 97.2 | 0.2 |
|  | N = 995 |  |  |  |
| 2017 | 1-100\% | 0.0 | 0.0 | 0.0 |
|  | 2-75\% | 0.2 | 0.2 | 0.0 |
|  | 3-50\% | 0.3 | 0.3 | 0.1 |
|  | 4-25\% | 1.3 | 1.1 | 0.2 |
|  | 5-0\% | 98.2 | 98.5 | -0.3 |
|  | N = 1,180 |  |  |  |

The adjusted spawn successes for the escapement between years 2010 and 2017 were generally greater than $97 \%$ (Table 45). The spawn success was lower for both hatchery and natural origin females in return year 2013 compared to other years. These observations coincide with an elevated pHOS which most likely resulted from high hatchery fish escapement and restrictive operations of the PRH volunteer trap during 2013. It is possible that a portion of these PRH origin females which were unable to enter the trap, died without spawning, and ended up being surveyed in the Hanford Reach carcass survey. The spawn success was high during 2014 and 2015 despite historically high record escapements to the Hanford Reach.

Table 45 Comparison of egg retention of natural and hatchery origin fall Chinook sampled in the Hanford Reach stream survey, Return Years 2010-2017.

| Return Year | Origin | Females Sampled | Egg Retention Categories |  |  |  |  | No Egg <br> Retention (\%) | Adj Spawn Success for Escapement (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 0 \% | $25 \%$ | $50 \%$ | 75\% | 100\% |  |  |
| 2004 | Combined | 1,176 | 1,151 | NA | 21 | NA | 4 | 97.9 | 98.8 |
| 2005 | Combined | 1,323 | 1,310 | NA | 6 | NA | 7 | 99.0 | 99.2 |
| 2006 | Combined | 352 | 343 | NA | 8 | NA | 1 | 97.4 | 98.6 |
| 2007 | Combined | 454 | 443 | NA | 8 | NA | 3 | 97.6 | 98.5 |
| 2008 | Combined | No spawn success data collected |  |  |  |  |  |  |  |
| 2009 | Combined | 499 | 484 | NA | 5 | NA | 10 | 97.0 | 97.5 |
| 2010 | Combined | 1,173 | 1,147 | 6 | 13 | 1 | 6 | 97.8 | 98.7 |
| 2011 | Combined | 1,264 | 1,203 | 1 | 52 | 5 | 3 | 95.2 | 97.4 |
| $2012{ }^{\text {b }}$ | Natural | 681 | 658 | 14 | 5 | 1 | 3 | 96.6 | 98.6 |
|  | Hatchery | 90 | 89 | 0 | 0 | 0 | 1 | 98.9 | 98.9 |
|  | Total | 771 | 747 | 14 | 5 | 1 | 4 | 96.9 | 98.6 |
| $2013{ }^{\text {b }}$ | Natural | 461 | 392 | 51 | 9 | 3 | 6 | 85.0 | 94.5 |
|  | Hatchery | 224 | 144 | 39 | 11 | 13 | 17 | 64.3 | 81.3 |
|  | Total | 685 | 536 | 90 | 20 | 16 | 23 | 78.2 | 90.1 |
| $2014{ }^{\text {b }}$ | Natural | 1,082 | 1,074 | 1 | 0 | 0 | 7 | 99.3 | 99.3 |
|  | Hatchery | 153 | 141 | 3 | 0 | 0 | 9 | 92.2 | 93.6 |
|  | Total | 1,235 | 1,215 | 4 | 0 | 0 | 16 | 98.4 | 98.6 |
| $2015{ }^{\text {b }}$ | Natural | 1256 | 1237 | 14 | 3 | 2 | 0 | 98.5 | 99.5 |
|  | Hatchery | 149 | 135 | 7 | 5 | 2 | 0 | 90.6 | 96.1 |
|  | Total | 1,405 | 1,372 | 21 | 8 | 4 | 0 | 97.7 | 99.1 |
| $2016{ }^{\text {b }}$ | Natural | 857 | 842 | 7 | 3 | 1 | 0 | 98.2 | 99.5 |
|  | Hatchery | 138 | 127 | 11 | 3 | 1 | 0 | 92 | 96.4 |
|  | Total | 995 | 969 | 18 | 6 | 2 | 0 | 97.4 | 99.1 |
| $2017^{\text {b }}$ | Natural | 1,071 | 1,062 | 8 | 1 | 0 | 0 | 99.2 | 99.8 |
|  | Hatchery | 109 | 100 | 5 | 2 | 2 | 0 | 91.7 | 96.6 |
|  | Total | 1,180 | 1,162 | 13 | 3 | 2 | 0 | 98.5 | 99.5 |
| Mean Natural Spawn Success (RY 2012 - 2016) |  |  |  |  |  |  |  | 96.1 | 98.5 |
| Mean Hatchery Spawn Success (RY 2012-2016) |  |  |  |  |  |  |  | 88.3 | 93.8 |
| Mean Combined Spawn Success (RY 2010-2016) |  |  |  |  |  |  |  | 95.0 | 97.7 |

The measure for reporting egg retention changed from that used for previous years beginning in 2010
${ }^{\mathrm{b}}$ Origins were determined the presence or absence of otolith marks, adipose clips and CWTs

### 14.0 Contribution to Fisheries

The contribution of fish produced at PRH to fisheries was estimated by querying the Regional Mark Processing Center (RMPC) database. This is central repository for all CWT and otherwise associated release, catch, sample, and recovery data of anadromous salmonids in the greater Pacific Coast Region of the United States of America. The Regional Mark Information System database (RMIS) within the RMPC provides specific recovery data for individual tag codes,
along with the sample rate used to derive the estimated total number of recoveries by fishery type.
The CWT data reported to RMPC are expanded by sample rates generated by the agency reporting the data. In some cases, the estimated number of tags reported is less than the number actually observed. This typically occurs when the sample rate is unknown, not reported, or biased (Gilbert Lensegrav, WDFW, personal communication). In these instances, the observed number was used instead of the estimated number to calculate the numbers of PRH origin fish recovered by location.
The RMIS database was queried on April 23, 2018 to provide CWT recoveries for active broods of PRH origin fish. The database for the 2011 brood should be complete for age-2 through age- 5 . The age-6 recovered during RY2017 may not be included until January 1, 2019 due to the lag in reporting field data to RMPC.
Beginning with the 2010 release year, portions of the non-adipose clipped smolts released from PRH received a CWT as part of a double index tag (DIT) study to evaluate the effect of various mark-selective fisheries occurring in Oregon, Washington, and British Columbia waters (PSC 2013). We are currently reviewing the data reported to the RMPC database to evaluate the results of the double index tagging for the PRH origin fish. Data for brood years 2009 through 2014 (some are incomplete) show that adipose clipped fish from the DIT groups are being recovered in mark selective fisheries occurring in ocean, marine, and freshwater zones. Comparisons of the demographics between the DIT groups recovered at PRH are very similar (Appendix I).
Therefore, mark selective fisheries do not appear to markedly influence the demographic data collected at PRH.
Fall Chinook salmon released from PRH supplement Pacific Ocean harvest for both commercial and sport fisheries from Washington to Southeast Alaska as well as Columbia River commercial, sport, and treaty tribal harvest. The Hanford Reach sport fishery for fall Chinook salmon is an extremely popular fishery. This fishery typically runs annually from August 1 to late October. In 2017, an estimated 12,368 fall Chinook salmon were harvested during this fishery; 11,496 adults and 872 jacks. Estimates generated from CWT recoveries from the Hanford Reach sport fishery suggest that $12.8 \%$ ( 1,582 fish) of the total sport harvest in the Hanford Reach was comprised of fall Chinook salmon released from PRH (Table 46). Likewise, fall Chinook salmon released from Ringold Springs Hatchery comprised 6.8\% (843 fish) of the sport fishery. Strays from other hatcheries combined represent $0.8 \%$ ( 35 fish) of the harvest. Sport harvest monitoring in the Hanford Reach and lower Yakima includes surveying both adipose intact and adipose clipped fish for CWT sampling. Recent data from otolith sampling indicates that CWT expansions may underestimate the number of PRH origin fall Chinook salmon annually returning to PRH. A similar situation may occur when evaluating hatchery contributions to the sport fishery.
The CWT data for PRH origin fall Chinook salmon that were marked with an adipose clip were reviewed to assess contributions to marine and freshwater, commercial, tribal, and sport fisheries. The largest proportion of the harvest of PRH origin fall Chinook salmon occurred in ocean fisheries followed by Zone-6 tribal harvest. For brood years 1997 through 2011, 49\% of the reported harvest was taken in ocean fisheries and the other $51 \%$ in the Columbia River fisheries (Table 47). The adipose clip CWT rate for the 2009 and 2011 broods notably increased from previous brood years. Not all CWT surveys locations check harvested adipose intact fish for the presence of a CWT. Therefore, the data presented in Table 47 includes harvest estimates based on recoveries of adipose clipped CWT tagged fish.

Table 46 Hatchery fall Chinook salmon contributions to harvest in the Hanford Reach fall Chinook salmon fishery. Coded-wire tag recoveries provided from RMIS database were expanded by sample rate and juvenile tag rate, Return Years 2003-2017.

| $\begin{gathered} \text { Return } \\ \text { Year } \end{gathered}$ | Harvest \& Sampling |  |  | CWT Expansions |  |  | \% of Harvest |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Harvest | Sampled | \% | PRH | RSH | Other Hatcheries | PRH | RSH | Other <br> Hatcheries |
| 2003 | 7,190 | 1,848 | 25.7 | 510 | 424 | 43 | 7.1 | 5.9 | 0.6 |
| 2004 | 8,787 | 2,255 | 25.7 | 276 | 62 | 23 | 3.1 | 0.7 | 0.3 |
| 2005 | 7,974 | 1,834 | 23.0 | 1,200 | 265 | 35 | 15.0 | 3.3 | 0.4 |
| 2006 | 4,508 | 1,296 | 28.7 | 683 | 66 | 10 | 15.1 | 1.5 | 0.2 |
| 2007 | 6,466 | 1,812 | 28.0 | 929 | 50 | 89 | 14.4 | 0.8 | 1.4 |
| 2008 | 7,013 | 1,593 | 22.7 | 304 | 66 | 22 | 4.3 | 0.9 | 0.3 |
| 2009 | 8,806 | 1,741 | 19.8 | 520 | 0 | 10 | 5.9 | 0.0 | 0.1 |
| 2010 | 12,499 | 2,475 | 19.8 | 1,157 | 399 | 10 | 9.3 | 3.2 | 0.1 |
| 2011 | 14,262 | 2,715 | 19.0 | 1,558 | 663 | 121 | 10.9 | 4.6 | 0.8 |
| 2012 | 18,854 | 3,615 | 19.2 | 3,974 | 1,974 | 237 | 21.1 | 10.5 | 1.3 |
| 2013 | 27,630 | 5,555 | 20.2 | 6,570 | 3,947 | 537 | 23.8 | 14.3 | 1.9 |
| 2014 | 32,417 | 8,319 | 25.7 | 3,987 | 1,419 | 332 | 12.3 | 4.4 | 1.0 |
| 2015 | 35,419 | 10,327 | 29.2 | 4,144 | 992 | 319 | 11.7 | 2.8 | 0.9 |
| 2016 | 17,927 | 5,544 | 30.9 | 2,177 | 822 | 339 | 12.1 | 4.6 | 1.9 |
| 2017 | 12,368 | 4,435 | 38.6 | 1,585 | 843 | 105 | 12.8 | 6.8 | 0.8 |
| Mean | 14,982 | 3,691 | 25.9 | 1,972 | 799 | 149 | 12.0 | 4.3 | 0.8 |

Table $47 \quad$ Priest Rapids Hatchery coded-wire tag recoveries provided from RMIS by brood year and harvest type expanded by sample rate and juvenile tag rate, Brood Years 1997-2011. Data only includes coded-wire tag recoveries from adipose clipped fish expanded by the juvenile tag rate.

| Brood Year | Ocean Fisheries |  | Columbia River Fisheries |  |  |  |  |  | Total Harvest | Ad- <br> CWT <br> Rate |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Tribal |  | Commercial |  | Recreational |  |  |  |
|  | \# | \% | \# | \% | \# | \% | \# | \% |  |  |
| 1997 | 1,100 | 36.7 | 1,506 | 50.2 | 304 | 10.1 | 91 | 3.0 | 3,001 | 0.030 |
| 1998 | 6,580 | 48.4 | 3,956 | 29.1 | 1,066 | 7.8 | 1,981 | 14.6 | 13,583 | 0.030 |
| 1999 | 14,190 | 54.6 | 5,908 | 22.8 | 2,410 | 9.3 | 3,458 | 13.3 | 25,966 | 0.029 |
| 2000 | 4,938 | 61.5 | 1,583 | 19.7 | 1,099 | 13.7 | 412 | 5.1 | 8,032 | 0.032 |
| 2001 | 17,758 | 56.5 | 6,612 | 21.1 | 1,554 | 4.9 | 5,484 | 17.5 | 31,408 | 0.052 |
| 2002 | 3,779 | 50.6 | 1,240 | 16.6 | 576 | 7.7 | 1,869 | 25.0 | 7,464 | 0.052 |
| 2003 | 1,871 | 54.6 | 570 | 16.6 | 226 | 6.6 | 757 | 22.1 | 3,424 | 0.059 |
| 2004 | 562 | 49.3 | 364 | 31.9 | 214 | 18.8 | 0 | 0.0 | 1,140 | 0.059 |
| 2005 | 10,699 | 52.1 | 5,975 | 29.1 | 998 | 4.9 | 2,871 | 14.0 | 20,543 | 0.030 |
| 2006 | 1,023 | 44.1 | 713 | 30.7 | 288 | 12.4 | 298 | 12.8 | 2,322 | 0.029 |
| 2007 | 13,838 | 44.4 | 10,620 | 34.1 | 2,160 | 6.9 | 4,523 | 14.5 | 31,141 | 0.030 |
| 2008 | 5,763 | 43.7 | 4,447 | 33.7 | 887 | 6.7 | 2,080 | 15.8 | 13,177 | 0.032 |
| 2009 | 24,872 | 43.4 | 21,121 | 36.8 | 2,581 | 4.5 | 8,761 | 15.3 | 57,335 | 0.091 |
| 2010 | 46,584 | 43.5 | 34,275 | 32.0 | 7,886 | 7.4 | 18,299 | 17.1 | 107,044 | 0.089 |
| 2011 | 18,235 | 44.2 | 11,813 | 28.6 | 3,874 | 9.4 | 7,310 | 17.7 | 41,232 | 0.084 |
| Mean | 11,453 | 48.5 | 7,380 | 28.9 | 1,742 | 8.7 | 3,880 | 13.9 | 24,454 | 0.049 |

### 15.0 Straying

The distribution of PRH origin fish spawning in areas outside of the target stream is presented to assess the level of straying and potential impacts on other populations. The presumptive target spawning location for PRH origin fish includes the section of Columbia River from McNary Dam to Wanapum Dam as well as the lower Yakima River below Prosser Dam.

The spawning escapement of PRH origin fish by brood year is determined from CWT recoveries collected during spawning surveys. The CWT recoveries are expanded by the juvenile mark rates and survey sampling rates to estimate the number of PRH origin fish recovered on spawning grounds.

The stray rates (i.e., fish that spawned outside of the presumptive target area $\div$ total escapement) for each brood year were calculated from the estimated recoveries of PRH origin fish from spawning grounds within and outside of the presumptive target area. CWT recoveries at nontarget hatcheries and adult fish traps are not included. These fish were not considered strays because these fish were not able to leave the facilities on their own volition.

There are two target rates for recipient population straying given in the Monitoring and Evaluation Plan for PUD Hatchery Programs (Hillman et al. 2017):
1). Stray rate for PRH origin fall Chinook salmon should be less than $5 \%$ of the spawning escapement for other non-target independent populations based on run year.
2). Stray rate for PRH origin fall Chinook salmon should be less than $10 \%$ of the spawning escapement of any non-target streams within the independent population based on run year.

In addition, the donor stray rate for each hatchery brood year is also monitored. With one exception (brood year 2006), less than 5\% of the PRH origin returns for each brood year are estimated to have spawned outside of the presumptive target spawning area (Table 48). Likewise, the CWT recoveries by return year for presumptive non-target streams or areas suggest that PRH fall Chinook salmon seldom exceeded more than $5 \%$ of the spawning escapement for other independent populations of fall Chinook salmon. However, for multiple return years, greater than $5 \%$ of the spawning escapement for the Chelan River may have consisted of PRH origin fall Chinook salmonError! Reference source not found.). The Chelan River spawning population is a mix of both summer and fall Chinook salmon strays and is not considered an independent population. This location was included to show contributions of PRH strays to this group of fish.

Table 48 Estimated number and proportions of Priest Rapids Hatchery fall Chinook salmon spawning escapement to Priest Rapids Hatchery and stream within and outside of the presumptive target stream by brood year (1992-2011). Coded-wire tag recoveries are expanded by juvenile mark rate and survey sample rate for each brood year.

| Brood Year | Number of PRH Origin Recoveries | Homing |  |  |  | Straying <br> Outside of Target Stream |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Target Hatchery |  | Target Stream ${ }^{1}$ |  |  |  |
|  |  | Number | Proportion | Number | Proportion | Number | Proportion |
| 1992 | 9,037 | 7,630 | 0.844 | 1,037 | 0.115 | 370 | 0.041 |
| 1993 | 25,965 | 21,144 | 0.814 | 4,821 | 0.186 | 0 | 0.000 |
| 1994 | 1,693 | 1,385 | 0.818 | 308 | 0.182 | 0 | 0.000 |
| 1995 | 30,655 | 23,414 | 0.764 | 7,207 | 0.235 | 34 | 0.001 |
| 1996 | 13,551 | 10,034 | 0.740 | 3,517 | 0.260 | 0 | 0.000 |
| 1997 | 3,173 | 2,690 | 0.848 | 483 | 0.152 | 0 | 0.000 |
| 1998 | 18,167 | 11,833 | 0.651 | 5,867 | 0.323 | 467 | 0.026 |
| 1999 | 27,334 | 15,467 | 0.566 | 11,867 | 0.434 | 0 | 0.000 |
| 2000 | 4,759 | 3,690 | 0.775 | 1,069 | 0.225 | 0 | 0.000 |
| 2001 | 25,375 | 15,875 | 0.626 | 9,469 | 0.373 | 31 | 0.001 |
| 2002 | 5,288 | 3,769 | 0.713 | 1,519 | 0.287 | 0 | 0.000 |
| 2003 | 3,034 | 2,034 | 0.670 | 949 | 0.313 | 51 | 0.017 |
| 2004 | 1,133 | 1,133 | 1.000 | 0 | 0.000 | 0 | 0.000 |
| 2005 | 21,378 | 17,103 | 0.800 | 4,241 | 0.198 | 34 | 0.002 |
| 2006 | 1,001 | 634 | 0.633 | 0 | 0.000 | 367 | 0.367 |
| 2007 | 22,206 | 19,220 | 0.866 | 2,964 | 0.133 | 22 | 0.001 |
| 2008 | 11,866 | 9,002 | 0.759 | 2,864 | 0.241 | 0 | 0.000 |
| 2009 | 28,153 | 13442 | 0.477 | 14,689 | 0.522 | 22 | 0.001 |
| 2010 | 107,961 | 67,060 | 0.621 | 40,574 | 0.376 | 327 | 0.003 |
| 2011 | 49,396 | 36,043 | 0.730 | 13,258 | 0.268 | 95 | 0.002 |
| Mean | 20,556 | 14,130 | 0.736 | 6,335 | 0.241 | 91 | 0.023 |

[^5]Table 49 Proportion of fall/summer Chinook spawning populations by return year (2000-2016) comprised of Priest Rapids Hatchery fall Chinook from 19982011 brood releases based on coded-wire tag recoveries.

| Return Year | Presumptive Non-Target Stream |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Snake Fall Chinook |  | Yakima Fall Chinook |  | Wenatchee Summer Chinook |  | Entiat River ${ }^{1}$ |  | Chelan River ${ }^{1}$ |  | Methow Summer Chinook |  | Okanogan Summer Chinook |  |
|  | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| 2000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2001 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 334 | 0.339 | 0 | 0.000 | 0 | 0.000 |
| 2002 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 274 | 0.471 | 0 | 0.000 | 0 | 0.000 |
| 2003 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2004 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2005 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2006 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2007 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2008 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 57 | 0.016 |
| 2009 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 228 | 0.177 | 0 | 0.000 | 0 | 0.000 |
| 2010 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 359 | 0.322 | 0 | 0.000 | 0 | 0.000 |
| 2011 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2012 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 50 | 0.038 | 0 | 0.000 | 0 | 0.000 |
| 2013 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 102 | 0.062 | 0 | 0.000 | 0 | 0.000 |
| 2014 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 83 | 0.075 | 0 | 0.000 | 0 | 0.000 |
| 2015 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 22 | 0.019 | 320 | 0.222 | 0 | 0.000 | 0 | 0.000 |
| 2016 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| Mean | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 1 | 0.001 | 103 | 0.100 | 0 | 0.000 | 3 | 0.001 |

${ }^{1}$ The Chelan and Entiat River spawning populations are a mix of both summer and fall Chinook salmon strays and are not considered independent populations. These locations were included to show contributions of PRH strays to these groups of fish.

As previously described in Section 4, approximately 3,000 smolts at PRH were annually PIT tagged at PRH from brood years 1995 through 2010. The annual release of PIT tagged smolts was increased to $\sim 43,000$ beginning with brood year 2011. The last known observations of individual PIT tag adult fall Chinook salmon originating from PRH at detection locations above McNary Dam are given in Table 50 for brood years 1999 through 2014. The number of observed PRH PIT tagged adults is increasing as anticipated due to the increased number of tags.
The majority of the PIT tagged PRH adults observed at McNary Dam have been observed at Priest Rapids Dam (PRD) adult fishways and/or PRH. Very few fish have been detected in the Snake River, which an the area of high concern for straying. In addition, notable proportions of the returns for several brood years have been observed at sites upstream of PRD. It is unclear whether fish spawned outside of the target areas because fish could return to a target location after being detected at a PIT tag array outside of the target stream without being detected again. Observations for PIT tagged presumptive Hanford Reach natural origin adults show very few detections above PRD.

Table 50 Last observations of unique PIT tagged adult fall Chinook from Priest Rapids Hatchery at detection sties upstream of McNary Dam, Brood Years 1999-2014.

|  | \# PIT | Number of last known detections of unique Priest Rapids Origin PIT tags by site |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year | tagged | MCN | ICH | PRO | PRH | PRD | RIA | LWE | RRF | EBO | ENL | WEA | LMR | Total |
| 1999 | 3000 | 9 |  |  |  | 7 | 1 |  |  |  |  | 1 |  | 18 |
| 2000 | 3000 | 3 |  |  |  | 4 |  |  |  |  |  |  |  | 7 |
| 2001 | 3000 | 5 |  |  |  | 6 |  |  |  |  |  |  |  | 11 |
| 2002 | 3000 | 7 |  |  |  | 1 |  |  |  |  |  |  |  | 8 |
| 2003 | 3000 |  |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 2004 | 3000 |  |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 2005 | 3000 | 9 |  |  |  | 4 | 1 |  |  |  |  |  |  | 14 |
| 2006 | 3000 |  |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 2007 | 3,000 | 20 |  |  | 1 | 12 | 2 |  | 2 |  |  | 1 | 1 | 39 |
| 2008 | 2,994 | 5 |  |  |  | 6 |  |  | 1 |  |  |  |  | 12 |
| 2009 | 1,995 | 4 |  |  | 16 |  | 2 |  |  |  |  |  |  | 22 |
| 2010 | 3,000 | 6 |  |  | 36 | 22 | 5 | 1 | 4 |  |  | 3 |  | 77 |
| 2011 | 42,844 | 215 |  |  | 78 | 222 | 8 | 3 | 29 |  |  | 22 | 5 | 582 |
| 2012 (age 2-5) | 42,908 | 112 |  |  | 442 | 126 | 7 | 1 | 21 |  |  | 14 | 3 | 726 |
| 2013 (age 2-4) | 42,988 | 74 |  |  | 3 | 32 | 1 |  | 7 |  |  | 1 | 1 | 119 |
| 2014 (age 2-3) | 42,621 | 8 |  |  |  | 7 |  |  |  |  |  |  |  | 15 |
| MCN | McNary Dam | m Adult | Fishway | RKM |  |  |  | LWE | Lower W | enatchee | River R | KM 754 |  |  |
| ICH | Ice Harbor D | Dam Adult | Fishw | ays RKM | 522 |  |  | RRF | Rocky R | each Dan | Adult | ishway R | KM 763 |  |
| PRO | Prosser Dive | ersion Da | m RKM | 539 |  |  |  | EBO | East Ban | $k$ Hatche | y Outfa | RKM 7 |  |  |
| PRH | Priest Rapid | s Hatcher | y Outfal | RKM |  |  |  | ENL | Lower E | tiat Rive | er RKM |  |  |  |
| PRD | Priest Rapid | s Dam A | dult Fish | ways R | M 639 |  |  | WEA | Well Da | Adult | Fishway | RKM 83 |  |  |
| RIA | Rock Island | Dam Ad | alt Fishw | ways RK | M 730 |  |  | LMR | Lower M | ethow R | ver at P | teros RK | M 843 |  |

### 15.1 Genetics

Genetic tissue was collected from each Chinook salmon spawned at PRH during 2017 by staff from the Columbia River Inter-Tribal Fish Commission (CRITFC). In total 5,122 specimens were collected at PRH to support their work associated with genetic stock identification and parentage-based tagging. The tissue samples collected from return years 2011 through 2017 is currently being archived by CRITFC. During 2010, WDFW staff collected 100 genetic tissue samples from both the Priest Rapids Hatchery broodstock and naturally spawning broodstock from the Hanford Reach.

### 15.2 Proportion of Natural Influence

The intent of integrated hatchery programs is to achieve management objectives while having hatchery and natural origin fish share a common gene pool. Gene flow and the associated risks within and between the hatchery and natural environments can be estimated using a simple ratio estimator using the proportion of natural origin fish in the hatchery broodstock (pNOB) and the proportion of hatchery origin fish in the natural spawning escapement (pHOS). This ratio of pNOB/(pHOS +pNOB ) is termed the Proportionate Natural Influence (PNI). The larger the PNI ratio, the greater selection that the natural environment has on the population relative to that of the hatchery environment. Alternatively, PNI estimates addressing gene flow from multiple
sources/hatchery programs can be calculated from a multiple population gene flow model based on the Ford model which has been extended to three or more populations (Busack 2015, 2016).
In order for the natural environment to dominate selection, PNI for either calculation should be greater than 0.5 and for integrated hatchery programs the Hatchery Scientific Review Group (HSRG) recommends a $\mathrm{PNI} \geq 0.67$ (HSRG/WDFW/NWIFC 2004). The HSRG recommends a minimum target of 0.15 for the proportion of natural origin Chinook salmon to be incorporated into the hatchery broodstock (pNOB) as well as a maximum target of 0.30 for the proportion of hatchery origin Chinook allowed to spawn in the natural environment (pHOS) for the Hanford Reach if it is to be managed as an integrated hatchery program.

Several estimates of PNI have been calculated to show the contributions of multiple programs on the overall PNI for the Hanford Reach. These programs include the hatchery production associated with the Grant PUD and USACE mitigation and the influence of strays. The different PNI estimates are based on pNOB and pHOS estimates specific to each source of spawning adults. The methods used to allocate pNOB and pHOS are described in the following sections.

### 15.3 Estimate of pNOB

Estimates of pNOB based on otolith samples are limited to return years 2012 through 2017. Otolith marking began with the 2007 brood. Therefore, otolith marks are only available for specific age classes of PRH origin fish during return years 2010 and 2011 and do not provide representative samples for estimating pNOB for the PRH broodstock.
The annual pNOB for fish spawned at PRH and used for Grant PUD and USACE smolt releases into the Hanford Reach during return years 2012 through 2017 is provided in Table 51.
Table 51 Origin of broodstock and pNOB apportioned to program for fall Chinook salmon spawned at Priest Rapids Hatchery, Return Years 20102-2017.

| Return Year | $\mathbf{N}$ | GCPUD pNOB | USACE pNOB | GCPUD and <br> USACE Combined <br> pNOB | Other Programs <br> pNOB $^{\mathbf{1}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2012 | 4,974 | 0.182 | 0.057 | 0.119 | N/A |
| 2013 | 5,442 | 0.225 | 0.026 | 0.127 | N/A |
| 2014 | 5,443 | 0.343 | 0.076 | 0.206 | 0.000 |
| 2015 | 5,524 | 0.313 | 0.045 | 0.179 | 0.000 |
| 2016 | 4,938 | 0.259 | 0.073 | 0.163 | 0.000 |
| 2017 | 5,668 | 0.433 | 0.091 | 0.254 | 0.000 |
| Mean | 5,332 | 0.293 | 0.061 | 0.175 | 0.000 |

${ }^{1}$ Represents pNOB associated with egg-takes utilized outside of the Hanford Reach.
The 2017 broodstock included 5,668 adults which were comprised of 4,511 fish from the volunteer trap, 809 from the OLAFT and 348 from the ABC. In general, broodstock from ABC and OLAFT are held in a specific holding pond (Pond 4) and mated with fish from this pond or with fish collected from the PRH volunteer trap and held in another specific holding pond (Pond 1). The fish culturists segregate the progeny resulting from these matings for release from PRH. Brood stock utilized for non-Grant PUD programs are collected from the PRH volunteer trap and held in a specific pond (Pond 2). Large portions of the progeny from the Pond 2 broodstock are shipped to other facilities for use by other programs.

Grant PUD funds the collection of non-marked or tagged broodstock from the ABC and OLAFT with the intent of improving the pNOB associated with the production of their 5.6 million smolt
mitigation requirement. The inclusion of these fish contributed greatly to the Grant PUD program's egg-take goal and the resulting pNOB (Table 52).
Table 52 Origin of broodstock and pNOB apportioned to program for fall Chinook salmon spawned at Priest Rapids Hatchery, Brood Year 2017.

| Program | Egg-Take | Facility Mean Fecundity | Natural Females | Hatchery Females | Natural <br> Males | Hatchery Males | Total Natural | Total <br> Hatchery | pNOB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GCPUD | 5,317,004 | 3,739 | 572 | 850 | 280 | 444 | 852 | 1,294 | 0.397 |
| GCPUD Alt Mating ${ }^{1}$ | 1,621,391 | 3,762 | 208 | 223 | 107 | 0 | 315 | 223 | 0.585 |
| GCPUD Combined | 6,938,395 | 3,744 | 780 | 1,073 | 387 | 444 | 1,167 | 1,517 | 0.435 |
| USACE - PRH | 2,011,997 | 3,517 | 50 | 522 | 40 | 251 | 89 | 774 | 0.103 |
| USACE - RSH | 4,788,524 | 3,430 | 88 | 1,308 | 93 | 613 | 181 | 1,921 | 0.086 |
| USACE Combined | 6,800,521 | 3,456 | 137 | 1,831 | 133 | 864 | 270 | 2,695 | 0.091 |
| Combined PRH and RSH Programs | 13,738,916 | 3,596 | 917 | 2,904 | 520 | 1,308 | 1,437 | 4,212 | 0.254 |
| Other Programs ${ }^{2}$ | 42,196 | 3,516 | 1 | 11 | 1 | 5 | 2 | 16 | 0.111 |

${ }^{1}$ Alternative mating strategy incorporates 1 natural origin male x 4 females.
${ }^{2}$ Includes eggs from presumed hatchery $x$ hatchery crosses shipped to educational organizations.
An alternative pNOB for calculating PNI was developed to account for the genetic influence on pNOB resulting from the PRH spawning protocol of spawning one male with one, two, or four females. It is intended to represent actual gene flow to the progeny instead of strictly the origin and number of parents. This information is presented in Appendix J for comparison to other conventional pNOB calculations.

### 15.4 Estimates of pHOS

Estimates of pHOS based on otolith samples are limited to return years 2012 through 2017. Otolith marking began with the 2007 brood. Hence, otolith marks are only available for specific age classes of PRH origin fish during return years 2010 and 2011 and do not provide representative samples for estimating population level pHOS. The population level pHOS estimates for recent annual Hanford Reach spawning escapements are presented Table 53.
Table 53 Proportion of hatchery Chinook salmon on the spawning grounds (pHOS) in the Hanford Reach, Brood Years 2012-2017.

| Return <br> Year | N | Total Escapement | PRH | RSH | Other $^{\mathbf{1}}$ | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1,609 | 57,631 | 0.062 | 0.066 | 0.005 | 0.135 |
| 2013 | 927 | 126,744 | 0.203 | 0.054 | 0.018 | 0.275 |
| 2014 | 2,426 | 183,750 | 0.052 | 0.015 | 0.028 | 0.096 |
| 2015 | 2,485 | 266,347 | 0.076 | 0.017 | 0.004 | 0.097 |
| 2016 | 1,648 | 116,421 | 0.066 | 0.022 | 0.027 | 0.115 |
| 2017 | 1,813 | 73,759 | 0.063 | 0.017 | 0.001 | 0.081 |
| Mean | 1,818 | 137,442 | 0.087 | 0.032 | 0.014 | 0.134 |

[^6]Estimates for pHOS were calculated for contributing sources of hatchery origin fall Chinook escapement in the Hanford Reach (Table 54). The pHOS associated with the PRH origin escapement was apportioned between the Grant PUD and USACE programs at PRH based on the annual mitigation requirement for the number of juveniles released by each program for brood
years 2008 through 2012. The pHOS estimate for return year 2017 includes 4,642 PRH origin fish in the escapement. Of these, $74.6 \%$ and $25.4 \%$ were allocated respectively to Grant PUD (3,463 fish) and USACE (1,179 fish) programs at PRH. The USACE's $25.4 \%$ portion of the PRH origin escapement was combined with the escapement associated with the USACE's RSH program (1,230 fish) to estimate the pHOS associated with the USACE programs in the Hanford Reach. There were 79 hatchery fish in the escapement associated with other hatchery programs located outside of the Hanford Reach.
Table 54 Origin of pHOS apportioned by program source for fall Chinook salmon spawning naturally in the Hanford Reach, Return Years 2012-2017.

| Return Year | Natural Origin | Hatchery Origin Spawners |  |  |  | pHOS by Source |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | GCPUD ${ }^{1}$ | USACE ${ }^{1,2}$ | Other ${ }^{3}$ | Total | GCPUD $^{1}$ | USACE ${ }^{1,2}$ | Other ${ }^{3}$ | Combined |
| 2012 | 50,072 | 3,943 | 3,598 | 261 | 7,803 | 0.068 | 0.062 | 0.005 | 0.135 |
| 2013 | 126,782 | 26,507 | 18,427 | 3,123 | 48,057 | 0.152 | 0.105 | 0.018 | 0.275 |
| 2014 | 166,183 | 7,185 | 5,262 | 5,120 | 17,567 | 0.039 | 0.029 | 0.028 | 0.096 |
| 2015 | 240,511 | 15,101 | 9,669 | 1,065 | 25,835 | 0.057 | 0.036 | 0.004 | 0.097 |
| 2016 | 103,033 | 5,732 | 4,513 | 3,143 | 13,388 | 0.049 | 0.039 | 0.027 | 0.115 |
| 2017 | 67,807 | 3,463 | 2,409 | 79 | 5,951 | 0.047 | 0.033 | 0.001 | 0.081 |

${ }^{1}$ Estimated number of PRH origin fish that spawned naturally in the Hanford Reach. Of these, $74.6 \%$ and $25.4 \%$ were apportioned to Grant PUD-PRH and USACE-PRH, respectively. The allocation of pHOS was based on the proportion of annual juvenile mitigation goals for each agency for brood years 2008 through 2012.
${ }^{2}$ Includes hatchery origin fish released from Ringold Springs Hatchery.
${ }^{3}$ Includes hatchery origin fish released from other hatcheries based on the presence of a hatchery mark without an otolith mark.

### 15.5 Estimates of PNI

We present a hierarchy of PNI estimates based on pNOB and pHOS values calculated to reflect differing methodologies driven by the type of data available to assign origin of adult Chinook salmon returns. The population level PNI for the Hanford Reach includes all hatchery origin fish regardless of hatchery program or funding source.
Prior to return year 2012, pHOS, pNOB and PNI rates were based on CWT recoveries from the adult returns. Historically, we used juvenile mark rate expansions of CWT recoveries in the hatchery and stream surveys for these calculations. The pNOB estimated from CWT requires the assumption that fish unaccounted for by the juvenile mark rate expansions are natural origin fish. As discussed in Appendix A of this report, this assumption significantly over estimates pNOB and PNI. This method of estimated pNOB for the 2015 through 2017 broodstock was not calculated due to culling fish possessing a CWT and or an adipose clip. Hence, the broodstock origin is poorly represented by CWT.
The pHOS estimates based on juvenile mark rate expansions of CWT recoveries also likely underestimate the presences of PRH and RSH origin fish as explained in Appendix A. For comparison, we present CWT based estimates of PNI derived from CWT adult-to-adult expansions for PRH and RSH origin adult recoveries at their respective hatcheries. An explanation of methods is given in Appendix M. Estimates of pNOB, pHOS, and PNI based on both methods of CWT expansions are presented in Table 55.
The pHOS and pNOB estimates from limited otolith datasets for recent complete brood years are more similar to the estimates produced by adult-to-adult CWT expansions versus juvenile mark rate expansions of CWT recoveries of returning adults.

Table 55 PNI of the Hanford Reach fall Chinook salmon supplementation program based on expanded coded-wire tag recoveries of all fish surveyed, Return Years 2001-2017.

| Return Year | pNOB ${ }^{1}$ | pHOS ${ }^{1}$ | pNOB ${ }^{2}$ | pHOS ${ }^{2}$ | PNI based on $\mathrm{pNOB}^{1}$ and $\mathrm{pHOS}^{1}$ | PNI based on $\mathrm{pNOB}^{2}$ and pHOS ${ }^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0.155 | 0.094 | 0.046 | 0.066 | 0.622 | 0.411 |
| 2002 | 0.145 | 0.101 | 0.046 | 0.125 | 0.589 | 0.269 |
| 2003 | 0.132 | 0.099 | 0.046 | 0.117 | 0.571 | 0.282 |
| 2004 | 0.229 | 0.081 | 0.046 | 0.099 | 0.739 | 0.317 |
| 2005 | 0.370 | 0.106 | 0.046 | 0.156 | 0.777 | 0.229 |
| 2006 | 0.507 | 0.057 | 0.046 | 0.124 | 0.899 | 0.271 |
| 2007 | 0.326 | 0.041 | 0.046 | 0.065 | 0.888 | 0.414 |
| 2008 | 0.501 | 0.046 | 0.046 | 0.087 | 0.916 | 0.346 |
| 2009 | 0.568 | 0.077 | 0.046 | 0.174 | 0.881 | 0.209 |
| 2010 | 0.392 | 0.040 | 0.046 | 0.076 | 0.907 | 0.377 |
| 2011 | 0.381 | 0.076 | 0.046 | 0.154 | 0.836 | 0.230 |
| 2012 | 0.304 | 0.074 | $0.119^{\text {a }}$ | 0.118 | 0.871 | 0.529 |
| 2013 | 0.252 | 0.217 | $0.127^{\text {a }}$ | 0.287 | 0.537 | 0.300 |
| 2014 | 0.443 | 0.054 | $0.206^{\text {a }}$ | 0.069 | 0.888 | 0.760 |
| 2015 | $\mathrm{N} / \mathrm{A}^{3}$ | 0.072 | $0.179^{\text {a }}$ | 0.075 | N/A ${ }^{3}$ | 0.691 |
| 2016 | N/A ${ }^{3}$ | 0.092 | $0.163^{\text {a }}$ | 0.097 | N/A ${ }^{3}$ | 0.627 |
| 2017 | $\mathrm{N} / \mathrm{A}^{3}$ | 0.116 | $0.254^{\text {a }}$ | 0.102 | $\mathrm{N} / \mathrm{A}^{3}$ | 0.713 |

pNOB $^{1}$ Assumes that all fish not accounted for by juvenile coded-wire tag expansions are natural origin. $\mathrm{pHOS}^{1}$ based on hatchery origin coded-wire recoveries expanded by juvenile mark rate and survey sample rate. $\mathrm{pNOB}^{2}$ is assigned to years 2001 - 2011 based on an average proportion of natural origin returns to PRH for return years 2012-2014 as determined by otolith and other hatchery marks.
pHOS ${ }^{2}$ is based on an adult coded-wire tag expansion rate for PRH and RSH origin adults recovered in the Hanford Reach escapement combined with juveniles coded-wire tag mark rate expansions for other hatchery strays. Both groups were expanded by the survey sample rate.
${ }^{3}$ Brood stock was generally high-graded to remove coded-wire tagged fish during ponding.
${ }^{2} \mathrm{pNOB}$ of broodstock used for production of PRH and RSH programs as determined from otoliths and other hatchery marks.

For return years 2012-2017 we present PNI estimates calculated from the multiple population gene flow model (Busack 2015) and otoliths (Table 56). The output from this model indicates that the PNI values for return years 2014 and 2017 have exceeded the goal of 0.670.

Table 56 PNI estimates for the Hanford Reach fall Chinook salmon supplementation programs based on otoliths, Return Years 2012-2017. Calculated from multiple population gene flow model based on the Ford model which has been extended to three or more populations.

| Return Year | pNOB |  |  | pHOS |  |  | pHOS <br> Reach ${ }^{7}$ | $\begin{gathered} \text { PNI } \\ \text { Population }^{8} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | GCPUD ${ }^{1}$ | USACE ${ }^{2}$ | Facility ${ }^{3}$ | GCPUD ${ }^{4}$ | USACE ${ }^{5}$ | Other ${ }^{6}$ |  |  |
| 2012 | 0.182 | 0.057 | 0.119 | 0.068 | 0.062 | 0.005 | 0.135 | 0.599 |
| 2013 | 0.225 | 0.027 | 0.127 | 0.152 | 0.105 | 0.018 | 0.275 | 0.463 |
| 2014 | 0.343 | 0.076 | 0.206 | 0.039 | 0.029 | 0.028 | 0.096 | 0.775 |
| 2015 | 0.313 | 0.045 | 0.179 | 0.057 | 0.036 | 0.004 | 0.097 | 0.762 |
| 2016 | 0.259 | 0.072 | 0.163 | 0.049 | 0.039 | 0.027 | 0.115 | 0.700 |
| 2017 | 0.433 | 0.091 | 0.254 | 0.047 | 0.033 | 0.001 | 0.081 | 0.835 |
| Mean | 0.293 | 0.061 | 0.175 | 0.069 | 0.051 | 0.014 | 0.133 | 0.689 |

${ }^{1}$ Includes broodstock associated with Grant PUD production at PRH.
${ }^{2}$ Includes broodstock associated with USACE production at PRH and RSH.
${ }^{3}$ Includes broodstock spawned at PRH for all production
${ }^{4}$ Includes pHOS associated with Grant PUD mitigation smolt releases at PRH
${ }^{5}$ Includes pHOS associated with USACE mitigation smolt releases at PRH and RSH
${ }^{6}$ Includes pHOS associated with strays from hatcheries outside of the Hanford Reach
${ }^{7}$ Population level pHOS in the Hanford Reach
${ }^{8}$ Population level PNI for the Hanford Reach. Assumes strays from hatcheries outside of the Hanford Reach have an associated pNOB of zero.

### 16.0 Natural and Hatchery Replacement Rates

The numbers of hatchery origin recruits (HOR) are estimated from CWT recoveries for brood year returns to the PRH and the Hanford Reach of the Columbia River. The recovered CWTs are expanded by sample rate of the survey and then by the juvenile tag rate. CWTs recovered from natural origin recruits (NOR) originating from the Hanford Reach are difficult to expand accurately because the juvenile tag rates are unknown. Therefore, an assumption was made that returns not accounted for by HOR CWT recoveries are NOR. Recent data indicates that that CWT data likely underestimates the true number of HORs and as a result, our assumption likely overestimates the number of NOR.

Hatchery replacement rates (HRR) were calculated as the ratio of HOR to the parent broodstock at PRH. This broodstock is an estimate of the number of fish spawned at PRH to produce the target release of subyearling fall Chinook salmon. Similarly, natural replacement rates (NRR) for the Hanford Reach URB fall Chinook salmon were calculated as the ratio of NOR to the parent population spawning naturally in the Hanford Reach natural environment. This spawning population is based on the escapement estimate to the Hanford Reach without adjustments for spawn success.
Harvest estimates for HOR were calculated from the proportion of the expanded CWT recoveries in the fisheries to the total number of the expanded CWTs recoveries included in fisheries, stream surveys, and hatchery racks. The CWT recoveries are expanded by sample rate of the survey and juvenile mark rate for the CWT group. Since there is not a CWT mark rate for NOR, the harvest rates for PRH origin returns (HOR) were used as an indicator for similar brood years of NOR.

The HRR and NRR for brood year 2011, includes harvest, was 32.03 and 4.93, respectively (Table 57). In comparison, the HRR and NRR for brood year without harvest was 17.46 and
4.93, respectively. The HRR should be greater than or equal to 5.30 (the target value in Murdoch and Peven 2005).
Table 57 Broodstock spawned at Priest Rapids Hatchery, estimated escapement to the Hanford Reach, natural and hatchery origin recruits (NOR and HOR), and natural and hatchery replacement rates (NRR and HRR, with and without harvest) for natural origin fall Chinook salmon in the Hanford Reach, Brood Years 1996-2011.

| Brood Year | Broodstock Spawned | $\begin{array}{\|c} \text { Hanford } \\ \text { Reach } \\ \text { Escapement }{ }^{1} \end{array}$ | Harvest not included |  |  |  | Harvest included ${ }^{2}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | HOR | NOR | HRR | NRR | HOR | NOR | HRR | NRR |
| 1996 | 2,859 | 43,249 | 13,584 | 28,849 | 4.75 | 0.67 | 26,205 | 59,899 | 9.17 | 1.38 |
| 1997 | 2,726 | 43,493 | 3,002 | 44,416 | 1.10 | 1.02 | 6,037 | 88,349 | 2.21 | 2.03 |
| 1998 | 3,027 | 35,393 | 18,464 | 93,999 | 6.10 | 2.66 | 31,932 | 222,865 | 10.55 | 6.30 |
| 1999 | 2,619 | 29,812 | 27,093 | 115,237 | 10.34 | 3.87 | 52,099 | 240,090 | 19.89 | 8.05 |
| 2000 | 2,619 | 48,020 | 4,665 | 56,422 | 1.78 | 1.17 | 12,508 | 89,983 | 4.78 | 1.87 |
| 2001 | 3,621 | 59,848 | 25,059 | 71,359 | 6.92 | 1.19 | 55,789 | 129,548 | 15.41 | 2.16 |
| 2002 | 3,630 | 84,509 | 5,277 | 47,813 | 1.45 | 0.57 | 12,744 | 81,600 | 3.51 | 0.97 |
| 2003 | 3,003 | 100,508 | 3,021 | 31,788 | 1.01 | 0.32 | 5,974 | 64,307 | 1.99 | 0.64 |
| 2004 | 3,014 | 87,696 | 1,109 | 22,747 | 0.37 | 0.26 | 3,262 | 34,465 | 1.08 | 0.39 |
| 2005 | 2,898 | 71,967 | 21,107 | 64,011 | 7.28 | 0.89 | 61,122 | 97,777 | 21.09 | 1.36 |
| 2006 | 2,911 | 51,701 | 998 | 54,288 | 0.34 | 1.05 | 3,347 | 77,344 | 1.15 | 1.50 |
| 2007 | 2,096 | 22,274 | 22,184 | 101,753 | 10.58 | 4.57 | 52,832 | 175,404 | 25.21 | 7.87 |
| 2008 | 2,959 | 29,058 | 11,867 | 41,809 | 4.01 | 1.44 | 25,166 | 79,116 | 8.51 | 2.72 |
| 2009 | 3,177 | 36,720 | 28,154 | 97,834 | 8.86 | 2.66 | 85,489 | 145,874 | 26.91 | 3.97 |
| 2010 | 3,320 | 87,016 | 97,567 | 281,364 | 29.38 | 3.23 | 209,338 | 526,972 | 63.05 | 6.06 |
| 2011 | 2,830 | 75,256 | 49,396 | 168,864 | 17.46 | 2.24 | 41,232 | 371,161 | 32.03 | 4.93 |
| Mean | 2,957 | 56,658 | 20,784 | 82,660 | 6.98 | 1.74 | 42,817 | 155,297 | 15.41 | 3.26 |
| Median | 2,935 | 49,861 | 16,024 | 60,217 | 5.43 | 1.18 | 29,068 | 93,880 | 9.86 | 2.10 |

${ }^{1}$ Includes estimated adult and jack escapement to the Hanford Reach natural environment.
${ }^{2}$ Harvest rates for NORs was estimated using the HRRs harvest rates for similar brood years as an indicator stock.

### 17.0 Smolt-to-Adult Survivals

Smolt-to-adult survival ratios (SAR) were calculated by dividing the expanded number of adults possessing a CWT recovered by the number of CWT smolts released. This estimate could be biased low for both hatchery and natural origin fish because of some of CWT bias identified previously in this report. The following data was obtained from the RMPC's RMIS online database: http://www.rmpc.org/. The 2011 brood year data was queried on March 9, 2018. This query should account for age 2 through 5 fall Chinook salmon sampled through December 2016. The lag in reporting field data for the 2017 return year likely excludes recoveries of a limited number of age-6 fish from the 2011 brood.
Annual SAR for hatchery fall Chinook salmon released from PRH for brood years 1992 through 2011 have a mean of 0.0066 with a median of 0.0045 (Table 58). The SAR for the PRH origin 2010 brood is 0.0304 , which is the highest SAR on record for PRH releases.

Table 58 Smolt-to-adult Survial Ratios (SAR) for Priest Rapids Hatchery fall Chinook salmon, Brood Years 1992-2011. Data includes all coded-wire tag recoveries from adipose clipped fish.

| Brood Year | Number of Tagged <br> Smolts Released | Estimated Adult <br> Captures | SAR |
| :---: | :---: | :---: | :---: |
| 1992 | 194,622 | 448 | 0.0023 |
| 1993 | 185,683 | 1,479 | 0.0080 |
| 1994 | 175,880 | 108 | 0.0006 |
| 1995 | 196,189 | 1,786 | 0.0091 |
| 1996 | 193,215 | 762 | 0.0040 |
| 1997 | 196,249 | 183 | 0.0009 |
| 1998 | 193,660 | 946 | 0.0049 |
| 1999 | 204,346 | 1,573 | 0.0077 |
| 2000 | 200,779 | 370 | 0.0018 |
| 2001 | 219,926 | 1,810 | 0.0082 |
| 2002 | 355,373 | 669 | 0.0019 |
| 2003 | 399,116 | 352 | 0.0009 |
| 2004 | 200,072 | 100 | 0.0005 |
| 2005 | 199,445 | 1,718 | 0.0086 |
| 2006 | 202,000 | 100 | 0.0005 |
| 2007 | 202,568 | 2,391 | 0.0118 |
| 2008 | 218,082 | 740 | 0.0034 |
| 2009 | 619,568 | 7,820 | 0.0126 |
| 2010 | 605,000 | 18,620 | 0.0308 |
| 2011 | 595,608 | 7,643 | 0.0128 |
| Mean | $\mathbf{2 7 7 , 8 6 9}$ | $\mathbf{2 , 4 8 1}$ | $\mathbf{0 . 0 0 6 6}$ |
| Median | $\mathbf{2 0 1 , 3 9 0}$ | $\mathbf{8 5 4}$ | $\mathbf{0 . 0 0 4 5}$ |

Annual SAR for Hanford Reach natural origin fall Chinook salmon for brood years 1992 through 2011 had a mean of 0.0043 with a median of 0.0024 (Table 59). The SAR for the Hanford Reach natural origin 2010 brood is 0.0164 which is the highest SAR on record for the Hanford Reach natural origin stock.

Table 59 Smolt-to-adult Survival Ratios (SAR) for Hanford Reach natural origin fall Chinook salmon, Brood Years 1992-2011. Data includes all coded-wire tag recoveries from adipose clipped fish.

| Brood Year | Number of Tagged <br> Smolts Released | Estimated Adult <br> Captures | SAR |
| :---: | :---: | :---: | :---: |
| 1992 | 203,591 | 829 | 0.0041 |
| 1993 | 95,897 | 485 | 0.0051 |
| 1994 | 148,585 | 74 | 0.0005 |
| 1995 | 146,887 | 340 | 0.0023 |
| 1996 | 92,262 | 111 | 0.0012 |
| 1997 | 199,896 | 365 | 0.0018 |
| 1998 | 129,850 | 784 | 0.0060 |
| 1999 | 213,259 | 2,378 | 0.0112 |
| 2000 | 204,925 | 362 | 0.0018 |
| 2001 | 127,758 | 519 | 0.0041 |
| 2002 | 203,557 | 338 | 0.0017 |
| 2003 | 207,168 | 199 | 0.0010 |
| 2004 | 163,884 | 147 | 0.0009 |
| 2005 | 203,929 | 301 | 0.0015 |
| 2006 | 263,478 | 356 | 0.0007 |
| 2007 | 53,618 | 456 | 0.0085 |
| 2008 | 203,947 | 520 | 0.0025 |
| 2009 | 201,606 | 1,597 | 0.0079 |
| 2010 | 179,727 | 2,956 | 0.0164 |
| 2011 | 166,610 | 1,063 | 0.0064 |
| Mean | $\mathbf{1 7 0 , 5 2 2}$ | $\mathbf{7 0 9}$ | $\mathbf{0 . 0 0 4 3}$ |
| Median | $\mathbf{1 8 9 , 8 1 2}$ | $\mathbf{4 1 1}$ | $\mathbf{0 . 0 0 2 4}$ |

### 18.0 ESA/HCP Compliance

### 18.1 Broodstock Collection

Section 10(a)(1)(B) Permit 1347 authorizes collection of fall Chinook broodstock at the OLAFT for the Priest Rapids hatchery program with an incidental take limit of 10 steelhead (an aggregate of hatchery or wild). Due to the absence of an identified steelhead take limit for operation of the PRH volunteer trap in permit 1347 and through ongoing coordination with NOAA Fisheries, the 10 fish take limit for broodstock collection at the OLAFT, on an interim basis (until a new permit is issued), has been re-conceptualized to include broodstock collection at the PRH volunteer trap, and in the ABC fishery. During the 2017 fall Chinook broodstock collection activities, a total of 5 steelhead, one of which was not adipose clipped, were encountered at the PRH volunteer trap with no incidental mortality reported. No steelhead mortalities were associated with broodstock collection at the OLAFT or in the ABC fishery (Table 60.

Table 60 Recoveries and disposition of steelhead at the Priest Rapids Hatchery volunteer trap, Return Year 2017.

|  |  | No Mark | Ad Only | Ad-RV | Total |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Released | Males | 1 | 1 | 1 | 3 |
|  | Females | 0 | 0 | 2 | 2 |
|  | Sub Total | 1 | 1 | 3 | 5 |
| Killed | Males | 0 | 0 | 0 | 0 |
|  | Females | 0 | 0 | 0 | 0 |
|  | Sub Total | 0 | 0 | 0 | 0 |
| Total |  | 1 | 1 | 3 | 5 |

### 18.2 Hatchery Rearing and Release

The juvenile fall Chinook salmon from the 2016 brood year reared throughout their life-stages at PRH without incident. The 2017 smolt release totaled 7,987,222 URB fall Chinook salmon, representing $109 \%$ of the production objective and was compliant with the $10 \%$ overage allowable in ESA Section 10 Permit 1347.

### 18.3 Distribution of Surpluses, Mortalities, and Spawned, Adult fall Chinook Salmon from Priest Rapids Hatchery

All adult Chinook salmon recovered at PRH are eventually distributed to multiple organizations depending on the condition and treatment of the individual fish while at the hatchery. A large majority of these fish are suitable for consumption and transported to Foodbanks (Table 61). In 2017, roughly $84 \%$ of the surplus fish were used for human consumption.
Table 61 Disposition of Chinook salmon removed from Priest Rapids Hatchery volunteer trap, Return Year 2001-2017.

| Return Year | Disposal of <br> Pet Food | ortalities <br> Landfill | WDFW <br> Nutrient Enhancement Projects | Donations to <br> Educational <br>  <br> Research | Donations to Foodbanks | Donations to Tribes | Sold to Fish Buyers | Fish Removed from Priest Rapids Hatchery |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 6,597 | 2,054 | 0 |  | 525 | 6,139 | 15,315 |
| 2002 | 0 | 6,572 | 2,192 | 0 | 3,130 | 502 | 0 | 12,396 |
| 2003 | 0 | 5,144 | 3,211 | 9 | 881 | 98 | 0 | 9,343 |
| 2004 | 350 | 2,661 | 2,756 | 88 | 9,371 |  | 595 | 15,821 |
| 2005 | 153 | 5,635 | 318 | 2 | 0 |  | 4,503 | 10,611 |
| 2006 | 0 | 5,467 | 0 | 250 | 0 | 340 | 2,146 | 8,203 |
| 2007 | 2,595 | 0 | 0 | 0 | 0 | 159 | 3,345 | 6,099 |
| 2008 | 5,384 | 90 | 0 | 340 | 0 | 375 | 13,428 | 19,617 |
| 2009 | 5,846 | 0 | 0 | 310 | 0 | 201 | 6,502 | 12,859 |
| 2010 | 5,412 | 1,937 | 1,937 | 452 | 3,548 | 8 | 8,259 | 21,553 |
| 2011 | 6,951 | 0 | 1,500 | 412 | 11,217 | 588 | 0 | 20,668 |
| 2012 | 7,554 | 0 | 0 | 460 | 20,628 |  | 0 | 28,642 |
| 2013 | 10,108 | 0 | 0 | 489 | 31,647 | 626 | 0 | 42,870 |
| 2014 | 10,805 | 0 | 0 | 237 | 67,684 | 783 | 0 | 79,509 |
| 2015 | 7,402 | 0 | 0 | 398 | 52,987 | 4,228 | 0 | 65,015 |
| 2016 | 7,833 | 0 | 0 | 411 | 19,424 | 1,948 | 0 | 29,616 |
| 2017 | 10,108 ${ }^{\text {a }}$ | 0 | 0 | 436 | 6,413 | 1,505 | 0 | 19,259 |
| Mean | 4,735 | 2,006 | 822 | 253 | 14,183 | 849 | 2,642 | 24,553 |
| Median | 5,412 | 0 | 0 | 310 | 4,981 | 514 | 0 | 19,259 |

[^7]
### 18.4 Hatchery Effluent Monitoring

Per ESA Permits 1196, 1347, and 1395, permit holders shall monitor and report hatchery effluents in compliance with applicable National Pollution Discharge Elimination Systems (NPDES) (EPA 1999) permit limitations. There were no NPDES violations reported at Grant PUD Hatchery facilities during the September 2017 through June 2018 collection and rearing periods.

### 18.5 Ecological Risk Assessment

One of the regional objectives in the Grant PUD M\&E plan is to conduct an ecological risk assessment on non-target taxa of concern to determine if additional M\&E is necessary (Pearsons and Langshaw 2009). The methodology that was used to assess risks was presented in Pearsons et al. (2012) and Pearsons and Busack (2012). This objective was completed through an approved report that summarized the methods and results of the risk assessment (Mackey et al. 2014).

### 19.0 Acknowledgments

We thank the many individuals and organizations that helped collect the data or provided data for inclusion in this report: Shawnaly Meehan and Dennis Werlau for the leadership of their WDFW M\&E crews that sorted and sampled over 26,000 fall Chinook salmon, along with entering and reviewing all the sample data collected in 2017; Paul Hoffarth for supervising the creel and carcass staff and summarizing the sport harvest data; and Alf Haukenes for providing helpful suggestions for the improvement of this report. Debbie Firestone (Grant PUD) assisted with final formatting of the report. Furthermore, we would like to thank the hatchery staff at Priest Rapids and Ringold Springs Hatcheries: Brian Lyon, Glen Pearson, Mike Erickson and their crews for accommodating the activities associated with M\&E. This work was primarily funded by GCPUD, the USACE, WDFW, and the CRCWTRP.

## Literature Cited

AFS-FHS (American Fisheries Society-Fish Health). 2014. FHS blue book: suggested procedures for the detection and identification of certain finfish and shellfish pathogens, 2014 edition.

Busack, C. 2015. Extending the Ford model to three or more populations. August 31, 2015. Sustainable Fisheries Division, West Coast Region, National Marine Fisheries Service. 5p.
Busack, C. 2016. Methow Gene Flow 2.xlsx. NMFS West Coast Region Portland, Oregon.
Elliot, D. G. 2012. Bacterial kidney disease. In AFS-FHS (American Fisheries Society- Fish Health Section. FHS blue book: suggested procedures for the detection and identification of certain finfish and shellfish pathogens, 2014 edition.
Fryer, J. K. 2018, Expansion of the 2017 Hanford Reach fall Chinook Salmon juvenile codedwire tagging and PIT tagging project. Report submitted to the Pacific Salmon Commission Technical Committee from the Columbia River Inter-Tribal Fish Commission, Portland, Oregon.
Ham, K. D., and T. N. Pearsons. 2001. A practical approach for containing ecological risks associated with fish stocking programs. Fisheries 25(4):15-23.
Harnish, R. A., R. Sharma, G. A. McMichael, R. B. Langshaw, and T. N. Pearsons. 2014.
Effect of hydroelectric dam operations on the freshwater productivity of a Columbia River fall Chinook salmon population. Canadian Journal of Fisheries and Aquatic Sciences 71:602615.

Harnish, R.A., R. Sharma, G. A. McMichael, R.B. Langshaw, T.N. Pearsons, and D.A. Bernard. 2012. Effects of Priest Rapids Dam Operations on Hanford Reach Fall Chinook Salmon Productivity and Estimation of Maximum Sustainable Yield, 1975-2004. Prepared for: Public Utility District No. 2 of Grant County, Ephrata, WA. Contract Number 430-2464.
Harnish, R. A. 2017. Hanford Reach Upriver Bright Productivity Analysis Update. Pacific Northwest National Laboratory, Richland, Washington.

Hillman, T., T. Kahler, G. Mackey, A. Murdoch, K. Murdoch, T. Pearsons, M. Tonseth and C. Willard. 2017. Monitoring and evaluation plan for PUD hatchery programs: 2017 update. Report to the HCP and PRCC Hatchery Committees, Wenatchee and Ephrata WA.

Hoffarth, P. A. and T. N. Pearsons. 2012a. Priest Rapids Hatchery Monitoring and Evaluation: Annual Report for 2010. Grant County Public Utility District, Ephrata, Washington.
Hoffarth, P. A. and T. N. Pearsons. 2012b. Priest Rapids Hatchery Monitoring and Evaluation: Annual Report for 2011. Grant County Public Utility District, Ephrata, Washington.

Langshaw, R. B., P. J. Graf and T. N. Pearsons. 2015. Effects of the Hanford Reach Fall Chinook Protection Program on Fall Chinook Salmon in the Hanford Reach -Summary, Conclusions, and Future Monitoring. Grant County Public Utility District, Ephrata, Washington.

Langshaw, R. B., P. J. Graf and T. N. Pearsons. 2017. Hydropower and high
productivity in the Hanford Reach: A synthesis of how flow management may benefit fall Chinook Salmon in the Columbia River, USA. WIREs Water. 2017;e1275. https://doi.org/10.1002/wat2.1275

Mackey, G., T. N. Pearsons, M. R. Cooper, K. G. Murdoch, A. R. Murdoch, and T. W. Hillman. 2014. Ecological risk assessment of upper Columbia hatchery programs on non-target taxa of concern. Report produced by the Hatchery Evaluation Technical Team (HETT) for the HCP Wells Hatchery Committee, HCP Rocky Reach Hatchery Committee, HCP Rock Island Hatchery Committee, and the Priest Rapids Hatchery Sub-Committee.

Murdoch, A.R , and C. Peven. 2005. Conceptual approach to monitoring and evaluating the Chelan County Public Utility District Hatchery Programs. Final report to the Chelan PUD Habitat Conservation Plan's Hatchery Committee.
Murdoch, A. R., T. N. Pearsons, and T. W. Maitland. 2010. Estimating the spawning escapement of hatchery- and natural-origin spring Chinook Salmon using redd and carcass data. North American Journal of Fisheries Management 30:361-375.

Norris, J.G., S.Y. Hyun, and J.J. Anderson, 2000. Ocean distribution of Columbia River upriver bright fall chinook salmon stocks. North Pacific Anadromous Fish Commission Bulletin 2:221-232.

Oldenburg, E.W., B.J. Goodman, G.A. McMichael, and R.B. Langshaw. 2012. Forms of Production Loss During the Early Life History of Fall Chinook Salmon in the Hanford Reach of the Columbia River. Prepared for the Public Utility District No. 2 of Grant County, Ephrata, WA. Contract Number 430-2464.

Pearsons, T. N. and C. A. Busack. 2012. PCD Risk 1: A tool for assessing and reducing ecological risks of hatchery operations in freshwater. Environmental Biology of Fishes 94:45-65. DOI:10.1007/s10641-011-9926-8.

Pearsons, T. N., A. R. Murdoch, G. Mackey, K. G. Murdoch, T. W. Hillman, M. R. Cooper, and J. L. Miller. 2012. Ecological risk assessment of multiple hatchery programs in the upper Columbia watershed using Delphi and modeling approaches. Environmental Biology of Fishes 94:87-100. DOI 10.1007/s10641-011-9884-1.

Pearsons, T. N., and C. W. Hopley. 1999. A practical approach for assessing ecological risks associated with fish stocking programs. Fisheries 24(9):16-23.

Pearsons, T. N., and R. B. Langshaw. 2009. Monitoring and evaluation plan for Grant PUDs Salmon and steelhead supplementation Programs. Grant PUD, Ephrata, Washington.

PSC (Pacific Salmon Commission). 2013. 2013 Exploitation Rate Analysis and Model Calibration - Volume One. A report of the Pacific Salmon Commission Joint Chinook Technical Committee. Technical Report (14)-1 V.1.

Richards, S. P., P. A. Hoffarth, and T. N. Pearsons. 2013. Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2012-13. Public Utility District Number 2 of Grant County, Ephrata, Washington.

Richards, S. P., and T. N. Pearsons. 2014. Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2013-14. Public Utility District Number 2 of Grant County, Ephrata, Washington.

Richards, S. P., and T. N. Pearsons. 2015. Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2014-15. Public Utility District Number 2 of Grant County, Ephrata, Washington.

Richards, S. P., and T. N. Pearsons. 2016. Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2015-16. Public Utility District Number 2 of Grant County, Ephrata, Washington.
Richards, S. P., and T. N. Pearsons. 2017. Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2016-17. Public Utility District Number 2 of Grant County, Ephrata, Washington.

Volk E.C., S.L. Schroder, and J.G. Grimm. 1999. Otolith Thermal Marking. Fisheries Research 43 (1999), pp. 205-219
USDOE - U.S. Department of Energy. In Press. Hanford Site Ecological Monitoring Report for Calendar Year 2016. HNF-61231, Rev. 0. U.S. Department of Energy, Richland Operations Office, Richland, Washington.

Weitkamp, L.A. 2010: Marine Distributions of Chinook Salmon from the West Coast of North America Determined by Coded Wire Tag Recoveries, Transactions of the American Fisheries Society, 139 (1), pp. 147-170
Zhou, S., 2002. Size-dependent recovery of Chinook salmon in carcass surveys. Transactions of the American Fisheries Society, 131(6), pp. 1194-1202.

## Appendix A Evaluation of Coded-Wire Tag Bias

We annually evaluate bias associated with estimates of the number of hatchery origin returns to PRH generated using coded-wire tags (CWT). Results from demographic sampling of the fall Chinook returns for 2010 through 2014 indicate that estimates of hatchery contributions to broodstock, the terminal sport fishery, and to escapement of the Hanford Reach and to the PRH trap calculated from otolith marks were substantially different from estimates generated using CWTs expanded by sampling rates and juvenile mark rates. This was of significant concern because many estimates such as stray rate, survival, origin, and harvest are dependent upon estimates generated from CWTs.

To assess the level of CWT recovery bias for any brood year, we used the following equation:

$$
\frac{\text { (\# of PRH Origin CWT Fish Recovered / \# of PRH Origin Fish Collected) }}{\text { CWT Mark Rate for Brood Year }}
$$

## Where:

\# of PRH origin fish collected = Estimate of the number of PRH origin fish for a specific age/brood year as determined by otoliths, scale aging, and expansion and pooling of age samples to represent total returns by age
\# of PRH Origin CWT Fish Recovered = Number of PRH origin CWT fish for a specific age/brood recovered at the hatchery ( $100 \%$ sample rate)
CWT Mark Rate = CWT marking rate for the specific brood year which is the number of CWT placed in fish divided by the estimated total number of fish at the time of marking.
If no CWT bias exists, the proportion of PRH CWT returns to the PRH CWT mark rate should equal 1.000. The values for CWT Recovery bias ranged from 0.573 to 4.620 for the different age/broods examined (Table A.1). Even though the datasets are not complete for recent brood years, it appears that the CWT Recovery bias is less pronounced since brood year 2011. The source of any bias is likely due to inappropriate expansion rate estimates resulting from nonrepresentative placement of CWT groups within the general population of rearing in the channel ponds. However, several other factors may contribute to the variation in CWT Recovery bias such as tag loss, CWT detection efficiency, or differential survival of tagged fish. In addition, the estimate of bias may be influenced by the level of precision of the estimated \# of PRH origin fish collected which varies for each age class of a given brood year due to size of the otolith subsample pulled from the demographic sample. In some cases, there are relatively few samples for age- 2 and 5 fish for a given brood year for this estimate.
Verification of the juvenile CWT rate at time of release is necessary to determine level of potential bias associated with reported juvenile CWT rates. Sampling for CWT rates at time of release has occurred at PRH since brood year 2014. Shortly prior to release, roughly 1,000 subyearlings from each of the five rearing ponds were captured and scanned with a V-detector to determine the proportions of adipose clipped CWT fish and adipose intact CWT fish within the sample. These proportions at release were compared to the proportions reported as ponded. In general, these two groups of proportions are similar for each brood year (Table A.2).

Table A. 1 Estimate of coded-wire tags bias for Priest Rapids origin returns to the hatchery, Brood Years 2007- 2015.

| Brood | Age | Proportion CWT Marked | \# of PRH <br> Origin <br> CWT Fish <br> Recovered | Estimated \# of PRH origin Fish Collected | Proportion of PRH Origin Brood Return CWT | Proportion of PRH CWT <br> Returns to the PRH CWT Mark Rate (CWT Recovery Bias) | Primary <br> Detector Type |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 5 | 0.045 | 48 | 928 | 0.052 | 1.161 | Blue Wand |
| 2007 | 4 | 0.045 | 280 | 10,977 | 0.026 | 0.573 | Blue Wand |
| 2007 | 3 | 0.045 | 410 | 14,073 | 0.029 | 0.654 | Blue Wand |
| 2007 | 2 |  | otolith dat | lected duri | turn year 200 |  |  |
| 2008 | 5 | 0.032 | 2 | 31 | 0.065 | 2.026 | Blue Wand |
| 2008 | 4 | 0.032 | 81 | 3,029 | 0.027 | 0.840 | Blue Wand |
| 2008 | 3 | 0.032 | 124 | 5,606 | 0.022 | 0.695 | Blue Wand |
| 2008 | 2 | 0.032 | 57 | 2,578 | 0.022 | 0.694 | Blue Wand |
| 2009 | 5 | 0.243 | 407 | 1,980 | 0.206 | 0.846 | R9500 |
| 2009 | 4 | 0.243 | 1,081 | 6,025 | 0.179 | 0.739 | Blue Wand |
| 2009 | 3 | 0.243 | 2,309 | 13,713 | 0.168 | 0.693 | Blue Wand |
| 2009 | 2 | 0.243 | 628 | 3,083 | 0.204 | 0.839 | Blue Wand |
| 2010 | 6 | 0.237 | 23 | 21 | 1.095 | 4.620 | R9500 |
| 2010 | 5 | 0.237 | 999 | 2,375 | 0.421 | 1.774 | R9500 |
| 2010 | 4 | 0.237 | 8,719 | 39,621 | 0.220 | 0.928 | R9500 |
| 2010 | 3 | 0.237 | 5,828 | 32,014 | 0.182 | 0.768 | Blue Wand |
| 2010 | 2 | 0.237 | 1,498 | 8,932 | 0.168 | 0.707 | Blue Wand |
| 2011 | 6 | 0.169 | 10 | 47 | 0.213 | 0.258 | R9500 |
| 2011 | 5 | 0.169 | 395 | 2,561 | 0.154 | 0.912 | R9500 |
| 2011 | 4 | 0.169 | 2,988 | 19,909 | 0.150 | 0.887 | R9500 |
| 2011 | 3 | 0.169 | 2,596 | 19,692 | 0.132 | 0.779 | R9500 |
| 2011 | 2 | 0.169 | 349 | 3,008 | 0.116 | 0.686 | R9500 |
| 2012 | 5 | 0.177 | 1,913 | 11,259 | 0.170 | 0.961 | R9500 |
| 2012 | 4 | 0.177 | 2,206 | 13,821 | 0.160 | 0.904 | R9500 |
| 2012 | 3 | 0.177 | 5,933 | 34,082 | 0.174 | 0.986 | R9500 |
| 2012 | 2 | 0.177 | 1,910 | 11,259 | 0.170 | 0.961 | R9500 |
| 2013 | 4 | 0.166 | 1,530 | 8,695 | 0.164 | 0.998 | R9500 |
| 2013 | 3 | 0.166 | 1,805 | 10,967 | 0.165 | 0.991 | R9500 |
| 2013 | 2 | 0.166 | 545 | 3,327 | 0.164 | 0.986 | R9500 |
| 2014 | 3 | 0.172 | 483 | 3289 | 0.147 | 0.856 | R9500 |
| 2014 | 2 | 0.172 | 78 | 486 | 0.160 | 0.935 | R9500 |
| 2015 | 2 | 0.166 | 183 | 1,219 | 0.150 | 0.903 | R9500 |
| CWT <br> Recovery Bias |  | Ages |  |  |  |  |  |
|  |  | Mean | 0.845 | 0.805 | 0.852 | 1.267 | 3.054 |

Table A.2. Proportions of coded-wire tagged juvenile fish reported ponded and the proportions of coded-wire tagged fish sampled at time of release, Brood Years 2014-17.

| Coded-wire sampling at release, Brood Year 2014 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# of Fish | Pond E | Pond D | Pond C | Pond B | Pond A | Total |
| Fish Released | 1,425,371 | 1,457,198 | 1,400,956 | 1,444,918 | 1,311,100 | 7,039,543 |
| $\mathrm{N}=$ | 1,040 | 1,024 | 1,018 | 1,023 | 1,565 | 5,670 |
| CWT Only Sampled | 98 | 85 | 79 | 67 | 220 | 549 |
| Ad-CWT Sampled | 102 | 69 | 73 | 86 | 165 | 495 |
| Proportion of Release Tagged |  |  |  |  |  |  |
| CWT Only | 8.5\% | 8.3\% | 8.6\% | 8.2\% | 9.0\% | 8.5\% |
| Ad-CWT | 8.5\% | 8.2\% | 8.6\% | 8.7\% | 8.7\% | 8.5\% |
| Proportion of Sample Tagged |  |  |  |  |  |  |
| CWT Only | 9.4\% | 8.3\% | 7.8\% | 6.5\% | 14.1\% | 9.7\% |
| Ad-CWT | 9.8\% | 6.7\% | 7.2\% | 8.4\% | 10.5\% | 8.7\% |
| Coded-wire sampling at release, Brood Year 2015 |  |  |  |  |  |  |
| \# of Fish | Pond E | Pond D | Pond C | Pond B | Pond A | Total |
| Fish Released | 1,445,733 | 1,448,510 | 1,507,753 | 1,512,437 | 1,327,621 | 7,242,054 |
| $\mathrm{N}=$ | 1,015 | 995 | 991 | 1,048 | 1,021 | 5,070 |
| CWT Only Sampled | 91 | 86 | 77 | 62 | 76 | 392 |
| Ad-CWT Sampled | 71 | 87 | 79 | 71 | 80 | 388 |
| Proportion of Release Tagged |  |  |  |  |  |  |
| CWT Only | 8.1\% | 8.6\% | 8.3\% | 7.5\% | 9.1\% | 8.3\% |
| Ad-CWT | 8.3\% | 8.6\% | 7.7\% | 8.0\% | 9.1\% | 8.3\% |
| Proportion of Sample Tagged |  |  |  |  |  |  |
| CWT Only | 9.0\% | 8.6\% | 7.8\% | 5.9\% | 7.4\% | 7.7\% |
| Ad-CWT | 7.0\% | 8.7\% | 8.0\% | 6.8\% | 7.8\% | 7.7\% |
| Coded-wire sampling at release, Brood Year 2016 |  |  |  |  |  |  |
| \# of Fish | Pond E | Pond D | Pond C | Pond $B$ | Pond A | Total |
| Fish Released | 1,401,157 | 1,455,960 | 1,450,776 | 1,487,339 | 1,211,019 | 7,006,251 |
| $\mathrm{N}=$ | 1,031 | 1,317 | 2,228 | 1,117 | 1,181 | 6,874 |
| CWT Only Sampled | 119 | 103 | 205 | 116 | 120 | 663 |
| Ad-CWT Sampled | 101 | 96 | 224 | 112 | 117 | 650 |
| Proportion of Release Tagged |  |  |  |  |  |  |
| CWT Only | 8.6\% | 8.3\% | 8.3\% | 8.1\% | 10.0\% | 8.6\% |
| Ad-CWT | 8.6\% | 8.3\% | 8.3\% | 8.1\% | 10.0\% | 8.6\% |
| Proportion of Sample Tagged |  |  |  |  |  |  |
| CWT Only | 11.5\% | 7.8\% | 9.2\% | 10.4\% | 10.2\% | 9.6\% |
| Ad-CWT | 9.8\% | 7.3\% | 10.1\% | 10.0\% | 9.9\% | 9.5\% |

Table A. 2 Continued

| Coded-wire sampling at release, Brood Year 2017 |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| \# of Fish | Pond E | Pond D | Pond C | Pond B | Pond A | Total |  |
| Fish Released | $1,632,887$ | $1,573,080$ | $1,615,297$ | $1,588,038$ | $1,594,137$ | $8,003,439$ |  |
| $\mathrm{~N}=$ | 1,046 | 1,260 | 1,022 | 1,173 | 1,044 | 5,545 |  |
| CWT Only Sampled | 88 | 143 | 74 | 87 | 85 | 477 |  |
| Ad-CWT Sampled | 81 | 164 | 71 | 77 | 67 | 460 |  |
| Proportion of Release Tagged |  |  |  |  |  |  |  |
| CWT Only | $7.5 \%$ | $7.6 \%$ | $7.5 \%$ | $7.6 \%$ | $7.6 \%$ | $7.6 \%$ |  |
| Ad-CWT | $7.2 \%$ | $7.7 \%$ | $7.5 \%$ | $7.6 \%$ | $7.6 \%$ | $7.5 \%$ |  |
| CWT Only | $8.4 \%$ | $11.3 \%$ | $7.2 \%$ | $7.4 \%$ | $8.1 \%$ | $8.6 \%$ |  |
| Ad-CWT | $7.7 \%$ | $13.0 \%$ | $6.9 \%$ | $6.6 \%$ | $6.4 \%$ | $8.3 \%$ |  |

Assessment of CWT detection efficiency has been conducted annually at PRH since 2010 during adult fish sampling with enhancement to these procedures developed over time. In 2013, M\&E staff randomly selected a total of 1,063 quality control fish being surplused with no CWT detected using the T-wand (Table A.3). These fish were then re-scanned with the older bluewand. If CWT was detected using a blue wand the fish was again scanned using the T-wand. In such a manner the missed CWT could be inferred as a result of operator error or the inability of the T-wand to detect the CWT. On a few occasions the T-wand did not detect a CWT identified by the blue-wand. In these instances, the snouts were removed from the fish to increase the likelihood of detection and then passed through a V-detector. Similar to quality control results for previous years, there were only a few ( 4 tags; $0.4 \%$ of the sample CWT detections observed in the quality control fish sampled that were not detected initially by the T-wands.

Table A. 3 Quality control results for coded-wire tag detection at Priest Rapids Hatchery, Brood Years 2013-2017.

| Brood Year | Initial Device | QC Device | \# Sampled | \# Missed CWT | P^ CWT Missed |
| ---: | :---: | :---: | :---: | :---: | ---: |
| 2013 | T-Wand | Blue Wand | 1,063 | 4 | 0.004 |
| 2014 | R9500 | T-Wand | 2,000 | 3 | 0.002 |
| 2015 | R9500 | T-Wand | 4,596 | 2 | 0.000 |
| 2016 | R9500 | T-Wand | 5,943 | 3 | 0.001 |
| 2017 | R9500 | T-Wand | 1,744 | 3 | 0.002 |

During 2013 and 2014, we found the T-wands to be overly sensitive leading to false positive detections and additional work related to processing snouts to extract CWTs. On October 2, 2014 we set up two series R9500 detectors to expedite scanning for CWTs (Figure 1). The detectors were checked for proper operation each day prior to scanning any fish. Informal quality control checks occurred daily during the first two weeks of operation in order to identify the detection efficiency of each detector. These checks involved running 100 fish through each machine and then re-scanning the fish with the T-wands. A total of 2,000 fish were passed through the R9500 units of which 422 were identified to possess a CWT. Of these fish, 419 signaled positive for a CWT during the initial scanning. The three fish possessing a CWT that were not identified by the R9500 during the initial scanning were correctly detected when re-ran though the detectors. The
missed fish were likely the result of passing fish through the detectors too rapidly which can interfere with the operation of the flip gates.

R9500 detectors were used to scan the vast majority of fish surplused at PRH during 2015, 2016 and 2017. During each of these years, the first group of fish handled each day was used to test the CWT detection of each R9500 detector. The test fish that a CWT was not detected were rescanned with a T-wand to assess the performance of the R9500 detectors. The results for all three years suggest that very few possessing a CWT are missed by the R9500 detectors.

The methods describe here do not provide a definitive estimate of undetected CWTs for fish sampled at PRH. We make an assumption, that if the CWT detection wands and R9500 units do not detect a CWT in a given fish, then it did not possess a tag. Based on this assumption, the CWT detection efficiency at PRH is likely greater than 99\%. Therefore, the magnitude of the CWT recovery bias expressed in Table 1 is not likely due to poor CWT detection efficiency.


Figure 1. Series R9500 Coded-wire tag detectors used at Priest Rapids Hatchery, 2014

## Appendix B

Recovery of coded-wire tags collected from adult returns to the Priest Rapids Hatchery Volunteer Trap during Return Year 2017

| Code | Tags | BY | Race | Age | Stock | CWT Release |  |  | Expansion |  | Return to PRH |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  | Date | AD CWT | CWT Only | All CWT | ADC WT |  |  |
| 090681 | 9 | 2012 | Fall | 5 | RSH | 2013 |  |  | 14.706 | 15.113 | 132 | 0.7\% |
| 090682 | 1 | 2012 | Fall | 5 | Umatilla R | 2014 | 229,6 |  | 1.040 | 2.046 | 1 | 0.0\% |
| 090683 | 3 | 2012 | Fall | 5 | Umatilla R | 2014 |  |  | 1.040 | 2.046 | 3 | 0.0\% |
| 090704 | 1 | 2012 | Fall | 5 | Umatilla R | 2013 | 141,0 |  | 1.986 | 1.987 | 2 | 0.0\% |
| 090816 | 5 | 2013 | Fall | 4 | Umatilla R | 2014 | 169,2 |  | 1.9 | 1.9 | 9 | 0.1\% |
| 090817 | 7 | 2013 | Fall | 4 | Umatilla R | 2014 | 163,1 |  | 1.9 | 1.9 | 13 | 0.1\% |
| 090863 | 35 | 2013 | Fall | 4 | RSH | 2011 |  |  | 15.096 | 15.287 | 528 | 3.0\% |
| 090867 | 2 | 2013 | Fall | 4 | Umatilla R | 2015 | 35,77 |  | 1.0 | 1.0 | 2 | 0.0\% |
| 090868 | 1 | 2013 | Fall | 4 | Umatilla R | 2015 | 28,92 |  | 1.027 | 1.030 | 1 | 0.0\% |
| 090870 | 3 | 2013 | Fall | 4 | Umatilla R | 2015 |  |  | 1.027 | 1.030 | 3 | 0.0\% |
| 090871 | 2 | 2013 | Fall | 4 | Umatilla R | 2015 |  |  | 1.027 | 1.030 | 2 | 0.0\% |
| 090888 | 1 | 2014 | Fall | 3 | Snake R | 2015 |  |  | 4.279 | 4.284 | 4 | 0.0\% |
| 090909 | 20 |  |  |  | ODFW |  |  |  |  |  | 0 | 0.0\% |
| 090917 | 7 | 2014 | Fall | 3 | Umatilla R | 2015 | 161,6 |  | 4.040 | 4.040 | 28 | 0.2\% |
| 090921 | 1 | 2014 | Fall | 3 | RSH | 2015 | 227,9 |  | 15.726 | 15.751 | 16 | 0.1\% |
| 090944 | 9 | 2014 | Fall | 3 | Umatilla R | 2016 |  |  | 1.007 | 1.009 | 9 | 0.1\% |
| 090945 | 8 | 2014 | Fall | 3 | Umatilla R | 2016 |  |  | 1.007 | 1.013 | 8 | 0.0\% |
| 090946 | 3 | 2014 | Fall | 3 | Umatilla R | 2016 |  |  | 1.002 | 1.007 | 3 | 0.0\% |
| 090981 | 2 | 2015 | Fall | 2 | Umatilla R | 2016 | 170,5 |  | 1.617 | 1.617 | 3 | 0.0\% |
| 090982 | 1 | 2015 | Fall | 2 | RSH | 2016 | 191,2 |  | 7.688 | 7.688 | 8 | 0.0\% |
| 090983 | 2 | 2015 | Fall | 2 | RSH | 2016 | 191,2 |  | 7.7 | 7.7 | 15 | 0.1\% |
| 091010 | 3 | 2015 | Fall | 2 | Umatilla R | 2016 | 167,3 |  | 1.6 | 1.6 | 5 | 0.0\% |
| 091013 | 1 | 2015 | Fall | 2 | Snake R |  | 247,4 |  | 4.208 | 4.215 | 4 | 0.0\% |
| 220237 | 1 | 2013 | Fall | 4 | Snake R | 2014 | 102,8 |  | 1.702 | 5.134 | 2 | 0.0\% |
| 220254 | 1 | 2015 | Fall | 2 | Snake R | 2016 | 104,4 |  | 2.085 | 6.280 | 2 | 0.0\% |
| 610441 | 1 | 2013 | Fall | 4 | Hanford R | 2014 | 4,831 |  |  |  | 0 | 0.0\% |
| 610447 | 1 | 2012 | Fall | 5 | Hanford R | 2013 | 6,884 |  |  |  | 0 | 0.0\% |
| 636371 | 5 | 2011 | Fall | 6 | PRH | 2012 |  | 598,0 | 5.9 | 11.8 | 0 | 0.0\% |
| 636372 | 5 | 2011 | Fall | 6 | PRH | 2012 | 595,6 |  | 5.9 | 11.8 | 0 | 0.0\% |
| 636505 | 1 | 2012 | Summ | 5 | Wells | 2014 |  |  | 1.024 | 1.028 | 0 | 0.0\% |
| 636507 | 189 | 2012 | Fall | 5 | PRH | 2013 | 603,9 |  | 5.662 | 11.297 | 0 | 0.0\% |
| 636508 | 215 | 2012 | Fall | 5 | PRH | 2013 |  | 601,0 | 5.662 | 11.297 | 30 | 0.2\% |
| 636676 | 1 | 2013 | Fall | 4 | L.Columbia | 2014 | 242,9 |  | 5.393 | 5.393 | 30 | 0.2\% |
| 636679 | 1 | 2014 | Fall | 3 | L.Columbia | 2015 | 449,3 |  | 1.000 | 1.000 | 1 | 0.0\% |
| 636681 | 731 | 2013 | Fall | 4 | PRH | 2014 | 603,7 |  | 6.017 | 12.093 | 1070 | 6.0\% |
| 636682 | 799 | 2013 | Fall | 4 | PRH | 2014 |  | 603,8 | 6.017 | 12.093 | 1217 | 6.8\% |
| 636738 | 1 | 2013 | Fall | 4 | Snake R | 2014 | 185,7 |  | 1.029 | 1.029 | 5 | 0.0\% |
| 636739 | 1 | 2013 | Fall | 4 | Snake R | 2014 |  |  | 1.997 | 1.997 | 1 | 0.0\% |
| 636836 | 244 | 2014 | Fall | 3 | PRH | 2015 | 604,8 |  | 5.826 | 11.660 | 4399 | 24.7 |
| 636837 | 239 | 2014 | Fall | 3 | PRH | 2015 |  | 604,8 | 5.826 | 11.660 | 4808 | 27.0 |
| 636967 | 85 | 2015 | Fall | 2 | PRH | 2016 | 605,4 |  | 5.982 | 11.960 | 1 | 0.0\% |
| 636968 | 98 | 2015 | Fall | 2 | PRH | 2016 |  | 605,0 | 5.982 | 11.960 | 2 | 0.0\% |
| 637184 | 1 | 2016 | Fall | 1 | PRH | 2017 |  | 120,3 | 5.788 | 11.580 | 1422 | 8.0\% |
| Total | 2753 |  | 17,799 |  | ered at PRH |  |  |  |  |  | 16,283 | 91.5 |

## Appendix C

Juvenile fish health inspections for Priest Rapids Hatchery fall Chinook salmon, Brood Years 1998-2017. The description in the Condition column indicates the presence of a certain condition within at least one of the fish examined.

| Hatchery/Stock | Date | Brood | Condition |
| :---: | :---: | :---: | :---: |
| Priest Rapids | 23-Feb-99 | 1998 | Healthy |
|  | 22-Mar-99 | 1998 | Healthy |
|  | 23-Apr-99 | 1998 | Healthy |
|  | 25-May-99 | 1998 | Dropout Syndrome \& Bacterial Gill Disease |
|  | 08-Jun-99 | 1998 | Bacterial Kidney Disease |
| Priest Rapids | 06-Mar-00 | 1999 | Healthy |
|  | 14-Apr-00 | 1999 | Healthy |
|  | 16-May-00 | 1999 | Healthy |
|  | 12-Jun-00 | 1999 | Healthy |
| Priest Rapids | 23-Feb-01 | 2000 | Healthy |
|  | 05-Apr-01 | 2000 | Healthy |
|  | 07-May-01 | 2000 | Healthy |
|  | 06-Jun-01 | 2000 | Healthy |
| Priest Rapids | 13-Feb-02 | 2001 | Healthy |
|  | 01-Mar-02 | 2001 | Coagulated Yolk Syndrome |
|  | 22-Apr-02 | 2001 | Healthy |
|  | 10-Jun-02 | 2001 | Healthy |
| Priest Rapids | 07-Mar-03 | 2002 | Healthy |
|  | 15-Apr-03 | 2002 | Healthy |
|  | 02-Jun-03 | 2002 | Healthy |
| Priest Rapids | 01-Apr-04 | 2003 | Healthy |
|  | 06-May-04 | 2003 | Healthy |
|  | 07-Jun-04 | 2003 | Healthy |
| Priest Rapids | 11-Mar-05 | 2004 | Healthy |
|  | 14-Apr-05 | 2004 | Healthy |
|  | 1-Jun-05 | 2004 | Healthy |
| Priest Rapids | 6-Mar-06 | 2005 | Healthy |
|  | 25-Apr-06 | 2005 | Healthy |
|  | 13-Jun-06 | 2005 | Healthy |
| Priest Rapids | 9-Mar-07 | 2006 | Healthy |
|  | 19-Apr-07 | 2006 | Healthy |
|  | 1-Jun-07 | 2006 | Healthy |
| Priest Rapids | 12-Feb-08 | 2007 | Coagulated Yolk Syndrome observed in some fish sampled |
|  | 23-Apr-08 | 2007 | Healthy |
|  | 4-Jun-08 | 2007 | Healthy |
| Priest Rapids | 12-Feb-09 | 2008 | Coagulated Yolk Syndrome observed in some fish sampled |
|  | 22-Apr-09 | 2008 | Healthy |
|  | 8-Jun-09 | 2008 | Healthy |
| Priest Rapids | 18-Feb-10 | 2009 | Coagulated Yolk Syndrome observed in some fish sampled |
|  | 1-Apr-10 | 2009 | Healthy |
|  | 19-May-10 | 2009 | Healthy |
| Priest Rapids | 25-Mar-11 | 2010 | Healthy |


| Hatchery/Stock | Date | Brood | Condition |
| :---: | :---: | :---: | :---: |
|  | 18-Apr-11 | 2010 | Healthy |
|  | 06-Jun-11 | 2010 | Healthy |
| Priest Rapids | 01-Mar-12 | 2011 | Healthy |
|  | 26-Apr-12 | 2011 | Healthy |
|  | 24-May-12 | 2011 | Healthy |
| Priest Rapids | 11-Feb-13 | 2012 | Healthy |
|  | 3-Mar-13 | 2012 | Healthy |
|  | 29-Apr-13 | 2012 | Healthy |
|  | 28-May-13 | 2012 | Healthy |
| Priest Rapids | 27-Mar-14 | 2013 | Dropout Syndrome present |
|  | 23-Apr-14 | 2013 | Dropout Syndrome present |
|  | 29-May-14 | 2013 | Healthy |
| Priest Rapids | 26-Feb-15 | 2014 | Coagulated Yolk Syndrome observed in some fish sampled |
|  | 26-Mar-15 | 2014 | Healthy |
|  | 21-Apr-15 | 2014 | Healthy |
|  | 28-May-15 | 2014 | Healthy |
|  | 22-June-15 | 2014 | Columnaris present in some fish sampled from CH Pond B. |
| Priest Rapids | 24-Feb-16 | 2015 | Healthy |
|  | 15-Mar-16 | 2015 | Coagulated Yolk Syndrome observed in some fish sampled |
|  | 15-June-16 | 2015 | Mild Ich infection but healthy and ready for release |
| Priest Rapids | 24-Feb-17 | 2016 | Presence of bacterial gill disease in Raceway Bank D and E |
|  | 21-Mar-17 | 2016 | Presence of bacterial gill disease in Raceway Pond B2 |
|  | 6-June-17 | 2016 | Mild Ich infection in Channel Ponds A, B, C |
| Priest Rapids | 21-Mar-18 | 2017 | Healthy |
|  | 19-Apr-18 | 2017 | Bacterial gill dieses present in Raceway Pond C4 |
|  | 7-May-18 | 2017 | Bacterial gill dieses present in Raceway Ponds C2 and C3 |
|  | 17-May-18 | 2017 | Re-examine Raceway Ponds C2 and C3 found fish healthy |
|  | 17-May-18 | 2017 | Pre-release examine Raceway Banks D and E found fish healthy C2 and C3 found fish healthy |
|  | 6-June-18 | 2017 | Pre-release examine of Raceway Banks A and B found fish healthy |

## Appendix D

Number and percent of fall Chinook salmon redds counted in different reaches of the Columbia River, 2001-2017. Data for years 2001-2010 was collected by staff with Pacific Northwest National Laboratory. Data for years 2001-2017 was collected by staff with Environmental Assessment Services, LLC.

| Location | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Islands 11-21 | 297 | 509 | 554 | 337 | 708 | 36 | 302 | 371 | 176 | 562 |
| Islands 8-10 | 480 | 865 | 1,133 | 867 | 1,067 | 435 | 338 | 416 | 722 | 870 |
| Near Island 7 | 350 | 280 | 455 | 415 | 500 | 873 | 311 | 360 | 380 | 457 |
| Island 6 (lower) | 750 | 940 | 1,241 | 1,084 | 1,229 | 289 | 615 | 753 | 878 | 1,135 |
| Island 4, 5,6 | 1,130 | 1,165 | 1,242 | 1,655 | 1,130 | 934 | 655 | 960 | 796 | 1,562 |
| Near Island 3 | 460 | 249 | 475 | 325 | 345 | 1,305 | 152 | 230 | 285 | 244 |
| Near Island 2 | 780 | 955 | 850 | 960 | 895 | 523 | 455 | 555 | 459 | 657 |
| Near Island 1 | 35 | 235 | 270 | 330 | 255 | 253 | 47 | 148 | 160 | 324 |
| Coyote Rapids | 16 | 63 | 354 | 180 | 304 | 150 | 10 | 29 | 34 | 49 |
| China Bar | 20 | 25 | 85 | 75 | 28 | 52 | 3 | 35 | 1,090 | 299 |
| Vernita Bar | 1,930 | 2,755 | 2,806 | 2,240 | 1,430 | 1,658 | 1,135 | 1,731 | 16 | 2,658 |
| Total | 6,248 | 8,041 | 9,465 | 8,468 | 7,891 | 6,508 | 4,023 | 5,588 | 4,996 | 8,817 |
| Location | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| Islands 11-21 | 5\% | 6\% | 6\% | 4\% | 9\% | 1\% | 8\% | 7\% | 4\% | 6\% |
| Islands 8-10 | 8\% | 11\% | 12\% | 10\% | 14\% | 7\% | 8\% | 7\% | 14\% | 10\% |
| Near Island 7 | 6\% | 3\% | 5\% | 5\% | 6\% | 13\% | 8\% | 6\% | 8\% | 5\% |
| Island 6 (lower) | 12\% | 12\% | 13\% | 13\% | 16\% | 4\% | 15\% | 13\% | 18\% | 13\% |
| Island 4, 5, 6 | 18\% | 14\% | 13\% | 20\% | 14\% | 14\% | 16\% | 17\% | 16\% | 18\% |
| Near Island 3 | 7\% | 3\% | 5\% | 4\% | 4\% | 20\% | 4\% | 4\% | 6\% | 3\% |
| Near Island 2 | 12\% | 12\% | 9\% | 11\% | 11\% | 8\% | 11\% | 10\% | 9\% | 7\% |
| Near Island 1 | 1\% | 3\% | 3\% | 4\% | 3\% | 4\% | 1\% | 3\% | 3\% | 4\% |
| Coyote Rapids | $>1 \%$ | 1\% | 4\% | 2\% | 4\% | 2\% | >1\% | 1\% | 1\% | 1\% |
| China Bar | $>1 \%$ | >1\% | 1\% | 1\% | >1\% | 1\% | >1\% | 1\% | 22\% | 3\% |
| Vernita Bar | 31\% | 34\% | 30\% | 26\% | 18\% | 25\% | 28\% | 31\% | >1\% | 30\% |
| Location | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 280 |  |  | (07-16) Mean |
| Islands 11-21 | 676 | 533 | 798 | 906 | 1,193 | 861 | 900 |  |  | 638 |
| Islands 8-10 | 814 | 807 | 2,200 | 1,565 | 3,145 | 1,735 | 670 |  |  | 1,261 |
| Near Island 7 | 670 | 700 | 655 | 1,100 | 800 | 670 | 900 |  |  | 610 |
| Island 6 (lower) | 1,181 | 1,375 | 3,340 | 2,530 | 2,315 | 1,807 | 911 |  |  | 1,593 |
| Island 4, 5,6 | 1,524 | 1,195 | 2,650 | 2,080 | 2,540 | 2,270 | 500 |  |  | 1,623 |
| Near Island 3 | 525 | 475 | 1,000 | 1,000 | 1,100 | 600 | 790 |  |  | 561 |
| Near Island 2 | 653 | 528 | 1,700 | 2,050 | 1,900 | 1,140 | 330 |  |  | 1,010 |
| Near Island 1 | 295 | 340 | 900 | 500 | 1,000 | 340 | 80 |  |  | 405 |
| Coyote Rapids | 44 | 29 | 520 | 500 | 765 | 255 | 75 |  |  | 224 |
| China Bar | 67 | 68 | 100 | 60 | 1,730 | 80 | 3210 |  |  | 353 |
| Vernita Bar | 2,466 | 2,318 | 3,535 | 3,650 | 4,190 | 3,510 | 8,646 |  |  | 2,521 |
| Total | 8,915 | 8,368 | 17,398 | 15,951 | 20,678 | 13,268 | 280 |  |  | 10,799 |
| Location | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |  |  | (07-16) Mean |
| Islands 11-21 | 8\% | 6\% | 5\% | 6\% | 6\% | 6\% | 3\% |  |  | 6\% |
| Islands 8-10 | 9\% | 10\% | 13\% | 10\% | 15\% | 13\% | 10\% |  |  | 12\% |
| Near Island 7 | 8\% | 8\% | 4\% | 7\% | 4\% | 5\% | 8\% |  |  | 6\% |
| Island 6 (lower) | 13\% | 16\% | 19\% | 16\% | 11\% | 14\% | 10\% |  |  | 14\% |
| Island 4, 5, 6 | 17\% | 14\% | 15\% | 13\% | 12\% | 17\% | 11\% |  |  | 15\% |
| Near Island 3 | 6\% | 6\% | 6\% | 6\% | 5\% | 5\% | 6\% |  |  | 5\% |
| Near Island 2 | 7\% | 6\% | 10\% | 13\% | 9\% | 9\% | 9\% |  |  | 9\% |
| Near Island 1 | 3\% | 4\% | 5\% | 3\% | 5\% | 3\% | 4\% |  |  | 4\% |
| Coyote Rapids | >1\% | >1\% | 3\% | 3\% | 4\% | 2\% | 1\% |  |  | 2\% |
| China Bar | 1\% | 1\% | 1\% | 0\% | 8\% | 1\% | 1\% |  |  | 3\% |
| Vernita Bar | 28\% | 28\% | 20\% | 23\% | 20\% | 26\% | 37\% |  |  | 24\% |

## Appendix E

Historical numbers of Chinook salmon carcasses recovered during the annual Hanford Reach fall Chinook salmon carcass survey, Return Years 1991-2017.

| Return Year | Total Recoveries | Total Escapement | Proportion of Escapement Recovered |
| :---: | :---: | :---: | :---: |
| 1991 | 2,519 | 52,196 | 0.048 |
| 1992 | 2,221 | 41,952 | 0.053 |
| 1993 | 3,340 | 37,347 | 0.089 |
| 1994 | 5,739 | 63,103 | 0.091 |
| 1995 | 3,914 | 55,208 | 0.071 |
| 1996 | 4,529 | 43,249 | 0.105 |
| 1997 | 5,053 | 43,493 | 0.116 |
| 1998 | 4,456 | 35,393 | 0.126 |
| 1999 | 4,412 | 29,812 | 0.148 |
| 2000 | 10,556 | 48,020 | 0.220 |
| 2001 | 6,072 | 59,848 | 0.101 |
| 2002 | 8,402 | 84,509 | 0.099 |
| 2003 | 13,573 | 100,840 | 0.135 |
| 2004 | 11,030 | 87,696 | 0.126 |
| 2005 | 8,491 | 71,967 | 0.118 |
| 2006 | 5,972 | 51,701 | 0.116 |
| 2007 | 3,115 | 22,272 | 0.140 |
| 2008 | 5,455 | 29,058 | 0.188 |
| 2009 | 5,318 | 36,720 | 0.145 |
| 2010 | 9,779 | 87,016 | 0.112 |
| 2011 | 8,391 | 75,256 | 0.111 |
| 2012 | 6,814 | 57,710 | 0.118 |
| 2013 | 13,071 | 174,651 | 0.075 |
| 2014 | 16,756 | 183,749 | 0.091 |
| 2015 | 17,738 | 266,346 | 0.086 |
| 2016 | 8,886 | 116,421 | 0.076 |
| 2017 | 5,591 | 73,759 | 0.076 |
| Mean | 7,452 | 75,159 | 0.110 |

## Appendix $F$

Estimated escapements for fall Chinook spawning in Hanford Reach and Priest Rapids Dam pool, Return Year 2017.

| Count Source |  | 2017 Hanford Reach Fall Chinook Escapement Estimate |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Adult | Jack | Total |
|  | McNary ${ }^{1}$ | 152,185 | 12,014 | 164,199 |
|  | Wanapum ${ }^{2}$ | 19,041 | 1,620 | 20,661 |
|  | Priest Rapids ${ }^{3}$ | 30,972 | 2,306 | 33,278 |
|  | Priest Rapids Fallback Adjustment ${ }^{4}$ | 8,224 | 612 | 8,836 |
|  | Ice Harbor ${ }^{5}$ | 26,393 | 5,057 | 31,450 |
|  | Prosser ${ }^{6}$ | 1,947 | 356 | 2,303 |
|  | Priest Rapids Hatchery | 15,572 | 1,441 | 17,012 |
|  | Angler Broodstock Collection | 492 | 0 | 492 |
|  | Ringold Springs Hatchery | 1,244 | 47 | 1,291 |
| 匆 | Hanford Sport Harvest | 11,496 | 872 | 12,368 |
|  | Yakima River Sport Harvest | 470 | 16 | 486 |
|  | Wanapum Tribal Fishery | 0 | 0 | 0 |
|  | Yakima River (Lower) ${ }^{7}$ | 520 | 75 | 595 |
|  | Hanford Reach + Priest Pool | 74,010 | 2,530 | 76,540 |
|  | Priest Pool Return | 2,707 | 74 | 2,781 |
|  | Hanford Reach Escapement | 71,303 | 2,456 | 73,759 |

${ }^{1}$ McNaryDam fish counts: August 9 - October 31
${ }^{2}$ Wanapum Dam fish counts, August 14 through November 5
${ }^{3}$ Priest Rapids Dam fish counts, August 18 through November 5. GCPUD continued counts through Nov 15 but McNary counts ended on Oct 31. Allowed 5 days to account for difference in passage timing
${ }^{4}$ Fallback/Reascension Adjustment estimate (26.6\%) based on 152 run of the river PIT tagged fish from the BOAFF and the lower Columbia River test fishery observed at Priest Rapids Dam and Priest Rapids Hatchery PIT tag arrays
${ }^{5}$ Ice Harbor counts ended on Oct 31
${ }^{6}$ Prosser counts, August 16 through November 5
${ }^{7}$ Escapement estimated by carcass counts versus Escapement regression (2000-2011)

| Count Source | 2017 Priest Rapids Pool Escapement |  |  |
| :--- | ---: | ---: | ---: |
|  | Adult |  | Jack |
| Total |  |  |  |
| Priest Rapids Adult Passage $^{3}$ | 30,972 | 2,306 | 33,278 |
| Priest Rapids Fallback Adjustment $^{2}$ | 8,224 | 612 | 8,836 |
| Wanapum Adult Passage $^{1}$ | 19,510 | 1,662 | 21,172 |
| Wanapum Dam Fallback Adjustment | Unknown | Unknown | Unknown |
| Wanapum Tribal Fishery Above PRD | 77 | 1 | 78 |
| OLAFT | 971 | 0 | 971 |
| Priest Rapids Pool Sport Fishery | 402 | 51 | 453 |
| Priest Rapids Dam Pool Escapement | $\mathbf{1 , 7 8 8}$ | $\mathbf{0}$ | $\mathbf{1 , 7 8 8}$ |

${ }^{1}$ Wanapum Dam fish counts, August 14 through November 5.
${ }^{2}$ Fallback/Reascension Adjustment estimate (26.6\%) based on 152 run of the river PIT tagged fish from the BOAFF and the lower Columbia River test fishery observed at Priest Rapids Dam and Priest Rapids Hatchery PIT tag arrays.
${ }^{3}$ Priest Rapids passage for fall Chinook based on counts from August 18 through November 15.

## Appendix G

## Carcass drift assessment

A common objective of hatchery monitoring and evaluation programs in the upper Columbia Watershed is to identify the spawning distribution of both hatchery and natural origin fish. Initially, we believed that the proportion of hatchery origin spawners (pHOS) could be calculated for each of five reaches within the Hanford Reach:

- Reach 1. Priest Rapids Dam to Vernita Bridge (14 km)
- Reach 2. Vernita Bridge to Island 2 ( 19 km )
- Reach 3. Island 2 to Powerline Towers at Hanford town site ( 21 km )
- Reach 4. Power line Towers to Wooded Island (21 km)
- Reach 5. Wooded Island to Interstate 182 Bridge (19 km),

However, previous carcass bias assessments within the Hanford Reach suggest a substantial amount of downstream carcass drift into lower reaches (Richards and Pearsons, 2013). Hence, it is uncertain that the carcass recovery locations directly represent spawner distributions in some locations. In order to gain a better understanding of natural post-spawn carcass drift, we tried two different approaches for tagging and releasing carcasses: tagging fish in place or releasing tagged fish over known spawning areas.
During 2014, we used a long pole to floy-tag 993 carcasses in place without moving them (Richards and Pearsons, 2015). Tagging occurred from October 26 through November 23. This prevented the collection of accurate size and gender data as many carcasses tagged were underwater. We anticipated that some carcasses would move downstream as river flows fluctuated. Recovery efforts occurred from November 1 through December 19. Recovery rates ranged from 31 - 37 \% for donor Sections 1, 2, and 4 (Table G.1).

## Table G. 1 Numbers of floy-tagged Chinook salmon carcasses released and recovered

 by donor reach within the Hanford Reach, Return Year 2014|  | Donor <br> Section 1 | Donor Section 2 | Donor Section 3 | Donor <br> Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Tagged by Donor Section | 486 | 107 | 225 | 176 | 994 |
| 1 | 143 |  |  |  | 146 |
| Fish Recoved by 2 | 1 | 32 |  |  | 34 |
| Fish Recovered by $\quad 3$ | 3 | 1 | 35 |  | 39 |
| Reciplent | 4 | 0 | 4 | 60 | 68 |
| 5 | 1 | 1 | 0 | 4 | 6 |
| P^ Recovered for each Donor Section | 0.319 | 0.327 | 0.173 | 0.364 | 0.295 |
| 1 | 0.942 |  |  |  |  |
| Proportion $\quad 2$ | 0.006 | 0.943 |  |  |  |
| Recovered by $\qquad$ | 0.019 | 0.029 | 0.897 |  |  |
| Section $\quad 4$ | 0.026 | 0.000 | 0.103 | 0.938 |  |
| 5 | 0.006 | 0.029 | 0.000 | 0.063 |  |
| 1 | 1.000 |  |  |  |  |
| Proportion $\qquad$ | 0.007 | 0.993 |  |  |  |
|  | 0.021 | 0.031 | 0.948 |  |  |
| $\begin{array}{ll}  \\ \\ \text { recipient Section } \end{array}$ | 0.025 | 0.000 | 0.097 | 0.879 |  |
| ¢ 5 | 0.066 | 0.297 | 0.000 | 0.636 |  |

Donor Section 3 had the lowest recovery rate at $17 \%$. We found that many tagged carcasses did not move from the tag sites; hence the results suggest that carcass drift was occurring at very low rates. We now believe that large portions of carcasses remain in their initial location of deposition.
During years 2015 through 2017, we adjusted our approach in attempt to mimic post-spawn fish dying near redd locations and subsequently drifting downstream. Each year, we operculumtagged roughly 1,000 intact carcasses, collected size and gender data and then redistributed them in the proximity of specific spawning areas within Sections $1-4$ (Figure G.1). Tagging occurred primarily in November for both years. Depths at release were visually estimated to range from 1 to 7 meters. River flow $\mathrm{m} / \mathrm{s}$ at release was not measured. No fish were released in eddies or slack water. Released carcasses were generally observed sinking quickly to the bottom and then slowly drifting downstream. Recovery efforts occurred during November and early December. During 2015, crews recovered 39 (3.9\%) tagged carcasses (Table G.2). The recovery rate was notably lower for fish released in Section 4 compared to the other sections. Although the numbers recovered were low, results show that large proportion of tagged fish recovered were found downstream of their adjacent donor section. During 2016 and 2017, crews recovered 45 and 42 tagged carcasses, respectively (Table G.3) (Table G.4). The recovery rate was notably lower for fish released in Section 2 compared to the other sections. Similar to the results of 2015, large proportions of tagged fish recovered during 2016 and 2017 were found downstream of their adjacent donor section.


Figure G. 1 Operculum Tagged Male fall Chinook in the Hanford Reach, 2015 Carcass Drift assessment.

Table G. 2 Numbers of operculum-tagged Chinook salmon carcasses released and recovered by donor reach within the Hanford Reach, Return Year 2015

|  | Donor Section 1 | Donor Section 2 | Donor Section 3 | Donor Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Ta | 231 | 62 | 343 | 362 | 998 |
| Fish Recovered by Recipient Section | 4 |  |  |  | 4 |
|  | 0 | 1 |  |  | 1 |
|  | 6 | 3 | 4 |  | 13 |
|  | 2 | 0 | 13 | 4 | 19 |
|  | 0 | 0 | 1 | 1 | 2 |
| P^ Recovered for each Donor Section | 0.052 | 0.065 | 0.052 | 0.014 | 0.039 |
| Proportion Recovered by Section | 0.333 |  |  |  |  |
|  | 0.000 | 0.250 |  |  |  |
|  | 0.500 | 0.750 | 0.222 |  |  |
|  | 0.167 | 0.000 | 0.722 | 0.800 |  |
|  | 0.000 | 0.000 | 0.056 | 0.200 |  |
| Proportion Recovered by Section into Recipient Section | 1.000 |  |  |  |  |
|  | 0.000 | 1.000 |  |  |  |
|  | 0.340 | 0.509 | 0.151 |  |  |
|  | 0.099 | 0.000 | 0.428 | 0.474 |  |
|  | 0.000 | 0.000 | 0.217 | 0.783 |  |

Table G. 3 Numbers of operculum-tagged Chinook salmon carcasses released and recovered by donor reach within the Hanford Reach, Return Year 2016

|  | Donor <br> Section 1 | Donor Section 2 | Donor Section 3 | Donor <br> Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Ta | 263 | 138 | 332 | 254 | 987 |
| Fish Recovered by Recipient Section | 3 |  |  |  | 3 |
|  | 0 | 0 |  |  | 0 |
|  | 10 | 0 | 3 |  | 13 |
|  | 7 | 2 | 10 | 5 | 24 |
|  | 0 | 0 | 1 | 4 | 5 |
| P^ Recovered for each Donor Section | 0.076 | 0.014 | 0.042 | 0.035 | 0.046 |
| Proportion Recovered by Section | 0.150 |  |  |  |  |
|  | 0.000 | 0.000 |  |  |  |
|  | 0.500 | 0.000 | 0.214 |  |  |
|  | 0.350 | 1.000 | 0.714 | 0.556 |  |
|  | 0.000 | 0.000 | 0.071 | 0.444 |  |
| Proportion Recovered by Section into recipient Section | 1.000 |  |  |  |  |
|  | 0.000 | 0.000 |  |  |  |
|  | 0.700 | 0.000 | 0.300 |  |  |
|  | 0.134 | 0.382 | 0.273 | 0.212 |  |
|  | 0.000 | 0.000 | 0.138 | 0.862 |  |

Table G4 Numbers of operculum-tagged Chinook salmon carcasses released and recovered by donor reach within the Hanford Reach, Return Year 2017

|  |  | Donor <br> Section 1 | Donor <br> Section 2 | Donor Section 3 | Donor <br> Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Tagged by Donor Section |  | 290 | 137 | 227 | 327 | 981 |
| Fish Recovered by Recipient Section | 1 | 0 |  |  |  | 0 |
|  | 2 | 2 | 0 |  |  | 2 |
|  | 3 | 3 | 2 | 3 |  | 8 |
|  | 4 | 1 | 2 | 11 | 8 | 22 |
|  | 5 | 1 | 1 | 5 | 3 | 10 |
| P^ Recovered for each Donor Section |  | 0.024 | 0.036 | 0.084 | 0.034 | 0.043 |
| Proportion Recovered by Section | 1 | 0.000 |  |  |  |  |
|  | 2 | 0.286 | 0.000 |  |  |  |
|  | 3 | 0.429 | 0.400 | 0.158 |  |  |
|  | 4 | 0.143 | 0.400 | 0.579 | 0.727 |  |
|  | 5 | 0.143 | 0.200 | 0.263 | 0.273 |  |
| Proportion Recovered by Section | 1 | 0.000 |  |  |  |  |
|  | 2 | 1.000 | 0.000 |  |  |  |
|  | 3 | 0.434 | 0.405 | 0.160 |  |  |
|  | 4 | 0.077 | 0.216 | 0.313 | 0.393 |  |
|  | 5 | 0.163 | 0.228 | 0.299 | 0.310 |  |

## Appendix H

## Carcass bias assessment results

Carcass surveys of Chinook salmon are conducted each fall to characterize fish spawning in the Hanford Reach. However, it is possible that carcasses collected during surveys do not represent the spawning population. There could be carcass collection bias against smaller/younger fish or males in the stream surveys (Zhou 2002; Murdoch et al. 2010; Richards and Pearsons, 2013). If true, this bias may compromise estimates associated with age and gender compositions by origin as well as escapement estimates of hatchery and natural origin fish. We began a pilot project to evaluate potential size and sex recovery bias in 2011. This work has occurred annually with the exception of 2014 when measurements were not taken on the mark sample so a bias estimate could not be estimated.

The methods for collecting, sampling, and releasing tagged carcass associated with this evaluation have varied slightly between years. In general, roughly 1,000 carcasses were collected for demographic data and tagged with numbered plastic tags. Depending on the year, the tagged fish were either systematically released either near shore or mid river or over known active redd locations.

The release strategy for years 2011-2013 included releasing tagged carcass either near shore or mid-channel near the point of initial recovery. Carcasses released near shore had higher proportions of recaptures compared to fish released mid channel. It was not uncommon for carcasses released near shore to be recovered the following day in the same vicinity of their release. In 2015, 2016, and 2017 we released tagged carcasses over active redd locations to better match the natural disposition of post spawn carcasses. After release into the river, the carcasses generally sunk quickly and gradually moved downstream along the bottom in a similar manner to that of post-spawn fish.
The annual recovery rates of tagged carcasses decreased annually from a high of 17.2\% in 2011 to a low of 3.8\% in 2015 (Tables H.1-H.7). The annual recovery rates may be influenced by the release method and by reduced chances of recovering tagged carcasses during large spawning escapements of fall Chinook salmon to the Hanford Reach.
In general, the level of carcass recovery bias was low and varied between years; suggesting that carcass samples collected may be reflective of the spawning population.

Table H. 1 Summary of mark recapture of post-spawn fall Chinook salmon in the Hanford Reach, 2011. Post orbital to hypural plate length (POHL) calculated from linear regression equation for fork length versus known POHL.

|  |  |  | Release Locations |  |  |  |  |  | Total Released |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |
| Released \# |  |  | 500 |  |  | 493 |  |  | 993 |  |
| Recaptured \# |  |  | 110 |  |  | 61 |  |  | 171 |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.220 |  |  | 0.124 |  |  | 0.172 |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | P^ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 26 | 0.026 | 66 | 0.066 | 172 | 0.173 | 137 | 0.138 | 401 | 0.404 |
| Female | 0 | 0.000 | 14 | 0.014 | 331 | 0.333 | 247 | 0.249 | 592 | 0.596 |
| Total | 26 | 0.026 | 80 | 0.081 | 503 | 0.507 | 384 | 0.387 | 993 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 3 | 0.018 | 10 | 0.058 | 34 | 0.199 | 26 | 0.152 | 73 | 0.427 |
| Female | 0 | 0.000 | 2 | 0.012 | 53 | 0.310 | 43 | 0.251 | 98 | 0.573 |
| Total | 3 | 0.018 | 12 | 0.070 | 87 | 0.509 | 69 | 0.404 | 171 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | 0.0 |  |  | . 008 |  | 026 |  | 014 |  | 023 |
| Female | 0.0 |  |  | . 002 |  | 23 |  | 003 |  | . 023 |
| Total | 0.0 |  |  | 010 |  | 002 |  | 017 |  | 000 |

Table H. 2 Summary of mark recapture of post-spawn fall Chinook salmon in the Hanford Reach, 2012. POHL calculated from linear regression equation for fork length versus know POHL.

|  |  |  | Release Locations |  |  |  |  |  | Total Released |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Bank |  |  | Mid Channel |  |  |  |  |
| Released \# |  |  | 489 |  |  | 500 |  |  | 989 |  |
| Recaptured \# |  |  | 103 |  |  | 34 |  |  | 137 |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.211 |  |  | 0.068 |  |  | 0.139 |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 49 | 0.050 | 172 | 0.174 | 157 | 0.159 | 142 | 0.144 | 520 | 0.526 |
| Female | 0 | 0.000 | 31 | 0.031 | 192 | 0.194 | 246 | 0.249 | 469 | 0.474 |
| Total | 49 | 0.050 | 203 | 0.205 | 349 | 0.353 | 388 | 0.392 | 989 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 10 | 0.073 | 25 | 0.182 | 20 | 0.146 | 19 | 0.139 | 74 | 0.540 |
| Female | 0 | 0.000 | 5 | 0.036 | 22 | 0.161 | 36 | 0.263 | 63 | 0.460 |
| Total | 10 | 0.073 | 30 | 0.219 | 42 | 0.307 | 55 | 0.401 | 137 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male |  |  |  |  |  |  | 0.0 |  |  | . 014 |
| Female | 0.0 |  |  |  |  |  | -0.01 |  |  | 014 |
| Total |  |  |  |  |  |  | -0.00 |  |  | 00 |

Table H. 3 Summary of mark recapture of post-spawn fall Chinook salmon in the Hanford Reach, 2013. POHL calculated from linear regression equation for fork length versus know POHL.

|  |  |  | Release Locations |  |  |  |  |  | Total Released |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Bank |  |  | Mid Channel |  |  |  |  |
| Released \# |  |  | 552 |  |  | 521 |  |  | 1,073 |  |
| Recaptured \# |  |  | 69 |  |  | 45 |  |  | 114 |  |
| Recapture P^ |  |  | 0.125 |  |  | 0.086 |  |  | 0.106 |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | $47-58 \mathrm{~cm}$ |  | $59-69 \mathrm{~cm}$ |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 206 | 0.192 | 332 | 0.309 | 183 | 0.170 | 60 | 0.056 | 781 | 0.727 |
| Female | 1 | 0.001 | 55 | 0.051 | 184 | 0.171 | 55 | 0.051 | 295 | 0.274 |
| Total | 206 | 0.192 | 387 | 0.360 | 367 | 0.341 | 115 | 0.107 | 1,075 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 16 | 0.140 | 42 | 0.368 | 19 | 0.167 | 8 | 0.070 | 85 | 0.746 |
| Female | 0 | 0.000 | 5 | 0.044 | 19 | 0.167 | 5 | 0.044 | 29 | 0.254 |
| Total | 16 | 0.140 | 47 | 0.412 | 38 | 0.333 | 13 | 0.114 | 114 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | 0.0 |  |  |  | 0.0 |  | -0.01 |  |  |  |
| Female | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  |
| Total | 0.0 |  |  |  | 0.0 |  | -0.0 |  |  |  |

Table H. 4 Summary of mark recapture of post-spawn fall Chinook salmon in the Hanford Reach, 2015, POHL.

| Total Release in Mid-Channel Redd Locations, RY2015 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Released \# |  |  | 997 |  |  |  |  |  |  |  |
| Recaptured \# |  |  | 38 |  |  |  |  |  |  |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.038 |  |  |  |  |  |  |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 39 | 0.039 | 128 | 0.128 | 183 | 0.184 | 172 | 0.173 | 522 | 0.524 |
| Female | 1 | 0.001 | 37 | 0.037 | 287 | 0.288 | 151 | 0.151 | 476 | 0.477 |
| Total | 39 | 0.039 | 165 | 0.165 | 470 | 0.471 | 323 | 0.324 | 997 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 0 | 0.000 | 6 | 0.158 | 9 | 0.237 | 8 | 0.211 | 23 | 0.605 |
| Female | 0 | 0.000 | 1 | 0.026 | 7 | 0.184 | 7 | 0.184 | 15 | 0.395 |
| Total | 0 | 0.000 | 7 | 0.184 | 16 | 0.421 | 15 | 0.395 | 38 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | 0.0 |  | -0.030 |  | -0.0 |  | -0.0 |  |  |  |
| Female | 0.0 |  | 0.0 |  | 0.1 |  | -0.0 |  |  |  |
| Total | 0.0 |  | -0. |  | 0.0 |  | -0.0 |  |  |  |

Table H. 5 Summary of mark recapture of post-spawn fall Chinook salmon in the Hanford Reach, 2016, POHL.

| Total Release in Mid-Channel Redd Locations, RY2016 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Released \# |  |  | 987 |  |  |  |  |  |  |  |
| Recaptured \# |  |  | 46 |  |  |  |  |  |  |  |
| Recapture P^ |  |  | 0.047 |  |  |  |  |  |  |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | $47-58 \mathrm{~cm}$ |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 43 | 0.044 | 171 | 0.173 | 181 | 0.183 | 119 | 0.121 | 514 | 0.521 |
| Female | 0 | 0.000 | 35 | 0.035 | 334 | 0.338 | 104 | 0.105 | 473 | 0.479 |
| Total | 43 | 0.044 | 206 | 0.209 | 515 | 0.522 | 223 | 0.226 | 987 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 4 | 0.087 | 7 | 0.152 | 11 | 0.239 | 4 | 0.087 | 26 | 0.565 |
| Female | 0 | 0.000 | 1 | 0.022 | 15 | 0.326 | 4 | 0.087 | 20 | 0.435 |
| Total | 4 | 0.087 | 8 | 0.174 | 26 | 0.565 | 8 | 0.174 | 46 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | -0.0 | . 043 |  |  |  |  |  | , 34 |  | . 044 |
| Female | 0.0 | 000 |  |  |  |  |  | 18 |  | . 044 |
| Total | -0.0 | 043 |  |  |  |  |  | 52 |  | 000 |

Table H. 6 Summary of mark recapture of post-spawn fall Chinook salmon in the Hanford Reach, 2017, POHL.

| Total Release in Mid-Channel Redd Locations, RY2017 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Released \# |  |  | 981 |  |  |  |  |  |  |  |
| Recaptured \# |  |  | 42 |  |  |  |  |  |  |  |
| Recapture P^ |  |  | 0.043 |  |  |  |  |  |  |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | $59-69 \mathrm{~cm}$ |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 11 | 0.011 | 173 | 0.176 | 193 | 0.197 | 121 | 0.123 | 498 | 0.508 |
| Female | 0 | 0.000 | 38 | 0.039 | 342 | 0.349 | 103 | 0.105 | 483 | 0.492 |
| Total | 11 | 0.011 | 211 | 0.215 | 535 | 0.545 | 224 | 0.228 | 981 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 0 | 0.000 | 7 | 0.167 | 15 | 0.357 | 4 | 0.095 | 26 | 0.619 |
| Female | 0 | 0.000 | 3 | 0.071 | 11 | 0.262 | 2 | 0.048 | 16 | 0.381 |
| Total | 0 | 0.000 | 10 | 0.238 | 26 | 0.619 | 6 | 0.143 | 42 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male |  | 011 | 0.0 |  |  |  |  | 28 |  | 111 |
| Female |  | 00 |  |  |  |  |  | 57 |  | 111 |
| Total |  | 11 |  |  |  |  |  | 85 |  | 000 |

Table H. 7 Mark and recapture bias post-spawn fall Chinook salmon in the Hanford Reach by size group (POHL), Return Years 2011-2013 and 2014-2017. Bias = $\mathbf{P}^{\wedge}$ Released - $\mathbf{P}^{\wedge}$ Recovered.

| Return |  |  | Post Orbital to Hypural Plate Length Size Groups |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | \# Tagged | \# Recovered | $<\mathbf{4 7} \mathbf{~ c m}$ | $\mathbf{4 7} \mathbf{- 5 8} \mathbf{~ c m}$ | $\mathbf{5 9} \mathbf{- 6 9} \mathbf{~ c m}$ | $\mathbf{> 6 9 \mathbf { c m }}$ |
| $2011^{\mathrm{a}}$ | 993 | 171 | 0.009 | 0.010 | -0.002 | -0.017 |
| $2012^{\mathrm{a}}$ | 989 | 137 | -0.023 | -0.014 | 0.046 | -0.009 |
| $2013^{\mathrm{a}}$ | 1073 | 114 | 0.051 | -0.052 | 0.008 | -0.007 |
| $2015^{\mathrm{b}}$ | 997 | 38 | 0.039 | -0.019 | 0.050 | -0.071 |
| $2016^{\mathrm{b}}$ | 987 | 46 | -0.043 | 0.035 | -0.043 | 0.052 |
| $2017^{\mathrm{b}}$ | 981 | 42 | 0.011 | -0.023 | -0.074 | 0.085 |
| Mean | $\mathbf{1 0 0 3}$ | $\mathbf{9 1}$ | $\mathbf{0 . 0 0 7}$ | $\mathbf{- 0 . 0 1 1}$ | $\mathbf{- 0 . 0 0 3}$ | $\mathbf{0 . 0 0 6}$ |

${ }^{\text {a }}$ Marked fish were released near shore or in mid Channel in roughly equal proportions. Lengths were calculated from linear regression equation for fork length versus known POHL
${ }^{\mathrm{b}}$ Marked fish were released over the top of known redd locations.

## Appendix I <br> Demographic comparisons for double index tag groups released from Priest Rapids Hatchery, Brood Years 2009-2014.

Double Index Tag (DIT) groups of fall Chinook salmon have been released annually from Priest Rapids Hatchery (PRH) starting with the progeny of the 2009 brood. Adipose clipped fish from these DIT groups have been recovered in various mark selective fisheries (MSF) occurring in ocean, marine, and freshwater zones. The Regional Mark Processing Center database was queried to identify mark selective fisheries occurring since 2012 that included recoveries of PRH DIT groups (Table I.1). Detailed descriptions of these fisheries are available at websites maintained by the RMPC, Oregon Department of Fish and Game, and WDFW. The level of contribution to these fisheries, some of which are summer Chinook salmon fisheries, is beyond the scope of this document. D

Survival estimates for DIT groups from release and recovery at PRH was calculated by dividing the total DIT recoveries at PRH for each brood year (ages $1-6$ ) by the corresponding number of juveniles marked for each DIT group. Similarities in gender composition, survival, age at maturity, and size at age between DIT groups within a brood year strongly suggest there is no difference for fish recovered at PRH (Tables I.2, I.3, I.4, and I.5).

Table I. 1 Regional Mark Processing Center mark selective fisheries showing recoveries of Priest Rapids Hatchery origin coded-wire tagged adipose clipped fish from brood years 2009-2014.

| Sampling Agency | Fishery |
| :---: | :--- |
| Alaska Dept. of Fish and Game | Ocean Selective Troll |
| Oregon Dept. of Fish and Game | Ocean Sport |
|  | Columbia River Sport |
|  | Columbia River Test Net |
|  | Columbia River Purse Seine |
|  | Columbia River Gillnet |
| Washington Dept. of Fish and Wildlife | Marine Sport |
|  | Columbia River Sport |

Table I. 2 Gender Composition of DIT groups by brood year. Brood years 2012-2014 not complete.

|  | Males |  | Females |  |
| :---: | :---: | :---: | :---: | :---: |
| Brood Year | Ad-CWT | CWT Only | Ad-CWT | CWT Only |
| 2009 | 0.720 | 0.718 | 0.280 | 0.282 |
| 2010 | 0.540 | 0.546 | 0.460 | 0.454 |
| 2011 | 0.644 | 0.638 | 0.346 | 0.362 |
| 2012 | 0.641 | 0.643 | 0.359 | 0.357 |
| 2013 | 0.650 | 0.652 | 0.350 | 0.348 |
| 2014 | 0.846 | 0.808 | 0.154 | 0.192 |
| Mean | $\mathbf{0 . 6 7 4}$ | $\mathbf{0 . 6 6 8}$ | $\mathbf{0 . 3 2 5}$ | $\mathbf{0 . 3 3 3}$ |

Table I. 3 Smolt to adult return proportion comparisons between DIT Groups by brood year. Brood years 2012-2014 not complete.

| Brood Year | Mark plus CWT | $\mathbf{P}^{\wedge}$ Survival by Age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age 2 | Age 3 | Age 4 | Age 5 | Age 6 | Total |
| 2009 | Ad-Clipped | 0.0004 | 0.0014 | 0.0006 | 0.0003 | 0.0000 | 0.0027 |
|  | No Mark | 0.0004 | 0.0014 | 0.0007 | 0.0002 | 0.0000 | 0.0027 |
| 2010 | Ad-Clipped | 0.0009 | 0.0033 | 0.0052 | 0.0006 | 0.0000 | 0.0100 |
|  | No Mark | 0.0009 | 0.0035 | 0.0050 | 0.0006 | 0.0000 | 0.0100 |
| 2011 | Ad-Clipped | 0.0003 | 0.0021 | 0.0024 | 0.0004 | 0.0000 | 0.0051 |
|  | No Mark | 0.0003 | 0.0023 | 0.0026 | 0.0003 | 0.0000 | 0.0055 |
| 2012 | Ad-Clipped | 0.0015 | 0.0046 | 0.0017 | 0.0003 |  | 0.0082 |
|  | No Mark | 0.0017 | 0.0046 | 0.0017 | 0.0003 |  | 0.0082 |
| 2013 | Ad-Clipped | 0.0005 | 0.0014 | 0.0012 |  |  | 0.0031 |
|  | No Mark | 0.0004 | 0.0016 | 0.0013 |  |  | 0.0034 |
| 2014 | Ad-Clipped | 0.0001 | 0.0004 |  |  |  | 0.0005 |
|  | No Mark | 0.0001 | 0.0004 |  |  |  | 0.0005 |
| Mean | Ad-Clipped | 0.0006 | 0.0022 | 0.0022 | 0.0003 | 0.0000 | 0.0049 |
|  | No Mark | 0.0006 | 0.0023 | 0.0019 | 0.0002 | 0.0000 | 0.0050 |

Table I. 4 Age composition of DIT Groups by brood year. Brood years 2012-2014 not complete.

| Brood Year | DIT Group | N | Age Composition (Genders Combined) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2009 | Ad-CWT | 1,648 | 0.137 | 0.520 | 0.244 | 0.099 | 0.000 |
|  | CWT Only | 2,787 | 0.145 | 0.526 | 0.242 | 0.088 | 0.000 |
| 2010 | Ad-CWT | 6,017 | 0.086 | 0.334 | 0.522 | 0.057 | 0.001 |
|  | CWT Only | 11,087 | 0.089 | 0.346 | 0.504 | 0.060 | 0.001 |
| 2011 | Ad-CWT | 3,021 | 0.054 | 0.406 | 0.469 | 0.070 | 0.000 |
|  | CWT Only | 3,316 | 0.057 | 0.413 | 0.474 | 0.056 | 0.000 |
| 2012 | Ad-CWT | 4,947 | 0.183 | 0.565 | 0.213 | 0.039 |  |
|  | CWT Only | 5,505 | 0.183 | 0.570 | 0.209 | 0.038 |  |
| 2013 | Ad-CWT | 1,857 | 0.150 | 0.454 | 0.396 |  |  |
|  | CWT Only | 2,023 | 0.131 | 0.476 | 0.393 |  |  |
| 2014 | Ad-CWT | 280 | 0.143 | 0.857 |  |  |  |
|  | CWT Only | 281 | 0.135 | 0.865 |  |  |  |
| Mean | Ad-Clipped | N/A | 0.125 | 0.523 | 0.369 | 0.066 | 0.001 |
|  | No Mark | N/A | 0.123 | 0.533 | 0.364 | 0.060 | 0.001 |

Table I. 5 Size at age for DIT Groups by brood year. Brood years 2012-2014 not complete.

| Brood Year | DIT Group | Fall Chinook fork length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2009 | AD-CWT | 226 | 49 | 4 | 857 | 67 | 5 | 402 | 78 | 5 | 163 | 85 | 5 | 0 | 0 | 0 |
|  | CWT Only | 404 | 48 | 4 | 1,465 | 66 | 5 | 674 | 77 | 6 | 244 | 84 | 6 | 0 | 0 | 0 |
| 2010 | AD-CWT | 519 | 48 | 4 | 2,011 | 68 | 4 | 3,138 | 77 | 5 | 340 | 81 | 5 | 9 | 89 | 5 |
|  | CWT Only | 985 | 48 | 4 | 3,840 | 68 | 5 | 5,585 | 77 | 5 | 663 | 82 | 5 | 14 | 81 | 6 |
| 2011 | AD-CWT | 162 | 47 | 4 | 1,227 | 66 | 5 | 1,417 | 76 | 5 | 210 | 82 | 6 | 5 | 84 | 2 |
|  | CWT Only | 188 | 47 | 4 | 1,369 | 66 | 5 | 1,571 | 77 | 5 | 185 | 82 | 6 | 3 | 85 | 4 |
| 2012 | AD-CWT | 904 | 59 | 5 | 2,794 | 67 | 5 | 1,055 | 78 | 5 | 194 | 82 | 5 |  |  |  |
|  | CWT Only | 1,006 | 50 | 5 | 3,139 | 67 | 5 | 1,151 | 78 | 5 | 209 | 81 | 6 |  |  |  |
| 2013 | AD-CWT | 279 | 45 | 5 | 843 | 66 | 5 | 735 | 75 | 5 |  |  |  |  |  |  |
|  | CWT Only | 266 | 45 | 5 | 962 | 66 | 5 | 795 | 75 | 5 |  |  |  |  |  |  |
| 2014 | AD-CWT | 40 | 49 | 3 | 240 | 66 | 5 |  |  |  |  |  |  |  |  |  |
|  | CWT Only | 38 | 50 | 4 | 243 | 66 | 5 |  |  |  |  |  |  |  |  |  |
| Mean | AD-CWT | 355 | 50 | 4 | 1,329 | 67 | 5 | 1,652 | 77 | 5 | 238 | 83 | 5 | 8 | 87 | 4 |
|  | CWT Only | 481 | 48 | 4 | 1,836 | 67 | 5 | 2,610 | 77 | 5 | 364 | 83 | 6 | 8 | 83 | 5 |

## Appendix J <br> Alternative pNOB and PNI Estimates

An alternative pNOB was developed to account for the genetic influence on pNOB resulting from the PRH spawning protocol of spawning one male with one, two, or four females. It is intended to represent actual gene flow to the progeny instead of strictly the origin and number of parents. However, it should be noted that although PNI was intended to index gene flow, the alternative method of estimating pNOB as described below has not been used elsewhere and is currently undergoing review. The PNI calculation using the alternative pNOB method is PNI = Alt pNOB/(Alt pNOB + pHOS)

The alternative pNOB is calculated by assigning scores to the estimated matings of males and females based on origin during the spawning of the PRH broodstock.

The hatchery x hatchery matings $=0.0$ points,
Hatchery x natural matings $=0.5$ points, and
Natural x natural matings $=1.0$ points.
The scores of all of the matings were averaged to generate the overall alternative pNOB. For example, the alternative pNOB calculation for the mating of one natural origin male x two hatchery origin females is $(0.5+0.5) / 2$ females $)=0.5$, whereas the conventional pNOB calculation for this mating equals (1 natural / (1 natural +2 hatchery) $=0.33$.

The origin assignments of fish spawned were based on a combination of otolith marks, adipose clips, and CWTs as done for the conventional pNOB calculation previously discussed. The fish from the OLAFT and ABC were spawned with either fish from those collections or adipose intact broodstock fish from the PRH volunteer trap to try to reduce the number of matings with hatchery origin fish; hence, improving the pNOB for the GCPUD program.

Similar to that done for estimates of pNOB by program, alternative pNOB and PNI estimates are given for the PRH facility as a whole and specific to the GCPUD production associated with each brood year. The pHOS used for these estimates are given in Error! Reference source not found..

The conventional and alternative pNOB values for GCPUD production spawned at PRH and GCPUD associated pHOS are presented in Error! Reference source not found. K.1. Both methods of calculating PNI associated with the GCPUD production provide PNI values in excess of the stated PNI target of 0.67 for most years.

Table J. 1 Conventional and alternative calculations of pNOB and PNI associated with the production specific to Grant County PUD, Return Years 2012-2017.

| Return Year | GCPUD Broodstock Combined | GCPUD pHOS ${ }^{\mathbf{1}}$ | PNI |
| :---: | :---: | :---: | :---: |
|  | 0.182 | 0.068 | 0.729 |
|  | 0.225 | 0.151 | 0.598 |
| 2014 | 0.343 | 0.039 | 0.898 |
| 2015 | 0.313 | 0.057 | 0.846 |
| 2016 | 0.260 | 0.049 | 0.841 |
| 2017 | 0.433 | 0.065 | 0.869 |
|  | Alternative pNOB = Total Score / Total Matings |  |  |
| Return Year | GCPUD Broodstock | GCPUD pHOS ${ }^{\mathbf{1}}$ | PNI |
| 2012 | 0.197 | 0.068 | 0.744 |
| 2013 | 0.284 | 0.151 | 0.653 |
| 2014 | 0.423 | 0.039 | 0.916 |
| 2015 | 0.434 | 0.057 | 0.884 |
| 2016 | 0.356 | 0.049 | 0.879 |
| 2017 | 0.473 | 0.065 | 0.879 |

${ }^{1}$ The proportion of the pHOS specific to the GCPUD mitigation smolt releases from PRH.

## Appendix K

## Explanation of methods for calculating adult-to-adult expansions based on coded-wire tag recoveries at Priest Rapids Hatchery

Expanding adult coded wire tag (CWT) recoveries of either PRH or RSH origin fish by the corresponding brood’s juvenile CWT rates has historically resulted in an under estimate of adult returns to locations within the Hanford Reach for each brood. A variety of factors may contribute to this problem; however, inappropriate juveniles tag expansion rates resulting from nonrepresentative placement of tag groups within the general population is likely the greatest contributing factor. For many years, WDFW fish management staff have addressed the issues related to problematic juvenile tag rates by employing adult-to-adult CWT expansions for the PRH origin returns to PRH for run-reconstruction associated with their annual fall Chinook salmon forecast. We used similar methods to expand PRH and RSH origin adult CWT recoveries in the vicinity of Hanford Reach to calculate PNI. An example of the calculations for the adult-to-adult expansion for the 2010 brood during return year 2014 is provided below. We make the assumption that the total number of PRH origin returns to PRH can be determined by removing other hatchery fish from the return: this is done by expanding the few other hatchery CWT recoveries by their corresponding juvenile CWT rates. Other hatchery CWT groups often have tag rates exceeding 50\%; therefore, we assume juvenile tag rate expansions are representative for these groups. In addition, we make the assumption that very few natural origin fish return to PRH.

Adult-to-Adult Expansion BY2010 $=$ Total BY2010 CWT Recoveries at PRH
Total By2010 PRH Origin Returns to PRH
Adult-to-Adult Expansion BY2010 $=\quad \underline{8,719}=0.211$
41,348
We then use the Adult-to-Adult Expansion by2010 to expand all recoveries of PRH by2010 in the Hanford Reach stream survey for return year 2014. This method is duplicated for each brood present in the given return year for both PRH and RSH to determine the total number of PRH and RSH origin fish in the escapement. The estimated number of PRH origin fish in the RY2014 Hanford Reach escapement based on the adult-to-adult expansion is higher than the number calculated using the conventional juvenile tag rate (Table 1).

Table K. 1 The number of Priest Rapids Hatchery origin fish in the RY 2014 Hanford Reach escapement calculated from Adult-to-Adult Expansions versus Juvenile Tag Rates.

| BY | CWT Recovered | Adult-to- <br> Adult Exp | Expanded <br> CWT | Survey Sample <br> Rate | Total PRH origin in <br> Escapement |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 5 | 0.216 | 23 | 0.1063 | 218 |
| 2010 | 139 | 0.211 | 659 | 0.1063 | 6,197 |
| 2011 | 18 | 0.127 | 142 | 0.1063 | 1,333 |
| 2012 | 5 | 0.160 | 31 | 0.019 | 1,645 |
| Adult-to-Adult Exp estimate for PRH origin fish in the Hanford Reach Escapement |  |  |  |  |  |


[^0]:    © 2018, PUBLIC UTILITY DISTRICT NO. 2 OF GRANT COUNTY, WASHINGTON.
    ALL RIGHTS RESERVED UNDER U.S. AND FOREIGN LAW, TREATIES AND CONVENTIONS.

[^1]:    ${ }^{1}$ Origin determined from "in-sample" otoliths, adipose clips and/or coded-wire tags.

[^2]:    ${ }^{1}$ The origin is assigned by survey
    ${ }^{\text {a }}$ Does not include age- 6 returns

[^3]:    ${ }^{\text {a }}$ Does not include age-6 returns

[^4]:    ${ }^{1}$ Includes both adult males and jacks.
    ${ }^{\text {a }}$ Includes age-2 through 5.

[^5]:    ${ }^{1}$ Target stream includes the Columbia River between McNary and Wanapum dams as well as the Yakima River below Prosser Dam.

[^6]:    ${ }^{1}$ Includes fish from other hatcheries based on presence of a coded-wire tag or adipose clip fish without an otolith mark.

[^7]:    ${ }^{\text {a }}$ Includes 1,724 fish made available to the Yakama Nation and Umatilla Tribe for broodstock to support their fall Chinook salmon programs.

