# Priest Rapids Hatchery Monitoring and Evaluation Annual Report for 2016-2017 

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## Disclaimer

This report is provided as an annual data update of the Public Utility District No. 2 of Grant County, Washington's (Grant PUD's) monitoring and evaluation plan for Priest Rapids Hatchery. All data are provisional and subject to change as new data and analyses become available. Readers are cautioned to use data at their own risk and should consult the most current report to obtain the most current and accurate information. Data sets will become final when they are published in peer reviewed scientific journals.

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## Executive Summary

This report is the seventh annual report dedicated to monitoring and evaluating the Priest Rapids Hatchery (PRH) production of fall Chinook salmon. The PRH is located below Priest Rapids Dam adjacent to the Columbia River and has been in operation since 1963. The monitoring and evaluation program associated with PRH is intended to evaluate the performance of the program in meeting hatchery and natural production goals. This report is intended to be cumulative, but also focus attention on the most recent year of data collection and production (2016-2017).
The PRH was originally built to mitigate for the construction and operation of Priest Rapids and Wanapum dams. The hatchery is operated as an integrated program for the purpose of increasing harvest while limiting undesirable risks to the naturally spawning population. The hatchery produces 5.6 million subyearling fall Chinook salmon for Public Utility District No. 2 of Grant County, Washington's (GPUD) mitigation requirement and 1.7 million subyearling fall Chinook salmon under contract with the United States Army Corps of Engineers for mitigation for the construction and operation of John Day Dam. These fish contribute significantly to a variety of fisheries, such as fisheries off the coasts of Alaska and Canada and fisheries in the Columbia River.

The estimated total escapement of fall Chinook salmon to the Hanford Reach in 2016 was 116,388 fish. This is substantially higher than mean historic abundances. The historical mean and median escapement for 1991 through 2016 is 73,550 and 55,208 fish, respectively.
The 2016 returns to PRH volunteer trap totaled 28,786 fall Chinook salmon. A total of 4,324 fish that returned to the volunteer trap at PRH were ponded at the hatchery for broodstock. An additional 247 fish were ponded from the Angler Broodstock Collection (ABC) fishery and 366 fish were ponded from Priest Rapids Dam Off-Ladder-Adult-Fish-Trap (OLAFT) in an effort to increase the number of natural-origin broodstock. In total, 4,938 fish were spawned to meet eggtake goals for multiple hatchery programs. Most of the fish that were surplus to broodstock needs were provided to food-banks and tribes for consumption.

There were a number of similarities and differences of hatchery and natural origin fall Chinook salmon. The hatchery origin fish appeared to return at a younger age than natural origin fish. It appears that age- 2 and 3 hatchery origin fish tend to be larger than natural origin fish of the same age. Likewise, age-4 and 5 natural origin fish tend to be larger than their hatchery origin counterparts The number of eggs, egg size, and egg mass produced by hatchery and natural origin females of similar length was similar. With the exception of one year, egg retention in female carcasses in the Hanford Reach has been low.

Hatchery origin fish released from PRH spawned throughout the Hanford Reach. In addition, the hatchery origin proportions of spawners relative to total spawners in the different sections of the Hanford Reach were similar. Recent evidence suggested that adult carcasses drift downstream of their spawning location and bias the estimated spawning distribution downstream. Stray rates into other populations appeared to be low based upon coded-wire tag recoveries and PIT tag detections of PRH adults in the Snake River were also low. However, there have been notable numbers of PIT tag detections of PRH adults above Priest Rapids Dam.

The PRH continued to contribute substantially to ocean and Columbia River fisheries and to have higher adult recruitment rates than the natural spawning fall Chinook salmon in the Hanford Reach of the Columbia River. Adult recruitment rate of brood year 2010 for PRH was the
highest that has been observed (61.69) for this program and was substantially higher than the fish spawning in the Hanford Reach (6.21).

PRH origin fish were estimated to make up $6.6 \%$ of the natural spawning population in the Hanford Reach during 2016. All hatchery fish combined (including fish released from Ringold Hatchery and strays from outside the Hanford Reach) comprised 11.5\% of the fall Chinook salmon on the spawning grounds. Otolith recoveries at the PRH volunteer trap indicated that a very high percentage of fish returning to the PRH were of PRH origin. The proportion of natural influence (PNI) for Hanford Reach fall Chinook salmon including all hatcheries was 0.70 in 2016. This value was calculated using a gene flow model based on the Ford model and exceeded the PNI target of 0.67 for the third consecutive year. Adult management of fish at the PRH volunteer trap and alternative broodstock collection techniques to increase natural origin fish in the broodstock have contributed to improvements in PNI for the PRH program.

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### 1.0 Introduction

The Public Utility District No. 2 of Grant County, Washington (Grant PUD) produces and releases 5.6 million subyearling fall Chinook salmon smolts from Priest Rapids Hatchery (PRH) as part of its mitigation for the construction and operation of Priest Rapids and Wanapum dams. The mitigation is the result of three components 1 ) inundation of historic spawning habitat ( 5 million), annual losses of fish that migrate through the project $(325,543)$, and flow fluctuation impacts in the Hanford Reach $(273,961)$. The PRH is located on the east bank of the Columbia River immediately downstream of Priest Rapids Dam (Figure 1 and Figure 2). The Washington Department of Fish \& Wildlife (WDFW) operates PRH which is owned, maintained, and funded by the Grant PUD. This report describes the monitoring and evaluation of the PRH M\&E program.

PRH also produces fish for other programs. PRH produces and releases 1.7 million subyearling smolts on-site for the U.S. Army Corps of Engineers (USACE) John Day Mitigation. An additional 3.7 million eyed eggs are provided for the USACE John Day Mitigation released at Ringold Springs Hatchery (RSH). The eggs for the RSH program are first transferred to Bonneville Hatchery for marking and ultimately $\sim 3.5$ million subyearlings are transported to, acclimated, and released as subyearling smolts from RSH. In previous years, PRH has accommodated egg-takes and/or incubated eggs for the Yakama Nation (YN) upper river bright (URB) fall Chinook salmon releases in the lower Yakima River at their Prosser facility. Additional eggs have also been taken for other programs such as Umatilla Hatchery, WDFW's Salmon in the Classroom program, and to support various research projects.

A cooperative effort between Grant, Douglas, and Chelan County Public Utility Districts and Washington Department of Fish and Wildlife (WDFW) has resulted in an updated Monitoring and Evaluation Plan for PUD Hatchery Programs (Hillman et al. 2013). This document provides guiding principles and approaches for the monitoring and evaluation (M\&E) of PRH. Objectives, hypotheses, measured and derived variables, and field methods that will be used to collect data are listed in this document.

This report of the PRH M\&E program is the seventh annual report (Hoffarth and Pearsons 2012a, 2012b, Richards et al. 2013, Richards and Pearsons 2014, 2015, and 2016) and encompasses data collected during the Washington State fiscal year (FY) 2016-2017 as well as earlier years where data were available. The data presented in this report are preliminary and subject to change as new data and analyses become available. Readers are encouraged to consult the most recent annual report in order to obtain the most current and accurate information.


Figure 1 Location of Priest Rapids and Ringold Spring hatcheries and the Hanford Reach (indicated by stars).


Figure 2 Priest Rapids Hatchery facility and Priest Rapids Dam Off Ladder Adult Fish Trap (OLAFT).

### 2.0 Objectives

The objective of the PRH M\&E plan is to evaluate the performance of the PRH program relative to the goals and objectives of the PRH program. The overarching goal of the PRH program is to meet Grant PUD's hatchery mitigation by producing fish for harvest while keeping genetic and ecological impacts within acceptable limits. The M\&E objectives of the PRH program are described below.

Objective 1: Determine if conservation programs have increased the number of naturally spawning and naturally produced adults of the target population and if the program has reduced the natural replacement rate (NRR) of the supplemented population.
Objective 2: Determine if the proportion of hatchery fish on the spawning ground affects the freshwater productivity of supplemented stocks.

Objective 3: Determine if the hatchery adult-to-adult survival (i.e., hatchery replacement rate, $H R R$ ) is greater than the natural adult-to-adult survival (i.e., natural replacement rate, NRR) and the target hatchery survival rate.
Objective 4: Determine if the proportion of hatchery origin spawners ( pHOS or PNI) is meeting management targets.

Objective 5: Determine if the run timing, spawn timing, and spawning distribution of the hatchery component is similar to the natural component of the target population or is meeting programs-specific objectives.
Objective 6: Determine if stray rate of hatchery fish is below the acceptable levels to maintain genetic variation among stocks.
Objective 7: Determine if genetic diversity, population structure, and effective population size have changed in natural spawning populations as a result of the hatchery program.
Objective 8: Determine if hatchery programs have caused changes in phenotypic characteristics of natural populations.
Objective 9: Determine if hatchery fish were released at programmed size and number.
Objective 10: Determine if appropriate harvest rates have been applied to the conservation, safety-net, and segregated harvest programs to meet the HCP/SSSA goal of provided harvest opportunities while also contributing to population management and minimalizing risk to natural populations.

We also present information in this report about two regional objectives that relate to disease and ecological interactions.

### 3.0 Project Coordination

WDFW M\&E staff partially assigned to PRH also conducts similar work at RSH. The M\&E staff also works in conjunction with multiple WDFW groups that include PRH fish culture staff, the Columbia River Coded-Wire Tag Recovery Program (CRCWTRP), Region 3 Fish Management staff, the Supplementation Research Team in Wenatchee, and the Grant PUD biological science staff to complete many of the tasks included in the M\&E Plan. In addition, samples collected at the hatchery and in the field were transported to and analyzed by WDFW laboratories including
the WFDW Scale Reading Lab and the WFW Otolith Lab. Coded-wire tags (CWT) were processed by the M\&E staff either at the WDFW District 4 office or the PRH wet lab. Data and analyses collected in association with the PRH M\&E and Hanford Reach population monitoring are incorporated into the WDFW Traps, Weirs, and Surveys (TWS) database which is administered by the WDFW staff stationed in the Region 5 Headquarters in Vancouver. Agency managers use these data for forecasting and managing fall Chinook salmon populations in the Columbia and Snake rivers and tributaries. WDFW and Grant PUD secured and held all environmental permits necessary for the work.

### 4.0 Life History - Hanford Reach Fall Chinook Salmon

The Hanford Reach is one of the last non-impounded reaches of the Columbia River and the location of the largest and most productive natural spawning fall Chinook salmon population in the United States (Harnish et al. 2012, Langshaw et al. 2015, Harnish 2017). The Hanford Reach extends 51 miles from the city of Richland to the base of Priest Rapids Dam. Natural origin fall Chinook salmon emerge from the substrate in the spring and rear in the Hanford Reach until outmigration in the summer. Egg-to-fry survival has been estimated to be about $71 \%$ in the Hanford Reach(Oldenburg et al. 2012) and egg-to-pre-smolt survival has been estimated to be about $40.2 \%$ (Harnish et al. 2012) or more recently at $63.4 \%$ (Harnish, 2017). Both of these estimates are high when compared to other Chinook salmon populations and flow management has resulted in improvements in survival (Harnish et al. 2012, Harnish 2017). The age at maturity for naturally produced fish in the Hanford Reach varies between age-1 mini-jack and age-6 adults: albeit recoveries of age- 1 and 6 fish are generally rare. The age of fish reported in this document begins with the first birthday occurring the year after the parents spawned. The abundance of mini-jacks which mature as age-1 males is currently not known. Age-2 male fall Chinook salmon (a.k.a jacks) return to the Hanford Reach after spending roughly one year in the ocean. The majority of the natural origin adults return after having spent three to four years in the ocean (age-4 and 5). A small portion, typically less than $2 \%$, will spend up to five years in the ocean and return as age-6. The ocean distribution of natural and hatchery origin Hanford Reach upriver brights are similar and range from the northern California coast to the Gulf of Alaska (Norris et al, 2000, Weitkamp, 2010). The majority of the adults migrate north of the Columbia River with the harvest primarily occurring in non-selective ocean and freshwater fisheries (Norris et al, 2000).

### 5.0 Sample Size Considerations

We attempted to strike an appropriate balance between statistical precision, logistics, and financial investment when setting sample size targets. A phased approach was used to collect biological samples with sufficient accuracy and precision. In general, we attempted to oversample the raw samples such as carcasses and trap recoveries and then use post season analysis to determine if sub-sampling was appropriate. The sample size target of systematic field sampling is 2,500 of the carcasses in the Hanford Reach, 1,000 at the hatchery trap, and 1,000 of the hatchery volunteer broodstock, and 200 broodstock collected from each other source such as OLAFT and ABC fishery.

All adult fall Chinook salmon recovered at PRH, in the Hanford Reach sport fishery, and in the stream surveys were sampled for the presence of CWT to maximize the precision of estimates generated from these data. Representative otolith samples by survey type were randomly selected as a sub-sample for processing to estimate origin by age class if numbers allowed. In some cases,
all otolith samples for a survey type were processed if the sampling rate provided relatively low numbers of otoliths collected or if there was a need for higher precision or accuracy. During return year 2016, randomly selected sub-samples of otoliths collected from the PRH volunteer trap and volunteer broodstock were submitted for processing. The methodologies for selecting otolith sub-samples have differed between return years as field methods changed and as new analyses facilitated improvements in approaches. In general, we randomly selected otoliths from various survey types to obtain roughly 120 otoliths for each age and gender. In some cases, all otoliths were submitted for stratified groups (age/gender) when specific age classes contain less than 100 samples. For example, typically all samples of age- 5 and 6 fish were submitted because of the low number of fish represented in the field collected sample. The stratified sub-sample size refinement process is described in Richards and Pearsons (2014). The sub-sample groups often included fish possessing a CWT within the biological sample which increased the number of fish sampled for origin with no additional cost.

### 6.0 Current Operation at Priest Rapids Hatchery

In 2016, 29,564 adult fall Chinook salmon were handled at PRH (Table 1). The 2016 broodstock for PRH were collected at the hatchery volunteer trap, the OLAFT, and from the ABC fishery. The majority of the broodstock were collected from the PRH volunteer trap which was operated from September 12 through December 12, 2016. The last fish were hauled from the volunteer trap on December 5 with no fish entering the trap afterwards. In attempt to increase pNOB , the broodstock ponded generally excluded known hatchery fish (i.e., possessing an adipose clip and or CWT); hence, increasing the potential number of natural origin broodstock ponded. Hatchery staff ponded 2,243 known hatchery fish in the event that there were not enough adipose fin intact/non CWT broodstock available to meet program goals. Many of these known hatchery origin fish ponded were surplused as they were replaced by adipose fin intact/non CWT fish during subsequent trapping and ponding operations.

Table 1 Source and disposition of Chinook salmon collected for broodstock at Priest Rapids Hatchery, Return Year 2016.

| Collection Location | Gender | Collected | Trap Surplused | Trap Mortalities | Ponded | Spawned ${ }^{1}$ | Pond Surplused | Pond Mortalities |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Males | 14,468 | 11,850 | 58 | 2,403 | 1,328 | 876 | 356 |
|  | Females | 13,436 | 8,561 | 121 | 4,443 | 2,997 | 1,407 | 350 |
|  | Jacks | 882 | 854 | 28 | 0 | 0 | 0 | 0 |
|  | Total | 28,786 | 21,265 | 207 | 6,846 | 4,325 | 2,283 | 706 |
| OLAFT <br> (Sept 14-Nov 07) | Males | 134 | 0 | 0 | 134 | 113 | 14 | 7 |
|  | Females | 310 | 0 | 0 | 310 | 253 | 3 | 54 |
|  | Jacks | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Total | 444 | 0 | 0 | 444 | 366 | 17 | 61 |
| $\begin{gathered} \mathbf{A B C} \\ (\text { Oct } 28,29 \& 30) \end{gathered}$ | Males | 132 | 0 | 0 | 132 | 96 | 10 | 26 |
|  | Females | 202 | 0 | 0 | 202 | 151 | 7 | 44 |
|  | Jacks | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  | Total | 334 | 0 | 0 | 334 | 247 | 17 | 70 |
| Facility | Total | 29,564 | 21,265 | 207 | 7,624 | 4,938 | 2,317 | 837 |

[^0]The pattern of arrival timing by week for adult fall Chinook salmon to the PRH volunteer trap was determined to help schedule future sampling and broodstock activities. Regular trap operations during 2016 should have provided unimpeded access to the trap. The return time in 2016 suggests it was a single modal peak return similar to patterns in 2015 (Figure 3). The annual return time prior to 2015 was typically a bi-modal peak in which a large proportion of fish returned in late September and late October.


Figure 3 Weekly counts of fish adult Chinook salmon collected at the Priest Rapids Hatchery volunteer trap, 2016.

PRH has four adult salmon holding ponds. Ponds 1 and 2 were used to hold broodstock collected at the PRH Volunteer Trap. Pond 4 was used to hold broodstock collected from the ABC and OLAFT. Pond 3 was used on occasion to temporarily hold males collected from ABC and OLAFT. Several hundred adipose clipped adults were held in Pond 4 to facilitate hatchery- xnatural origin crosses during spawning. The PRH staff generally transported fish from the volunteer trap seven days per week to collect broodstock and or to surplus the excess fish. Male fall Chinook salmon, both adult and jack, typically comprised the majority of the fish surplused from the trap. Spawning days generally occurred on Mondays and Tuesdays each week from October 24 through December $5(\mathrm{~N}=10)$. The hatchery staff generally used the electroanesthesia system for spawning fish. However, on November 8, they used a seine to collect fish out of pond 2 for spawning in order to expedite the day's operation.
The egg-take goal from the 2016 PRH brood was $13,329,318$ eggs. The actual egg-take from the 2016 broodstock was $12,411,530$ ( $93 \%$ of the goal). During routine spawn days, the eggs from two females were stripped into a five-gallon bucket and then the milt from a single male was mixed with the eggs. Two buckets of fertilized eggs were then combined to ensure fertilization. Fertilized eggs were then transferred to the incubation room, combined with multiple egg-takes, weighed to estimate numbers of eggs, and then placed in vertical incubation trays at roughly 7,000 eggs per tray.
Since 2014, a cooperative effort between WDFW and Grant PUD staff to perform real-time otolith reading (RTOR) coinciding with an alternative mating strategy occurred on November 7
and 8. This activity entailed examining 204 otoliths from unmarked males collected at OLAFT or the ABC fishery to identify 151 natural origin fish. These 151 males were used for $1 x 4$ matings with unmarked females collected either at the volunteer trap, OLAFT, or the ABC fishery. Milt from natural origin males was mixed with 4 females in a five-gallon bucket and then two, 5 gallon buckets of eggs were combined before being transferred to the incubation room. An estimated 2,220,002 green eggs were taken during the RTOR 1:4 crosses.

After shipping two large groups of eyed eggs to other facilities, fry from nine egg-takes were moved from the vertical trays in the incubation building to outdoor raceways between January 28 and March 14, 2017. The fry were reared in the raceways until they were of sufficient size that a portion of them could be marked in some manner (i.e., adipose clipped and or tagged). Fish receiving marks and or tags were collected directly from the raceways banks and then released into the corresponding concrete rearing ponds (e.g., fish moved from raceway bank E to channel pond E). Fish not selected for marking were transferred from the raceway banks into the corresponding rearing ponds. The growth of smolts from ponds E and D was accelerated for early releases that occurred on May 23 and 25, respectively. The remaining smolts were released between June 9 and June 19. All releases occurred at night. These fish migrate down the old one mile long spawning channel and then down the hatchery discharge channel to the Columbia River.

### 7.0 Origin of Adult Returns to Priest Rapids Hatchery

There were three sources for collection of adult Chinook salmon broodstock for PRH during the 2016 return: PRH volunteer trap, OLAFT, and ABC. The origin of fish collected at these locations was determined by examination of hatchery marks (i.e., otolith marks, adipose clips, and CWTs) for the fish within the demographic sample groups. PRH origin fish were identified by their otolith mark or a CWT. The fish that did not possess a thermal mark or other hatchery marks and tags were classified as natural origin. Historically, the very low recovery ( $<1 \%$ ) of CWT strays at PRH suggests that a high percentage of the fish not possessing any type of hatchery mark may be of natural origin (See Section 9.0). In some sections of the report, we make a simplifying assumption that fish without hatchery marks are of natural origin. Similar to that observed in previous years, there is a discrepancy between estimates of origin based on CWT and those based on otoliths. It's believed that estimates of origin based on otolith sampling provides the most accurate data under the current marking regime at PRH due to discrepancies in the data associated with CWT results. The error rate associated with determination of origin by otoliths is reported at less than $1 \%$ (J. Grimm, WDFW Otolith Lab, personal communication). Each otolith is independently read by two experienced lab staff. Upon completion of the second read, any discrepancies are read a third time to resolve the conflict. If the marks are poor quality, three staff independently read the otoliths. The otolith marks created by the PRH fish culture staff are high quality and generally require only two readings. Most discrepancies related to otolith data are clerical in nature (data entry). Discrepancies associated with the data collected by the $M \& E$ team were generally clerical and easy to resolve and correct.
We present estimates of abundance based on CWTs (1:1 sample rate) and estimates based on sub-samples of hatchery marked fish collected from specific groups (varying sample rates) to illustrate differences in the estimates for the proportions of natural and hatchery origin fish recovered at PRH as well as the potential for creating a method to correct the historical database that was generated using CWT recoveries.

### 7.1 Origin Based on Hatchery Marks

For return year 2016, the proportion of broodstock obtained from the PRH volunteer trap that was natural origin is estimated at 0.059 . Overall, it is estimated that 0.028 of the fish surplused from the PRH volunteer trap were natural origin. The proportion of natural origin fish used as broodstock from the OLAFT and ABC was estimated to be 0.730 and 0.955 , respectively. The estimated numbers of natural and hatchery origin broodstock spawned annually since return year 2013 are given in (Table 2).

For return years 2014 through 2016, a minimum fork-length threshold of $\sim 73 \mathrm{~cm}$ was generally used to reduce the number of age- 2 and 3 broodstock collected at OLAFT along with the exclusion of hatchery marks and tags. Historical data suggests that a larger proportion of age- 2 and 3 fall Chinook salmon returning to the Hanford Reach are of hatchery origin versus age-4 and 5 fish. Similarly, broodstock collected from the PRH volunteer trap were generally culled to exclude fish $<73 \mathrm{~cm}$ with no hatchery marks or tags.

Table 2 Total fish handled, numbers sampled, and estimates of hatchery and natural origin Chinook salmon collected at Priest Rapids Hatchery, Priest Rapids Dam Off-Ladder Adult Fish Trap, and Angler Broodstock Collection fishery. Origin determined by otolith thermal marks, presence of coded-wire tages, and/or adipose clips, Brood Years 2013-2016.

| Priest Rapids Hatchery Broodstock ${ }^{1}$ |  |  | Estimate (95\% CI) |  |
| :---: | :---: | :---: | :---: | :---: |
| Brood Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| 2013 | 4,476 | 503 | 4,395 [4,319, 4,436] | 81 [40, 157] |
| 2014 | 4,427 | 574 | 4,228 [4,130, 4,294] | 199 [133, 297] |
| 2015 | 4,875 | 682 | 4,482 [4,368, 4,573] | 393 [302, 507] |
| 2016 | 4,324 | 827 | 4,067 [4,034, 4,095] | 257 [227, 290] |
| Priest Rapids Hatchery Surplused from Trap |  |  | Estimate (95\% CI) |  |
| Brood Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| $2013{ }^{\text {a }}$ | 37,355 | 608 | 36,085 [35,375, 36,533] | 1,270 [822, 1,980] |
| $2014{ }^{\text {b }}$ | 73,352 | 639 | 69,024 [67,484, 70,271] | 4,328 [3,081, 5,868] |
| $2015{ }^{\text {b }}$ | 57,625 | 619 | 54,646 [53,418, 55,551] | 2,979 [2,075, 4,207] |
| $2016{ }^{\text {a }}$ | 24,461 | 1,033 | 23,790 [23,737, 23,837] | 668 [619, 719] |
| Off Ladder Fish Trap Broodstock ${ }^{1}$ |  |  | Estimate (95\% CI) |  |
| Brood Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| 2013 | 763 | 169 | 343 [242, 370] | 420 [392, 416] |
| 2014 | 825 | 225 | 143 [122, 166] | 682 [659, 703] |
| 2015 | 348 | 164 | 45 [29, 66] | 303 [282, 319] |
| 2016 | 366 | 211 | 99 [83, 117] | 267 [249, 283] |
| Angler Broodstock Collection Broodstock |  |  | Estimate (95\% CI) |  |
| Brood Year | Total | (N) | Hatchery Origin | Natural Origin ${ }^{2}$ |
| 2013 | 308 | 293 | $59[46,75]$ | 249 [233, 262] |
| 2014 | 221 | 111 | 17 [9, 34] | 204 [187, 212] |
| 2015 | 301 | 141 | 11 [4, 26] | 290 [275, 297] |
| 2016 | 247 | 94 | 11 [6, 20] | 236 [227, 241] |

${ }^{1}$ Includes only fish that were spawned.
${ }^{2}$ Origin based on the absence of otolith marks, coded-wire tags, or adipose clips.
${ }^{\text {a }}$ This data was collected from samples intermittently high-graded for broodstock and may not be representative of the entire return to the Priest Rapids Hatchery volunteer trap.
${ }^{\mathrm{b}}$ This data is representative of the entire volunteer return to the Priest Rapids Hatchery volunteer trap.

### 7.2 Origin Based on Coded-Wire Tag Recoveries

The expansions of CWT recoveries at PRH have until recent years frequently under estimated the returns of PRH origin fish by return year and brood year. This bias and steps taken to identify the source are provided in Appendix A.
All Chinook salmon returning to PRH and broodstock collected from the OLAFT and ABC were sampled for the presence of CWT. A total of 4,669 CWT fish were recovered from Chinook salmon sampled at PRH in 2016, of which 297 were obtained from the broodstock collected from the PRH volunteer trap (Appendix B and Appendix C). The broodstock collected from the PRH volunteer trap were generally culled to exclude CWT fish. Therefore, this CWT group is not representative of the volunteer broodstock. There were seven CWT recovered in the ABC broodstock and subsequently surplused. The ABC fish were not screened for a CWT during
collection but were later scanned for CWT at the hatchery. The staff collecting the OLAFT fish effectively screened out CWT fish during the brood stock collection. The juvenile mark rate expansions of CWT at PRH in 2016 suggest that $93.6 \%$ of the returns to the PRH volunteer trap were hatchery origin fish. If we were to make the assumption that these CWT expansions accurately reflected the proportion of hatchery origin fish, then the remaining $6.4 \%$ of the unaccounted fish could potentially be natural origin (Table 3).

During return year 2016, PRH origin CWT tags accounted for $91.2 \%$ of the total return to the PRH volunteer trap. There were 11 natural origin CWT Hanford Reach fall Chinook salmon recovered at the hatchery in 2016; all were excluded from the broodstock while sorting out adipose clipped fish. There is not an expansion factor for the natural origin CWT fish so there was no attempt to estimate the proportion of natural origin fish based on these 11 CWT recoveries.

Table 3 Estimated proportion of hatchery and natural origin adult Chinook salmon returning to the Priest Rapids Hatchery volunteer trap based on coded-wire tag expansion. The entire collection was sampled for coded-wire tags, Return Years 2005-2016.

| Return Year | Returns to Priest Rapids Hatchery Volunteer Trap | Origin based on Coded-Wire Tag expansions |  | Natural Origin ${ }^{1}$ |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Priest Rapids Hatchery | Other Hatchery |  |
| 2005 | 10,616 | 0.622 | 0.006 | 0.329 |
| 2006 | 8,223 | 0.490 | 0.006 | 0.436 |
| 2007 | 6,000 | 0.671 | 0.004 | 0.525 |
| 2008 | 19,586 | 0.491 | 0.008 | 0.409 |
| 2009 | 12,778 | 0.428 | 0.003 | 0.540 |
| 2010 | 19,169 | 0.602 | 0.003 | 0.486 |
| 2011 | 20,823 | 0.613 | 0.006 | 0.381 |
| 2012 | 28,039 | 0.692 | 0.004 | 0.304 |
| 2013 | 41,831 | 0.713 | 0.034 | 0.252 |
| 2014 | 77,259 | 0.809 | 0.020 | 0.170 |
| 2015 | 63,978 | 0.914 | 0.015 | 0.071 |
| 2016 | 28,786 | 0.912 | 0.024 | 0.064 |
| Mean | 28,027 | 0.663 | 0.011 | 0.331 |
| Median | 19,586 | 0.647 | 0.006 | 0.355 |

${ }^{1}$ The proportion not accounted for by coded-wire tag expansion is assumed to be of natural origin.

### 8.0 Brood stock Collection and Sampling

The broodstock collected at the PRH volunteer trap were systematically sampled at a rate of 1:5 for otoliths (origin), scales (age), gender, and length. The broodstock collected at the OLAFT and ABC were sampled at a 1:2 rate for otoliths (origin), scales (age), gender, and length.

### 8.1 Broodstock Age Composition

A combined total of 4,938 fish were spawned from the three sources of broodstock. In general, hatchery origin broodstock tend to be younger than natural origin broodstock (Table 4). The historical broodstock age compositions are not directly comparable to the 2012 through 2015 broodstock age compositions due to inconsistent methodology for assigning origin. Prior to

2012, the origin of broodstock was estimated by adult CWT recoveries which in turn were expanded by the specific juvenile tag rates.
Table 4 Age composition for hatchery and natural origin fall Chinook salmon spawned at Priest Rapids Hatchery (includes all sources of broodstock), Return Years 20017-2016. Proportions calculated from expanded age compositions by origin for each source of broodstock to account for differing sample rates.

| Return Year | Origin | Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural ${ }^{1}$ | 0.000 | 1.000 | 0.000 | 0.000 | 0.000 |
|  | Hatchery ${ }^{1}$ | 0.081 | 0.274 | 0.486 | 0.138 | 0.020 |
| 2008 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery ${ }^{1}$ | 0.011 | 0.848 | 0.100 | 0.039 | 0.002 |
| 2009 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery ${ }^{1}$ | 0.012 | 0.086 | 0.883 | 0.019 | 0.000 |
| 2010 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery | 0.016 | 0.755 | 0.111 | 0.118 | 0.000 |
| 2011 | Natural ${ }^{1}$ | -- | -- | -- | -- | -- |
|  | Hatchery ${ }^{1}$ | 0.010 | 0.229 | 0.753 | 0.008 | 0.000 |
| 2012 | Natural ${ }^{2}$ | 0.032 | 0.435 | 0.400 | 0.131 | 0.002 |
|  | Hatchery ${ }^{2}$ | 0.006 | 0.487 | 0.376 | 0.130 | 0.000 |
| 2013 | Natural ${ }^{2}$ | 0.000 | 0.446 | 0.517 | 0.037 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.001 | 0.658 | 0.339 | 0.002 | 0.000 |
| 2014 | Natural ${ }^{2}$ | 0.000 | 0.045 | 0.886 | 0.070 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.064 | 0.897 | 0.039 | 0.000 |
| 2015 | Natural ${ }^{2}$ | 0.000 | 0.183 | 0.506 | 0.305 | 0.006 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.210 | 0.680 | 0.110 | 0.000 |
| 2016 | Natural ${ }^{2}$ | 0.000 | 0.101 | 0.761 | 0.138 | 0.000 |
|  | Hatchery ${ }^{2}$ | 0.000 | 0.099 | 0.700 | 0.196 | 0.007 |

${ }^{1}$ Origin determined from coded-wire tag expansions of juvenile mark rate.
${ }^{2}$ Origin determined from presence of hatchery marks (i.e., coded-wire tags, adipose clips, and otoliths)
In recent years, the broodstock selected from the PRH volunteer trap consisted primarily of age- 4 fish (Table 5). The hatchery origin broodstock for return years 2012 and 2013 had higher proportions of age- 3 fish due to the scarcity of older fish returning to the trap. The hatchery and natural origin fish broodstock recovered and high-graded at the OLAFT were primarily age- 4 (Table 6). Adipose clipped fish and all jack were generally excluded from the fish collected from the ABC fishery. In recent years, both the PRH origin and natural origin fish spawned from the ABC broodstock were mostly age-4 (Table 7).

Table 5 Age composition for hatchery and natural origin fall Chinook broodstock collected from the Priest Rapids Hatchery volunteer trap, Return Years 2012-2016.

| Return Year |  | Age Composition |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Origin $^{\mathbf{1}}$ | $\mathbf{N}$ |  |  |  |  |  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
|  | Natural | 39 | 0.000 | 0.295 | 0.585 | 0.121 | 0.000 |  |  |  |  |  |  |
|  | Hatchery | 646 | 0.000 | 0.477 | 0.389 | 0.134 | 0.000 |  |  |  |  |  |  |
| 2013 | Natural | 11 | 0.000 | 0.390 | 0.610 | 0.000 | 0.000 |  |  |  |  |  |  |
|  | Hatchery | 497 | 0.000 | 0.656 | 0.342 | 0.002 | 0.000 |  |  |  |  |  |  |
| 2014 | Natural | 26 | 0.000 | 0.115 | 0.885 | 0.000 | 0.000 |  |  |  |  |  |  |
|  | Hatchery | 548 | 0.000 | 0.065 | 0.899 | 0.036 | 0.000 |  |  |  |  |  |  |
| 2015 | Natural | 55 | 0.000 | 0.218 | 0.491 | 0.273 | 0.018 |  |  |  |  |  |  |
|  | Hatchery | 627 | 0.000 | 0.215 | 0.668 | 0.116 | 0.000 |  |  |  |  |  |  |
| 2016 | Natural | 49 | 0.000 | 0.102 | 0.776 | 0.122 | 0.000 |  |  |  |  |  |  |
|  | Hatchery | 778 | 0.000 | 0.100 | 0.763 | 0.136 | 0.000 |  |  |  |  |  |  |

${ }^{1}$ Origin determined from "in-sample" otoliths, adipose clips and/or coded-wire tags.
Table 6 Age composition for hatchery and natural origin fall Chinook salmon broodstock collected from the Off Ladder Adult Fish Trap at Priest Rapids Dam, Return Years 2012-2016.

| Return Year |  | Age Composition |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Origin $^{\mathbf{1}}$ | $\mathbf{N}$ |  |  |  |  |  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
|  | Natural | 281 | 0.048 | 0.540 | 0.257 | 0.151 | 0.004 |  |  |  |  |  |  |
|  | Hatchery | 219 | 0.106 | 0.687 | 0.136 | 0.071 | 0.000 |  |  |  |  |  |  |
| 2013 | Natural | 116 | 0.000 | 0.353 | 0.595 | 0.052 | 0.000 |  |  |  |  |  |  |
|  | Hatchery | 85 | 0.000 | 0.588 | 0.400 | 0.012 | 0.000 |  |  |  |  |  |  |
| 2014 | Natural | 186 | 0.000 | 0.000 | 0.902 | 0.098 | 0.000 |  |  |  |  |  |  |
|  | Hatchery | 39 | 0.000 | 0.000 | 0.870 | 0.130 | 0.000 |  |  |  |  |  |  |
| 2015 | Natural | 143 | 0.000 | 0.132 | 0.514 | 0.347 | 0.007 |  |  |  |  |  |  |
|  | Hatchery | 21 | 0.000 | 0.211 | 0.563 | 0.226 | 0.000 |  |  |  |  |  |  |
| 2016 | Natural | 155 | 0.000 | 0.058 | 0.677 | 0.245 | 0.019 |  |  |  |  |  |  |
|  | Hatchery | 56 | 0.000 | 0.089 | 0.643 | 0.250 | 0.018 |  |  |  |  |  |  |
| Mean | Natural | $\mathbf{1 7 6}$ | $\mathbf{0 . 0 1 0}$ | $\mathbf{0 . 2 1 7}$ | $\mathbf{0 . 5 8 9}$ | $\mathbf{0 . 1 7 9}$ | $\mathbf{0 . 0 0 6}$ |  |  |  |  |  |  |
|  | Hatchery | $\mathbf{8 4}$ | $\mathbf{0 . 0 2 1}$ | $\mathbf{0 . 3 1 5}$ | $\mathbf{0 . 5 2 2}$ | $\mathbf{0 . 1 3 8}$ | $\mathbf{0 . 0 0 4}$ |  |  |  |  |  |  |

[^1]Table 7 Age composition for hatchery and natural origin fall Chinook salmon broodstock collected form Angler Broodstock Collection, Return Years 20122016.

| Return Year | Origin ${ }^{1}$ | Age Composition |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | N | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2012 | Natural | 59 | 0.000 | 0.542 | 0.339 | 0.119 | 0.000 |
|  | Hatchery | 6 | 0.000 | 0.667 | 0.333 | 0.000 | 0.000 |
| 2013 | Natural | 237 | 0.000 | 0.511 | 0.468 | 0.021 | 0.000 |
|  | Hatchery | 56 | 0.000 | 0.839 | 0.161 | 0.000 | 0.000 |
| 2014 | Natural | 102 | 0.000 | 0.126 | 0.830 | 0.044 | 0.000 |
|  | Hatchery | 9 | 0.059 | 0.369 | 0.572 | 0.000 | 0.000 |
| 2015 | Natural | 136 | 0.000 | 0.196 | 0.499 | 0.305 | 0.000 |
|  | Hatchery | 5 | 0.000 | 0.397 | 0.603 | 0.000 | 0.000 |
| 2016 | Natural | 90 | 0.000 | 0.156 | 0.656 | 0.189 | 0.000 |
|  | Hatchery | 4 | 0.000 | 0.250 | 0.750 | 0.000 | 0.000 |
| Mean | Natural | 125 | 0.000 | 0.306 | 0.558 | 0.136 | 0.000 |
|  | Hatchery | 16 | 0.012 | 0.504 | 0.484 | 0.000 | 0.000 |

${ }^{1}$ Origin determined from "in-sample" otoliths, adipose clips and/or coded-wire tags.

### 8.2 Length by Age Class of Broodstock

The mean fork length (cm) by age for each source of broodstock is provided in Table 8. Both the hatchery origin and natural origin age- 3 fish collected at the OLAFT appear to be slightly larger than age- 3 fish collected at other locations. This may be due to the size high-grading processes.
Table 8 Mean fork length (cm) at age (total age) of fall Chinook salmon sampled from each source of broodstock spawned at Priest Rapids Hatchery, Return Year 2016. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation.

| Source of Broodstock | Origin ${ }^{1}$ | Fall Chinook Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| Volunteer Returns | Natural | 0 |  |  | 78 | 73 | 3 | 594 | 79 | 4 | 106 | 85 | 6 | 0 |  |  |
|  | Hatchery | 0 |  |  | 133 | 71 | 4 | 437 | 80 | 4 | 79 | 84 | 5 | 0 |  |  |
| OLAFT | Natural | 0 |  |  | 9 | 76 | 2 | 105 | 81 | 4 | 38 | 89 | 6 | 3 | 95 | 8 |
|  | Hatchery | 0 |  |  | 5 | 79 | 1 | 36 | 82 | 5 | 14 | 86 | 8 | 1 | 99 | 0 |
| ABC | Natural | 0 |  |  | 14 | 66 | 5 | 59 | 79 | 6 | 17 | 89 | 4 | 0 |  |  |
|  | Hatchery | 0 |  |  | 2 | 67 | 3 | 3 | 80 | 4 | 0 |  |  |  |  |  |

[^2]Table $9 \quad$ Mean fork length (cm) at age (total age) of hatchery and natural origin fall Chinook salmon collected from volunteer broodstock for the Priest Rapids Hatchery program. $\mathrm{N}=$ sample size and $\mathrm{SD}=$ standard deviation, Return Years 2012-2016.

| Return Year | Origin ${ }^{1}$ | Fall Chinook Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2012 | Natural | 0 |  |  | 12 | 71 | 4 | 25 | 82 | 4 | 5 | 86 | 4 | 0 |  |  |
|  | Hatchery | 0 |  |  | 298 | 70 | 4 | 253 | 81 | 5 | 91 | 88 | 7 | 0 |  |  |
| 2013 | Natural | 0 |  |  | 4 | 76 | 4 | 7 | 78 | 4 | 0 |  |  | 0 |  |  |
|  | Hatchery | 0 |  |  | 288 | 71 | 4 | 200 | 80 | 5 | 2 | 85 | 4 | 0 |  |  |
| 2014 | Natural | 0 |  |  | 3 | 74 | 2 | 23 | 80 | 5 | 0 |  |  | 0 |  |  |
|  | Hatchery | 0 |  |  | 36 | 70 | 3 | 491 | 78 | 5 | 21 | 87 | 6 | 0 |  |  |
| 2015 | Natural | 0 |  |  | 12 | 74 | 7 | 30 | 79 | 6 | 15 | 86 | 4 | 1 | 87 | 0 |
|  | Hatchery | 0 |  |  | 133 | 71 | 4 | 437 | 80 | 4 | 79 | 84 | 5 | 0 |  |  |
| 2016 | Natural | 0 |  |  | 78 | 73 | 3 | 594 | 79 | 4 | 106 | 85 | 6 | 0 |  |  |
|  | Hatchery | 0 |  |  | 133 | 71 | 4 | 437 | 80 | 4 | 79 | 84 | 5 | 0 |  |  |

${ }^{1}$ It is assumed for this analysis that all fish not possessing an otolith mark, ad-clipped or hatchery origin coded-wire tag were natural origin.

### 8.3 Gender Ratios

PRH staff sort and select broodstock from the trap to meet their egg-take goals and male-tofemale spawner ratio which is generally 1:2. Additional broodstock were collected from the OLAFT and ABC. The 2016 broodstock population was comprised of $68.9 \%$ females, resulting in an overall male to female ratio of $0.45: 1.00$ which is lower than the historic mean ratio of 0.53:1.00 (Table 10). This lower ratio of males to females resulted from the 138 matings of 1male x 4 -females during the real-time otolith read/alternative mating strategy study.

Table 10 Number of male and female hatchery fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 2001-2016. Ratios of males to females are also provided.

| Return Year | Males (M) | Females (F) | M/F Ratio |
| :---: | :---: | :---: | :---: |
| 2001 | 1,697 | 3,289 | $0.52: 1.00$ |
| 2002 | 1,936 | 3,628 | $0.53: 1.00$ |
| 2003 | 1,667 | 3,176 | $0.52: 1.00$ |
| 2004 | 1,688 | 3,099 | $0.54: 1.00$ |
| 2005 | 1,962 | 3,326 | $0.59: 1.00$ |
| 2006 | 1,777 | 3,322 | $0.53: 1.00$ |
| 2007 | 850 | 1,301 | $0.65: 1.00$ |
| 2008 | 1,823 | 3,195 | $0.57: 1.00$ |
| 2009 | 1,531 | 3,000 | $0.51: 1.00$ |
| 2010 | 1,809 | 3,447 | $0.52: 1.00$ |
| 2011 | 1,858 | 3,000 | $0.62: 1.00$ |
| 2012 | 1,749 | 3,225 | $0.54: 1.00$ |
| 2013 | 1,865 | 3,578 | $0.52: 1.00$ |
| $2014^{\mathrm{a}}$ | 1,805 | 3,688 | $0.49: 1: 00$ |
| $2015^{\mathrm{a}}$ | 1,697 | 3,827 | $0.44: 1: 00$ |
| $2016^{\mathrm{a}}$ | 1,537 | 3,401 | $0: 45: 1: 00$ |
| Mean | $\mathbf{1 , 7 0 3}$ | $\mathbf{3}, 219$ | $\mathbf{0 . 5 3 : 1 . 0 0}$ |

${ }^{\mathrm{a}}$ Includes broodstock used in the 1-male x 4-females alternative mating strategy.

### 8.4 Fecundity

The annual mean fecundity for PRH was calculated as the proportion of the total number of females spawned to the total estimated take of green eggs. The total number of green eggs is calculated after the first pick of dead eggs from the incubation trays. Fish culture staff weigh large lots of either dead or live eggs and then sub-sample the lots to calculate a mean individual egg weight. The number of eggs per lot is estimated by dividing the weight of the each egg lot by the calculated mean individual egg weight. The egg count for each lot is summed to estimate the facility egg-take. Each egg lot likely contained slightly varying amounts of interstitial water which might overestimate the egg count.
Fecundity for the 2016 broodstock averaged 3,649 eggs per female which is similar to that observed in 2015 but less than the historical mean of 3,966 (Table 11). Pre-spawn egg loss was often observed during the electro-anesthetic and pneumatic fish euthanizing process and may contribute to the reduced fecundity of fish in recent years.

Table 11 Mean fecundity of fall Chinook salmon collected for broodstock at Priest Rapids Hatchery, Return Years 2001-2016.

| Return Year | Egg-Take | Viable Females | Fecundity/Female |
| :---: | :---: | :---: | :---: |
| 2001 | $10,750,000$ | 3,161 | 3,401 |
| 2002 | $12,180,000$ | 3,489 | 3,491 |
| 2003 | $12,814,000$ | 3,078 | 4,163 |
| 2004 | $12,753,500$ | 3,019 | 4,224 |
| 2005 | $14,085,000$ | 3,211 | 4,386 |
| 2006 | $13,511,200$ | 3,217 | 4,200 |
| $2007^{\mathrm{a}}$ | $5,067,319$ | 1,249 | 4,057 |


| Return Year | Egg-Take | Viable Females | Fecundity/Female |
| :---: | :---: | :---: | :---: |
| 2008 | $12,643,600$ | 3,074 | 4,113 |
| 2009 | $13,074,798$ | 2,858 | 4,575 |
| 2010 | $11,903,407$ | 3,342 | 3,562 |
| 2011 | $12,693,000$ | 3,038 | 4,178 |
| 2012 | $12,398,389$ | 3,053 | 4,061 |
| 2013 | $12,947,070$ | 3,473 | 3,728 |
| 2014 | $14,321,183$ | 3,563 | 4,019 |
| 2015 | $13,530,988$ | 3,706 | 3,651 |
| 2016 | $12,411,530$ | 3,401 | 3,649 |
| Mean | $\mathbf{1 2 , 3 1 7 , 8 1 2}$ | $\mathbf{3 , 1 2 1}$ | $\mathbf{3 , 9 6 6}$ |

Fecundities of individual females were taken from sub-samples at PRH during the spawn of 2010 through 2016 broodstock to estimate fecundity by length and age. For the 2013 through 2016 brood year data, we show comparisons between hatchery and natural origin fall Chinook salmon sampled at PRH which include fork length/fecundity, fork length/egg size (weight) and fork length and gamete mass. For these years, we attempted to stratify the females sampled by fork length categories to obtain fecundity samples for all sizes of fish to better estimate the relationship between length and fecundity. Comparisons between age classes are not representative of the females spawned from 2013 through 2016 broodstock populations.

M\&E staff performed the fecundity estimates on green eggs during the spawn days. The entire gamete mass was drained of most all ovarian fluid and weighed within 0.1 gram. Sub-sample sizes ranged between years from 60 or 100 green eggs which were counted out and weighed within 0.01 gram to estimate individual egg weight (g) for each female. This sample size was determined to be sufficient based upon previous work that examined different samples sizes (Richards and Pearsons, 2014). The total fecundity of each female was estimated by dividing the weight of the total egg mass by the calculated mean individual egg weight. Each sample of the total egg mass likely contained slight varying amounts of ovarian fluid which might over estimate fecundity.

The fecundity data was pooled for return year 2010 through 2016 to provide a simple linear regression to predict fecundity based on fork-length (natural and hatchery females combined). This data shows a strong positive correlation between size and fecundity (Figure 4). The regression formula may be useful for coarse predictions of egg production for different size fish.


Figure 4 Linear relationship between fecundity and fork length for combined samples of natural and hatchery origin fall Chinook salmon spawned at Priest Rapids Hatchery, Return Years 2010-2016.

Fecundity samples collected in years 2010 through 2012 were not identified as to the origin of the females. For years 2013 through 2016, fecundity samples were taken from the broodstock at PRH to collect data associated with fecundity by size, age and origin (hatchery or natural).

Females were selected from both the PRH volunteer broodstock as well as from ponds which possessed broodstock primarily from the OLAFT and ABC. For the most part, the origin of fish during sampling was unknown. Therefore, we made a concerted effort to select females that were not adipose clipped so as to increase the chances of obtaining natural origin fish which were less common than hatchery origin fish. The origins of females sampled for fecundity were determined by hatchery marks (i.e., otoliths, adipose clips and CWTs). We make the assumption that fish not possessing any type of hatchery marks were of natural origin.
The mean fecundity by age is given in Table 12. This information is useful for forecasting potential egg-takes based on the numbers and age composition of the forecasted return.

Table 12 Mean fecundity at age for fall Chinook salmon sampled at the Priest Rapids Hatchery, Return Years 2010-2016. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation.

| Return Year | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{N}$ | Mean | SD | $\mathbf{N}$ | Mean | $\mathbf{S D}$ | $\mathbf{N}$ | Mean | SD |
| 2010 | 273 | 3,658 | 834 | 17 | 3,664 | 585 | 1 | 4,217 |  |
| 2011 | 30 | 3,538 | 842 | 206 | 4,276 | 884 | 1 | 4,380 |  |
| 2012 | 2 | 3,639 | 882 | 3 | 4,282 | 1089 | 0 |  |  |
| 2013 | 105 | 3,488 | 768 | 68 | 4,152 | 788 | 4 | 5,339 | 805 |
| 2014 | 1 | 3,358 | 0 | 73 | 4,126 | 755 | 5 | 4,416 | 407 |
| 2015 | 1 | 3,169 | 382 | 53 | 3,662 | 606 | 25 | 4,746 | 691 |
| 2016 | 14 | 3,192 | 559 | 101 | 3,676 | 639 | 36 | 4,173 | 693 |
| Mean | 61 | 3,435 | 711 | 74 | 3,977 | 764 | 12 | 4,545 | 649 |

The data collected from return years 2013 through 2016 were pooled to increase the number of samples for a given fork length. The linear relationships between fork length and variables including fecundity, mean egg weight, and total egg mass weight for natural and hatchery origin females subsampled are plotted Figure 5,Figure 6 andFigure 7. All relationships show a positive correlation with fork length. In addition, the relationships between fish size and egg data were similar for hatchery and natural origin fish.


Figure 5 Fecundity versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 2013-2016.


Figure 6 Mean egg weight versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 20132016.


Figure $7 \quad$ Total egg mass weight versus fork length for natural and hatchery origin fall Chinook salmon sub-sampled at Priest Rapids Hatchery, Return Years 20132016.

### 9.0 Hatchery Rearing

### 9.1 Number of Eggs Taken

In 2016, an estimated total of $12,411,530$ eggs were collected at PRH (Table 13). The egg-take goal for return year 2016 was $13,329,318$. The egg-take goal is calculated annually based on current program needs. This goal is established to meet the fall Chinook salmon production goals at both PRH and RSH as well as provide eggs for the Salmon in the Classroom Program.

PRH incubates approximately 8.4 million eyed eggs to produce the 7.3 million smolt release at the hatchery. Roughly an additional 4.1 million eyed eggs are needed to meet the program goal of eyed eggs delivered to Bonneville Hatchery for the 3.5 million subyearling releases from RSH.
Table 13 Number of eggs taken from fall Chinook salmon broodstock collected at Priest Rapids Hatchery, Return Years 1984-2016.

| Return Year | Number of Eggs Taken | Return Year | Number of Eggs Taken |
| :---: | :---: | :---: | :---: |
| 1984 | 10,342,000 | 2001 | 10,750,000 |
| 1985 | 10,632,000 | 2002 | 12,180,000 |
| 1986 | 22,126,100 | 2003 | 12,814,000 |
| 1987 | 24,123,000 | 2004 | 12,753,500 |
| 1988 | 16,682,000 | 2005 | 14,085,000 |
| 1989 | 13,856,500 | 2006 | 13,511,200 |
| 1990 | 9,605,000 | 2007 | 5,067,319 |
| 1991 | 6,338,000 | 2008 | 12,643,600 |
| 1992 | 11,156,400 | 2009 | 13,074,798 |
| 1993 | 14,785,000 | 2010 | 11,903,407 |
| 1994 | 16,074,600 | 2011 | 12,693,000 |
| 1995 | 17,345,900 | 2012 | 12,398,389 |
| 1996 | 14,533,500 | 2013 | 13,276,000 |
| 1997 | 17,007,000 | 2014 | 14,321,818 |
| 1998 | 13,981,300 | 2015 | 12,692,400 |
| 1999 | 16,089,600 | 2016 | 12,411,530 |
| 2000 | 15,359,500 | 9 year (08-16) Mean ${ }^{1}$ | 12,823,882 |

${ }^{1}$ Began additional annual egg-takes starting in return year 2008 for the 3.5 million Ringold Springs Hatchery Program

### 9.2 Number of Acclimation Days

The 2016 brood were incubated on a combination of well and river water before being transferred to intermediate concrete raceways and then transferred to the concrete holding ponds for final acclimation before release into the Columbia River in late May and June 2017. The eggtakes for the 2016 brood were distributed into eleven batches associated with the dates in which fish were spawned. The number of acclimation days ranged from 97 for the later egg-takes to 116 for the earlier egg-takes (Table 14).

Table 14 Number of days fall Chinook salmon fry were reared at Priest Rapids Hatchery prior to release, Brood Year 2016.

| Batch | Egg Tray to Raceway Transfer Date | Release Date | Number of Days |
| :---: | :---: | :---: | :---: |
| 1 | January 27 into Bank E | May 23 | 116 |
| 3 | January 27 into Bank E | May 23 | 116 |
|  | February 14 into Bank E | May 23 | 98 |
|  | February 14 into Bank D | May 25 | 100 |
|  | February 14 into Bank C | June 9 | 118 |
|  | February 28 into Bank C | June 9 | 104 |
|  | February 28 Bank B | June 12 | 111 |
|  | February 28 Back A | June 19 | 111 |
|  | February 28 into Bank A | June 19 | 111 |
| 9 | March 14 into Bank A | June 19 | 97 |
| 10 | March 14 into Bank A | June 19 | 97 |
| 11 | March 14 into Bank A | June 19 | 97 |

### 9.3 Annual Releases, Tagging and Marking

The annual release of fall Chinook salmon smolts from PRH range considerably since the initial release of roughly 2.38 million smolts from the 1979 brood year to over roughly 10.30 million from the 1982 brood year (Table 15). The 2016 release goal for PRH is $7,299,504$ smolts. This goal includes a recent increase in the Grant PUD mitigation from 5,000,000 to 5,599,504 combined with the ongoing USACE's John Day mitigation of $1,700,000$ smolts.
In 2016, PRH released an estimated 7,242,054 subyearling fall Chinook salmon from the 2015 broodstock (Table 16). Fish were released between June 16 and June 24.
Various mark types and rates have occurred at PRH over the years for both the Grant PUD and USACE mitigation fish. In 1976, PRH began adipose fin clipping and coded-wire tagging a portion of the juvenile fall Chinook released to determine PRH contributions to ocean and river fisheries. All smolts associated with the USACE John Day mitigation have been adipose clipped, but only small fractions included a CWT. Poor returns in 2007 precluded the production of USACE John Day mitigation fish for the 2008 release.

All PRH releases for both mitigation programs were $100 \%$ otolith marked beginning with the 2008 release. All intra-annual releases from PRH have the same annual otolith pattern, but the pattern differs between years. The eyed eggs produced for the RSH program have received an otolith mark for brood years 2010 through 2016. Otolith sampling at PRH and in the Hanford Reach should provide increased precision in the determination of PRH origin returns to the hatchery and Hanford Reach compared to CWT estimates. Given sufficient samples sizes, the otolith mark rate of $100 \%$ may provide better estimates than those from CWTs.
Since 1987, the U.S. Section of the Pacific Salmon Commission (PSC) has supported a coordinated project which seeks to capture and CWT 200,000 naturally produced juvenile fall Chinook salmon in the Hanford Reach. Fish are collected with seines over a ten day period between late May and early June. Fish are approximately 40-80 mm long at the time of capture. Recoveries from these tagged fish are used to estimate harvest exploitation rates and interception
rates for Hanford Reach natural origin fall Chinook salmon. These data have also more recently been used to estimate the number of natural origin juveniles produced in the Hanford Reach (Harnish et al. 2012, Harnish 2017).
WDFW operates the OLAFT at Priest Rapids Dam three days per week beginning in July and continuing through mid to late October. This project began in 1986 and was designed to sample steelhead to (1) determine upriver run size, (2) estimate hatchery to natural origin (wild) fish ratios, (3) determine age class distribution, and (4) evaluate the need for managing returning hatchery steelhead consistent with ESA recovery objectives. In 2009, WDFW began sampling fall Chinook salmon at the trap for run composition assessment. Beginning in 2010, the OLAFT was used to collect potential natural origin fall Chinook salmon for incorporation into the broodstock at PRH.

Table 15 Number of marked, unmarked, and tagged fall Chinook salmon smolts released from Priest Rapids Hatchery, Brood Years 1977-2016.

| Brood Year | Total Released | Non Ad-Clip Released | AD/CWT | CWT Only | AD Only | PIT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 150,625 | 0 | 147,338 | 0 | 3,287 |  |
| 1978 | 153,840 | 0 | 152,532 | 0 | 1,308 |  |
| 1979 | 3,005,654 | 2,858,509 | 147,145 | 0 |  |  |
| 1980 | 4,832,591 | 4,581,054 | 251,537 | 0 |  |  |
| 1981 | 5,509,241 | 5,198,365 | 310,876 | 0 |  |  |
| 1982 | 10,296,700 | 9,888,989 | 407,711 | 0 |  |  |
| 1983 | 9,742,700 | 9,517,263 | 222,055 | 0 | 3,382 |  |
| 1984 | 6,363,000 | 6,253,240 | 106,960 | 0 | 2,800 |  |
| 1985 | 6,048,000 | 5,843,176 | 203,534 | 0 | 1,290 |  |
| 1986 | 7,709,000 | 7,506,142 | 201,843 | 0 | 1,015 |  |
| 1987 | 7,709,000 | 7,501,578 | 196,221 | 0 | 11,201 |  |
| 1988 | 5,404,550 | 5,200,080 | 201,608 | 0 | 2,862 |  |
| 1989 | 6,431,100 | 6,224,770 | 194,530 | 0 | 11,800 |  |
| 1990 | 5,333,500 | 5,134,031 | 199,469 | 0 |  |  |
| 1991 | 7,000,100 | 6,798,453 | 201,647 | 0 |  |  |
| 1992 | 7,134,159 | 6,939,537 | 194,622 | 0 |  |  |
| 1993 | 6,705,836 | 6,520,153 | 185,683 | 0 |  |  |
| 1994 | 6,702,000 | 6,526,120 | 175,880 | 0 |  | 1,500 ${ }^{\text {c }}$ |
| 1995 | 6,700,000 | 6,503,811 | 196,189 | 0 |  | $3,000^{\text {c }}$ |
| 1996 | 6,644,100 | 6,450,885 | 193,215 | 0 |  | $3,000^{\text {c }}$ |
| 1997 | 6,737,600 | 6,541,351 | 196,249 | 0 |  | $3,000^{\text {c }}$ |
| 1998 | 6,504,800 | 6,311,140 | 193,660 | 0 |  | $3,000^{\text {c }}$ |
| 1999 | 6,856,000 | 6,651,664 | 204,336 | 0 |  | $3,000^{\text {c }}$ |
| 2000 | 6,862,550 | 6,661,771 | 200,779 | 0 |  | $3,000^{\text {c }}$ |
| 2001 | 6,779,035 | 6,559,109 | 219,926 | 0 |  | $3,000^{\text {c }}$ |
| 2002 | 6,777,605 | 6,422,232 | 355,373 | 0 |  | $3,000^{\text {c }}$ |
| 2003 | 6,814,560 | 6,415,444 | 399,116 | 0 |  | $3,000^{\text {c }}$ |
| 2004 | 6,599,838 | 6,399,766 | 200,072 | 0 |  | $3,000^{\text {c }}$ |
| 2005 | 6,876,290 | 6,676,845 | 199,445 | 0 |  | $3,000^{\text {c }}$ |
| 2006 | 6,743,101 | 4,912,487 | 202,000 | 0 | 1,628,614 | $3,000^{\text {c }}$ |
| $2007{ }^{\text {a }}$ | 4,548,307 | 4,344,926 | 202,568 | 0 | $813^{\text {b }}$ | $3,000^{\text {c }}$ |
| $2008{ }^{\text {a }}$ | 6,788,314 | 4,850,844 | 218,082 | 0 | 1,719,388 | 2,994 ${ }^{\text {c }}$ |


| Brood Year | Total <br> Released | Non Ad-Clip <br> Released | AD/CWT | CWT Only | AD Only | PIT |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $2009^{\mathrm{a}}$ | $6,776,651$ | $3,413,334$ | 619,568 | $1,026,561$ | $1,717,188$ | $1,995^{\mathrm{c}}$ |
| $2010^{\mathrm{a}}$ | $6,798,390$ | $3,383,859$ | 602,580 | $1,108,990$ | $1,702,961$ | $3,000^{\mathrm{c}}$ |
| $2011^{\mathrm{a}}$ | $7,056,948$ | $3,094,666$ | 595,608 | 598,031 | $2,768,643$ | $42,844^{\mathrm{c}}$ |
| $2012^{\mathrm{a}}$ | $6,82,861$ | $2,905,694$ | 603,930 | 601,009 | $2,712,228$ | $42,908^{\mathrm{c}}$ |
| $2013^{\mathrm{a}}$ | $7,267,248$ | $3,347,417$ | 603,417 | 603,439 | $2,712,975$ | $42,908^{\mathrm{c}}$ |
| $2014^{\mathrm{a}}$ | $7,039,543$ | $3,125,734$ | 600,688 | 600,730 | $2,712,392$ | $42,621^{\mathrm{c}}$ |
| $2015^{\mathrm{a}}$ | $7,242,054$ | $3,317,992$ | 602,116 | 601,770 | $2,720,176$ | $42,999^{\mathrm{c}}$ |
| $2016^{\mathrm{a}}$ | $7,006,251$ | $3,045,689$ | 603,539 | 603,864 | $2,710,302$ | $42,858^{\mathrm{d}}$ |

${ }^{\mathrm{a}}$ Entire release was otolith marked
${ }^{\mathrm{b}}$ Low returns to PRH precluded the production of the USACE adipose clipped release.
${ }^{\text {c }}$ PIT tagged are included in the Non Ad-Clip totals
${ }^{\text {d}}$ PIT tagged are not included in the Non Ad-Clip totals

### 9.4 Fish Size and Condition at Release

The data associated with fish size and condition at release from PRH prior to brood year 2013 was obtained from the hatchery staff. The mean fish weight was obtained by weighing groups of roughly 300 fish sampled from each pond to the nearest gram and then dividing the group weight by the total number of fish weighed. The fork length of each fish from the group weight was measured to the nearest millimeter to calculate mean length and coefficient of variation. Samples from each of the rearing ponds were taken the day of release. The results were pooled to provide mean estimates for the facility as a whole. The size and condition data for the 2013 through 2016 broods were collected by M\&E staff the day prior to release for each pond. We attempted to collect representative samples by capturing multiple groups of fish with a cast net from the lower, middle, and upper third of the rearing pond. Each fish sampled was individually weighed to the nearest 0.1 gram and measured for fork length to the nearest millimeter. The results were pooled to provide mean estimates for the facility as a whole.
The goal for PRH is to release fall Chinook salmon smolts at 50 fish per pound. At release, the smolts from the 2016 brood averaged 49 fish per pound with an mean fork length of 89 mm , and a mean CV of 6.1 (Table 16). For brood years 1991 through 2016, smolts released from PRH have averaged 48 fish per pound with a mean fork length of 95 and a mean CV of 7.3.
Table 16 Mean length ( $\mathrm{FL}, \mathrm{mm}$ ), weight ( g and fish/pound), and coefficient of variations (CV) of fall Chinook smolts released from Priest Rapids Hatchery, Brood Years 1991-2016.

|  |  | Fork Length (mm) |  | Mean Weight <br> Brood year |  | Release Year |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Mean | $\mathbf{C V}$ | Grams (g) | Fish/pound | $\mathbf{N}$ |  |  |
| 1991 | 1992 | 93 | 8.7 | 8.3 | 55 | 1,500 |
| 1992 | 1993 | 92 | 8.6 | 8.3 | 54 | 1,500 |
| 1993 | 1994 | 95 | 6.9 | 9.3 | 49 | 1,500 |
| 1994 | 1995 | 96 | 6.7 | 9.7 | 47 | 1,500 |
| 1995 | 1996 | 97 | 6.6 | 10 | 45 | 1,500 |
| 1996 | 1997 | 95 | 11 | 8.7 | 52 | 1,500 |
| 1997 | 1998 | 103 | 8.9 | 10.1 | 45 | 1,500 |
| 1998 | 1999 | 95 | 6.5 | 9.6 | 48 | 1,500 |
| 1999 | 2000 | 93 | 6.6 | 8.9 | 51 | 1,500 |
| 2000 | 2001 | 97 | 6.3 | 10.2 | 45 | 1,500 |


| Brood year | Release Year | Fork Length (mm) |  | Mean Weight |  | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Mean | CV | Grams (g) | Fish/pound |  |
| 2001 | 2002 | 96 | 6.9 | 10.1 | 45 | 1,500 |
| 2002 | 2003 | 95 | 6.9 | 9.5 | 48 | 1,500 |
| 2003 | 2004 | 96 | 6.8 | 9.6 | 48 | 1,500 |
| 2004 | 2005 | 95 | 5.9 | 9.4 | 48 | 1,500 |
| 2005 | 2006 | 98 | 6.3 | 10.1 | 45 | 1,500 |
| 2006 | 2007 | 98 | 7.0 | 9.9 | 46 | 1,500 |
| 2007 | 2008 | 101 | 8.3 | 10.2 | 45 | 1,200 |
| 2008 | 2009 | 94 | 6.7 | 9.3 | 49 | 1,500 |
| 2009 | 2010 | 94 | 7.3 | 9.2 | 49 | 1,500 |
| 2010 | 2011 | 92 | 9.1 | 9.7 | 47 | 1,500 |
| 2011 | 2012 | 94 | 7.1 | 9.2 | 49 | 1,500 |
| 2012 | 2013 | 95 | 7.6 | 9.7 | 47 | 1,500 |
| 2013 | 2014 | 92 | 8.4 | 9.0 | 50 | 648 |
| 2014 | 2015 | 91 | 6.6 | 8.7 | 52 | 1,728 |
| 2015 | 2016 | 92 | 6.1 | 9.3 | 49 | 1,595 |
| 2016 | 2017 | 89 | 6.1 | 9.3 | 49 | 1,788 |
| Mean |  | 95 | 7.3 | 9.4 | 48 | 1,479 |

### 9.5 Survival Estimates

The survival proportion $\left(\mathrm{P}^{\wedge}\right)$ for egg to juvenile release for brood year 2016 was 0.819 which is lower than the historic mean of 0.852 (Table 17). The green egg to eyed egg stage is the most critical life stage at PRH during incubation/juvenile rearing because the greatest level of loss annually occurs at this stage. The green egg to eyed egg survival $\mathrm{P}^{\wedge}$ for brood year 2016 was 0.899 which is similar to the historical mean of 0.901 .

In 2016, survival $\mathrm{P}^{\wedge}$ of fish ponded for broodstock was 0.843 which is similar to the historic mean of 0.846 . The trapping operations in 2014 through 2016 were carried out in a manner which generally reduced fish densities in the trap; possibly resulting in reduced ponding mortality.
Table 17 Hatchery life stage survival ( $\mathbf{P}^{\wedge}$ ) for fall Chinook salmon at Priest Rapids Hatchery, Brood Years 1989-2016.

|  | PRH Volunteers Ponded to Spawned |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood year | Female | Male | Jack | Total | Unfertilized to <br> Eyed Egg | Eyed egg to <br> Ponding | Ponding to <br> Release | Fertilized Egg <br> to Release |
| 1989 |  |  |  | 0.919 | 0.866 | 0.976 | 0.950 | 0.821 |
| 1990 |  |  |  | 0.947 | 0.869 | 0.996 | 0.984 | 0.852 |
| 1991 |  |  |  | 0.973 | 0.948 | 0.993 | 0.998 | 0.922 |
| 1992 |  |  |  | 0.952 | 0.945 | 0.991 | 0.965 | 0.901 |
| 1993 |  |  |  | 0.917 | 0.941 | 0.984 | 0.974 | 0.902 |
| 1994 |  |  |  | 0.710 | 0.935 | 0.985 | 0.953 | 0.878 |
| 1995 |  |  |  | 0.897 | 0.914 | 0.980 | 0.962 | 0.862 |
| 1996 |  |  |  | 0.908 | 0.924 | 0.997 | 0.983 | 0.897 |
| 1997 |  |  |  | 0.900 | 0.915 | 0.996 | 0.970 | 0.790 |
| 1998 |  |  |  | 0.834 | 0.914 | 0.998 | 0.970 | 0.884 |


|  | PRH Volunteers Ponded to Spawned |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood year | Female | Male | Jack | Total | Unfertilized to <br> Eyed Egg | Eyed egg to <br> Ponding | Ponding to <br> Release | Fertilized Egg <br> to Release |
| 1999 |  |  |  | 0.759 | 0.897 | 0.997 | 0.995 | 0.888 |
| 2000 |  |  |  | 0.868 | 0.898 | 0.995 | 0.985 | 0.884 |
| 2001 | 0.776 | 0.732 | 0.665 | 0.757 | 0.886 | 0.994 | 0.975 | 0.859 |
| 2002 | 0.835 | 0.829 | 0.705 | 0.828 | 0.880 | 0.995 | 0.979 | 0.858 |
| 2003 | 0.893 | 0.817 | 0.698 | 0.858 | 0.882 | 0.989 | 0.989 | 0.868 |
| 2004 | 0.958 | 0.915 | 0.646 | 0.845 | 0.881 | 0.975 | 0.985 | 0.846 |
| 2005 | 0.890 | 0.890 | 0.782 | 0.886 | 0.914 | 0.976 | 0.991 | 0.884 |
| 2006 | 0.918 | 0.924 | 0.695 | 0.913 | 0.897 | 0.975 | 0.981 | 0.859 |
| 2007 | 0.967 | 0.748 | 0.642 | 0.861 | 0.858 | 0.996 | 0.981 | 0.898 |
| 2008 | 0.943 | 0.896 | 0.877 | 0.924 | 0.902 | 0.973 | 0.877 | 0.877 |
| 2009 | 0.848 | 0.901 | 0.916 | 0.864 | 0.912 | 0.977 | 0.891 | 0.891 |
| 2010 | 0.803 | 0.831 | 0.803 | 0.809 | 0.913 | 0.985 | 0.977 | 0.841 |
| 2011 | 0.611 | 0.847 | 0.737 | 0.679 | 0.903 | 0.985 | 0.985 | 0.875 |
| 2012 | 0.643 | 0.786 | 0.630 | 0.688 | 0.873 | 0.970 | 0.962 | 0.787 |
| 2013 | 0.698 | 0.660 | 0.333 | 0.684 | 0.884 | 0.983 | 0.951 | 0.806 |
| 2014 | 0.830 | 0.880 | N/A | 0.847 | 0.870 | 0.970 | 0.973 | 0.817 |
| 2015 | 0.841 | 0.810 | N/A | 0.830 | 0.917 | 0.977 | 0.965 | 0.827 |
| 2016 | 0.873 | 0.782 | N/A | 0.843 | 0.899 | 0.973 | 0.964 | 0.819 |
| Mean | $\mathbf{0 . 8 3 4}$ | $\mathbf{0 . 8 3 8}$ | $\mathbf{0 . 7 0 5}$ | $\mathbf{0 . 8 2 3}$ | $\mathbf{0 . 8 9 2}$ | $\mathbf{0 . 9 8 0}$ | $\mathbf{0 . 9 6 3}$ | $\mathbf{0 . 8 5 2}$ |

### 9.6 Juvenile PIT Tag Detections at the Priest Rapids Hatchery Array

Roughly 3,000 sub-yearlings at PRH were annually PIT tagged and released from PRH for brood years 1995 through 2010 to assess timing, migration speed, and juvenile survival from PRH to McNary Dam. The analysis for these measures is reported annually by the Fish Passage Center and can be found at www.fpc.org/documents/FPC_memos.html.
Beginning with the 2011 brood, approximately 40,000 additional juveniles were annually PIT tagged and released to bolster the data collected for estimation of juvenile abundance at release and adult straying. These tags can also be used to estimate adult migration timing, conversion rates from Bonneville Dam to McNary Dam to PRH, smolt to adult survival rates, as well as fallback and re-ascension estimates at McNary, Ice Harbor, and Priest Rapids dams. The annual detection rates are given in Table 18. Prior to the 2012 release (brood year 2011), a PIT tag array consisting of six antennas was installed in the hatchery discharge channel to detect both juvenile out-migrants and adult returns. The detection rates reported below account for the relatively few shed PIT tags found in the rearing raceways. Prior to the release of the 2016 brood, the mortalities routinely recovered from the rearing ponds were not scanned for PIT tags. This prohibits us from knowing the actual total number of PIT tagged fish released. Hence, the overall proportion of released PIT tagged fish detected would likely be higher than reported if we knew the actual number of live PIT tagged fish that left the ponds.
The overall detection rate for the releases of the 2011 brood year was $70.4 \%$. The releases occurred over an eight day period, with only two days of consecutive releases. Detection rates for the 2011 brood year release may have been reduced as a result of the array being inundated by high river elevations during portions of releases. The overall detection rate for the 2012 brood
year was $3.4 \%$. The low detection rates were likely due to force releasing all of the smolts in four consecutive days which appears to have overwhelmed the PIT tag detection equipment. The restricted release period was necessitated by the construction schedule of the new hatchery.
A concerted effort was made during both the 2013 and 2014 brood year releases to improve the PIT tag detection efficiency at the PRH array. First, the automatic upload function of the array was discontinued to reduce the usage demand on the system's processor. Secondly, the five releases from the hatchery were conducted over a fourteen day period beginning on June 12 to spread out over time the number of PIT tags passing the array. This was managed by pulling the individual weir boards for each pond over a two day period. The percentage of PIT tagged subyearlings detected for the 2013 and 2014 brood years were $92.9 \%$ and $94.5 \%$, respectively.
The releases of the 2015 brood occurred every two days between June 16 and June 24, 2016 to accommodate a day versus night release evaluation. During the evaluation, all weir boards for a given pond where incrementally pulled over an eight hour period on the date of release. Overall, $84.3 \%$ of the PIT tagged subyearlings were detected. The detected rate between release groups varied from $33.6 \%$ to $97.0 \%$. These values are lower than the previous two years. It's possible that forced releases over an eight hour period may have resulted in high rates of tag collision at the array resulting in poor detection efficiency.
The releases of the 2016 brood were initiated at 9PM for each pond. All weir boards were pulled by 3AM. Releases occurred irregularly between May 23 and June 19. We anticipated river flows during May and June to exceed 240kcfs which results in the inundation of the PRH array. A temporary two antenna array was installed at a higher elevation near the upper end of the discharge channel to complement the PRH array. The overall detection rate was $95.4 \%$ for the combined release of all ponds, ranging from $89.7 \%$ to $97.5 \%$.

Table 18 Number of sub-yearlings PIT tagged, mark and release dates, and the number of unique tags detected at the array in the Priest Rapids discharge channel, Brood Years 2011-2016.

| Brood Year | Tag File | Tagging Date | Release Date | \# Tagged | \# of Tags Recovered from Facility Mortalities | \# of Unique Detections | \% <br> Detected |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | CSM12114.A01 | 4/23/2012 | 6/20/2012 | 9937 | No Data | 6,277 | 63.2 |
| 2011 | CSM12114.A04 | 4/23/2012 | 6/14/2012 | 9948 | No Data | 6,674 | 67.1 |
| 2011 | CSM12114.A03 | 4/24/2012 | 6/15/2012 | 9997 | No Data | 6,963 | 69.7 |
| 2011 | CSM12115.A02 | 4/24/2012 | 6/16/2012 | 9967 | No Data | 8,115 | 81.4 |
| 2011 | CSM12115.A02 | 5/30/2012 | 6/20/2012 | 1000 | No Data | 499 | 49.9 |
| 2011 | SMP12151.PR2 | 5/30/2012 | 6/16/2012 | 998 | No Data | 806 | 80.8 |
| 2011 | SMP12152.PR3 | 5/31/2012 | 6/12/2012 | 996 | No Data | 810 | 81.3 |
|  |  |  | Totals | 42,844 | N/A | 30,144 | 70.4 |
| 2012 | CSM13143.A06 | 5/23/2013 | 6/14/2013 | 9,982 | No Data | 317 | 3.2 |
| 2012 | CSM13143.A07 | 5/23/2013 | 6/13/2013 | 9,983 | No Data | 267 | 2.7 |
| 2012 | CSM13144.A08 | 5/24/2013 | 6/12/2013 | 9,974 | No Data | 335 | 3.4 |
| 2012 | CSM13144.A09 | 5/24/2013 | 6/15/2013 | 9,977 | No Data | 325 | 3.3 |
| 2012 | SMP13149.PR1 | 5/29/2013 | 6/15/2013 | 997 | No Data | 131 | 13.1 |
| 2012 | SMP13149.PR2 | 5/29/2013 | 6/14/2013 | 996 | No Data | 33 | 3.3 |
| 2012 | SMP13150.PR3 | 5/30/2013 | 6/12/2013 | 999 | No Data | 48 | 4.9 |
|  |  |  | Totals | 42,908 | N/A/ | 1,456 | 3.4 |
| 2013 | CSM14148.PRA | 5/28/2014 | 6/25/2014 | 7,994 | 21 | 7,215 | 90.5 |
| 2013 | CSM14148.PRB | 5/28/2014 | 6/23/2014 | 7,998 | 14 | 7,215 | 92.5 |
| 2013 | CSM14149.PRC | 5/29/2014 | 6/18/2014 | 7,996 | 11 | 7,443 | 93.2 |
| 2013 | CSM14149.PRD | 5/29/2014 | 6/16/2014 | 7,993 | 6 | 7,662 | 95.9 |
| 2013 | CSM14149.PRE | 5/29/2014 | 6/12/2014 | 7,998 | 7 | 7,407 | 92.7 |
| 2013 | SMP14148.PR1 | 5/29/2014 | 6/25/2014 | 996 | 0 | 914 | 91.8 |
| 2013 | SMP14148.PR2 | 5/29/2014 | 6/18/2014 | 994 | 0 | 927 | 93.3 |
| 2013 | SMP14149.PR3 | 5/30/2014 | 6/12/2014 | 998 | 0 | 951 | 95.3 |
|  |  |  | Total | 42,967 | 59 | 39,908 | 92.9 |
| 2014 | CSM15147.PRE | 5/27/2015 | 6/12/2015 | 7,999 | 169 | 7,438 | 95.0 |
| 2014 | CSM15147.PRD | 5/27/2015 | 6/15/2015 | 7,996 | 39 | 7,685 | 96.6 |
| 2014 | CSM15147.PRC | 5/27/2015 | 6/18/2015 | 7,996 | 63 | 7,524 | 94.8 |
| 2014 | CSM15147.PRB | 5/28/2015 | 6/22/2015 | 7,998 | 50 | 7,696 | 96.8 |
| 2014 | CSM15147.PRA | 5/28/2015 | 6/25/2015 | 7,994 | 31 | 7,447 | 93.5 |
| 2014 | SMP15140.PR1 | 5/20/2015 | 6/25/2015 | 993 | 0 | 940 | 94.7 |
| 2014 | SMP15140.PR2 | 5/20/2015 | 6/18/2015 | 998 | 0 | 946 | 94.8 |
| 2014 | SMP15141.PR3 | 5/21/2015 | 6/12/2015 | 999 | 0 | 935 | 93.6 |
|  |  |  | Total | 42,973 | 352 | 40,611 | 95.3 |
| 2015 | CSM16153.PRE | 6/01/2016 | 6/16/2016 | 7,996 | 13 | 6,032 | 75.6 |
| 2015 | CSM16153.PRD | 6/01/2016 | 6/18/2016 | 7,998 | 224 | 7,537 | 97.0 |
| 2015 | CSM16153.PRC | 6/01/2016 | 6/20/2016 | 7,985 | 137 | 6,777 | 86.4 |
| 2015 | CSM16154.PRB | 6/02/2016 | 6/22/2016 | 7,993 | 13 | 7,136 | 89.4 |
| 2015 | CSM16154.PRA | 6/02/2016 | 6/24/2016 | 7,990 | 26 | 6,590 | 82.7 |
| 2015 | SMP16153.PR1 | 6/01/2016 | 6/24/2016 | 995 | 88 | 513 | 56.6 |
| 2015 | SMP16153.PR2 | 6/01/2016 | 6/20/2016 | 998 | 5 | 795 | 80.1 |
| 2015 | SMP16154.PR3 | 6/02/2016 | 6/16/2016 | 1001 | 109 | 300 | 33.6 |
|  |  |  | Totals | 42,956 | 615 | 35,680 | 84.3 |
| 2016 | BMI17129.PRE | 5/09/2017 | 5/23/2017 | 7,996 | 18 | 7,279 | 91.2 |


| Brood Year | Tag File | Tagging Date | Release Date | \# Tagged | \# of Tags <br> Recovered <br> from <br> Facility <br> Mortalities | \# of Unique Detections | $\begin{gathered} \% \\ \text { Detected } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2016 | BMI17129.PRD | 5/09/2017 | 5/25/2017 | 7,998 | 7 | 7,790 | 97.5 |
| 2016 | BMI17143.PRC | 5/23/2017 | 6/09/2017 | 7,981 | 32 | 7,714 | 97.0 |
| 2016 | BMI17143.PRB | 5/23/2017 | 6/12/2017 | 7,995 | 24 | 7,633 | 95.8 |
| 2016 | BMI17144.PRA | 5/24/2017 | 6/19/2017 | 7,995 | 46 | 7,633 | 96.0 |
| 2016 | SMP17128.PR1 | 5/08/2017 | 5/23/2017 | 600 | 0 | 538 | 89.7 |
| 2016 | SMP17129.PR2 | 5/09/2017 | 5/25/2017 | 600 | 0 | 579 | 96.5 |
| 2016 | SMP17144.PR3 | 5/24/2017 | 6/09/2017 | 598 | 0 | 568 | 95.0 |
| 2016 | SMP17144.PR4 | 5/24/2017 | 6/12/2017 | 601 | 0 | 581 | 96.7 |
| 2016 | SMP17144.PR5 | 5/24/2017 | 6/19/2017 | 600 | 2 | 570 | 95.3 |
|  |  |  | Totals | 42,964 | 129 | 40,885 | 95.4 |

### 10.0 Adult Fish Pathogen Monitoring

At spawning, adult fall Chinook are sampled for infectious hemotopoietic necrosis virus (IHNV), infectious pancreatic necrosis (IPN), viral hemorrhagic septicemia (VHS), paramyxovirus, aquaroviruses, as well as Renibacterium salmoninarum, the causative agent for bacterial kidney disease (BKD). Viral and bacterial inspections included sampling the ovarian fluid and kidney/spleen for pathogens (AFS-FHS 2014). Annual testing for BKD was initiated with the 2008 broodstock to address concerns associated with shipping eyed-eggs to Bonneville Hatchery for the USACE RSH production. All results of viral testing in 2016 were negative (Table 19). The risk of BKD was assayed using the enzyme linked immunosorbent assay (ELISA) for $R$. salmoninarum antigen (Elliot 2012). Results of adult broodstock BKD monitoring in 2016 indicated that 59 of the 60 ( $98.3 \%$ ) females tested had ELISA values less than an optical density of 0.10 (Below Low); 1 of the 60 samples was in the Low category between 0.10 and 0.199 (Table 20).
Table 19 Viral inspections of fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 1991-2016.

| Year | Date(s) | Stock | Life stage | Ovarian Fluid | Kidney/Spleen | Results |
| :---: | ---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 28-Oct, 4, 13-Nov | Priest Rapids | Adult | 150 | 60 | Negative |
| 1992 | 2,9-Nov | Priest Rapids | Adult | 150 | 60 | Negative |
| 1993 | 25-Oct, 1-Nov | Priest Rapids | Adult | 150 | 60 | Negative |
| 1994 | 7-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1995 | $9,13,19,21-N o v$ | Priest Rapids | Adult | 160 | 160 | Negative |
| 1996 | $17-N o v$ | Priest Rapids | Adult | 60 | 60 | Negative |
| 1997 | 17-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1998 | 16-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 1999 | 8-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2000 | 13-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2001 | 13-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2002 | 13-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2003 | 17-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2004 | 8-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2005 | 14-Nov | Priest Rapids | Adult | 60 | 60 | Negative |


| Year | Date(s) | Stock | Life stage | Ovarian Fluid | Kidney/Spleen | Results |
| :---: | ---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 6-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2007 | 5-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2008 | 3-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2009 | 2-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2010 | 15-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2011 | 7,14, 21-Nov | Priest Rapids | Adult | 180 | 180 | Negative |
| 2012 | 5-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2013 | 18-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2014 | 18-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2015 | 11-Nov | Priest Rapids | Adult | 60 | 60 | Negative |
| 2016 | 8-Nov | Priest Rapids | Adult | 60 | 60 | Negative |

Table 20 ELISA test results to determine risk of bacterial kidney disease of adult female fall Chinook salmon broodstock at Priest Rapids Hatchery, Return Years 2008-2016.

| Year | Stock |  | \%Below-Low | \% Low | \% Mod | \% High |
| :--- | :--- | ---: | :---: | :---: | :---: | :---: |
| 2008 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2009 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2010 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2011 | Priest Rapids | 135 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2012 | Priest Rapids | 60 | $98.3 \%$ | $0.0 \%$ | $1.7 \%$ | $0.0 \%$ |
| 2013 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2014 | Priest Rapids | 60 | $100.0 \%$ | $0.0 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2015 | Priest Rapids | 60 | $98.3 \%$ | $1.7 \%$ | $0.0 \%$ | $0.0 \%$ |
| 2016 | Priest Rapids | 60 | $98.3 \%$ | $1.7 \%$ | $0.0 \%$ | $0.0 \%$ |

### 10.1 Juvenile Fish Health Inspections

Juvenile fish are inspected for the presence of pathogens and other conditions on a monthly basis following ponding (AFS-FHS 2014). The 2016 brood year juveniles were generally healthy throughout the rearing period (Table 21). Inspection results for brood years 1995 through 2009 are provided in Appendix D.
Table 21 Juvenile fish health inspections for Priest Rapids Hatchery fall Chinook salmon, Brood Years 2006-2016.

| Date | Stock | Brood <br> Year | Condition |
| :--- | :--- | :--- | :--- |
| 18-Feb-10 | Priest Rapids | 2009 | Coagulated Yolk Syndrome observed in some fish sampled |
| 1-Apr-10 | Priest Rapids | 2009 | Healthy |
| 19-May-10 | Priest Rapids | 2009 | Healthy |
| 25-Mar-11 | Priest Rapids | 2010 | Healthy |
| 18-Apr-11 | Priest Rapids | 2010 | Healthy |
| 06-Jun-11 | Priest Rapids | 2010 | Healthy |
| 01-Mar-12 | Priest Rapids | 2011 | Healthy |
| 26-Apr-12 | Priest Rapids | 2011 | Healthy |
| 24-May-12 | Priest Rapids | 2011 | Healthy |
| 11-Feb-13 | Priest Rapids | 2012 | Healthy |
| 3-Mar-13 | Priest Rapids | 2012 | Healthy |


| Date | Stock | Brood <br> Year | Condition |
| :--- | :--- | :--- | :--- |
| 29-Apr-13 | Priest Rapids | 2012 | Healthy |
| 28-May-13 | Priest Rapids | 2012 | Healthy |
| 27-Mar-14 | Priest Rapids | 2013 | Dropout Syndrome present |
| 23-Apr-14 | Priest Rapids | 2013 | Dropout Syndrome present |
| 29-May-14 | Priest Rapids | 2013 | Healthy |
| 26-Feb-15 | Priest Rapids | 2014 | Coagulated Yolk Syndrome observed in some fish sampled |
| 26-Mar-15 | Priest Rapids | 2014 | Healthy |
| 21-Apr-15 | Priest Rapids | 2014 | Healthy |
| 28-May-15 | Priest Rapids | 2014 | Healthy |
| 22-June-15 | Priest Rapids | 2014 | Columnaris present in some fish sampled from Channel Pond B. |
| 24-Feb-16 | Priest Rapids | 2015 | Healthy |
| 15-Mar-16 | Priest Rapids | 2015 | Coagulated Yolk Syndrome observed in some fish sampled |
| 15-June-16 | Priest Rapids | 2015 | Mild Ich infection but healthy and ready for release |
| 24-Feb-17 | Priest Rapids | 2016 | Presence of bacterial gill disease in Raceway Ponds D and E |
| 21-Mar-17 | Priest Rapids | 2016 | Presence of bacterial gill disease in Raceway Pond B2 |
| 6-June-17 | Priest Rapids | 2016 | Mild Ich infection in Channel Ponds A, B, C |

### 11.0 Redd Survey

Fall Chinook salmon redd surveys were performed in the Hanford Reach during 2016 by staff with Environmental Assessment Services, LLC under contract with Mission Support Alliance. WDFW M\&E staff performed fall Chinook salmon redd surveys in the PRH discharge channel during 2016.

### 11.1 Hanford Reach Aerial Redd Counts

Aerial redd counts in the Hanford Reach were performed by Mission Support Alliance on October 23, November 6, 13, and 20 during 2016 (Nugent 2017). Redd counts should be considered an index of the total number of redds in the Hanford Reach. Redds may not be visible during flights due to wind, turbidity, ambient light, and depth. The surveys occurred on Sundays when outflows at Priest Rapids Dam were lowered to nearly 40 kcfs in conjunction with the Vernita Bar Settlement Agreement surveys performed by Grant PUD and WDFW. It is reported that viewing conditions during the surveys were generally good; however, a cloudy plume stretched along the eastern shore from the White Bluffs down to Richland during the last three flights (USDOE In Press). The peak fall Chinook Salmon redd count for the Hanford Reach in 2016 was 13,268 (Table 22).

## Table 22 Summary of fall Chinook salmon peak redd counts for the 1948-2016 aerial

 surveys in the Hanford Reach, Columbia River.| Year | Redds | Year | Redds | Year | Redds | Year | Redds |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1948 | 787 | 1967 | 3,267 | 1986 | 8,291 | 2005 | 7,891 |
| 1949 | 313 | 1968 | 3,560 | 1987 | 8,616 | 2006 | 6,508 |
| 1950 | 265 | 1969 | 4,508 | 1988 | 8,475 | 2007 | 4,023 |
| 1951 | 297 | 1970 | 3,813 | 1989 | 8,834 | 2008 | 5,588 |
| 1952 | 528 | 1971 | 3,600 | 1990 | 6,506 | 2009 | 4,996 |
| 1953 | 139 | 1972 | 876 | 1991 | 4,939 | 2010 | 8,817 |
| 1954 | 160 | 1973 | 2,965 | 1992 | 4,926 | 2011 | 8,915 |


| Year | Redds | Year | Redds | Year | Redds | Year | Redds |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1955 | 60 | 1974 | 728 | 1993 | 2,863 | 2012 | 8,368 |
| 1956 | 75 | 1975 | 2,683 | 1994 | 5,619 | 2013 | 17,398 |
| 1957 | 525 | 1976 | 1,951 | 1995 | 3,136 | 2014 | 15,951 |
| 1958 | 798 | 1977 | 3,240 | 1996 | 7,618 | 2015 | 20,678 |
| 1959 | 281 | 1978 | 3,028 | 1997 | 7,600 | 2016 | 13,268 |
| 1960 | 258 | 1979 | 2,983 | 1998 | 5,368 |  |  |
| 1961 | 828 | 1980 | 1,487 | 1999 | 6,068 |  |  |
| 1962 | 1,051 | 1981 | 4,866 | 2000 | 5,507 |  |  |
| 1963 | 1,254 | 1982 | 4,988 | 2001 | 6,248 |  |  |
| 1964 | 1,477 | 1983 | 5,290 | 2002 | 8,083 |  |  |
| 1965 | 1,789 | 1984 | 7,310 | 2003 | 9,465 |  | 8,468 |

### 11.2 Redd Distribution

The main spawning areas observed during the 2016 counts were located near Vernita Bar and among Islands 4-6 (Table 23 \& Figure 8). Historical redd counts by location from 2001 through 2016 are included in Appendix E of this report.
Table 23 Number of fall Chinook salmon redds counted in difference reaches on the Hanford Reach area of the Columbia River during October 2016 through November 2016 aerial redd counts. (Data provided by Mission Support Alliance).

| General Location | $\begin{gathered} \text { Start } \\ \text { KM } \\ \hline \end{gathered}$ | End <br> KM | Total Length | 23-Oct | 6-Nov | 13-Nov | 20-Nov | Max Count | Avg. Redd Per River KM |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Islands 17-21 | 545 | 558 | 13 | 0 | 0 | 0 | 0 | 0 | 0 |
| Islands 11-16 | 558 | 573 | 15 | 0 | 380 | 830 | 861 | 861 | 57 |
| Islands 8-10 | 587 | 593 | 6 | 35 | 1,020 | 1,685 | 1,735 | 1,735 | 289 |
| Near Island 7 | 593 | 594 | 1 | 0 | 650 | 660 | 670 | 670 | 670 |
| Island 6 (lower half) | 594 | 599 | 5 | 54 | 1,135 | 1,805 | 1,807 | 1,807 | 361 |
| Island 4, 5 and upper 6 | 599 | 602 | 3 | 68 | 2,140 | 2,262 | 2,270 | 2,270 | 757 |
| Near Island 3 | 602 | 604 | 2 | 30 | 380 | 550 | 600 | 600 | 300 |
| Near Island 2 | 604 | 606 | 2 | 40 | 810 | 1,120 | 1,140 | 1,140 | 570 |
| Near Island 1 | 606 | 608 | 2 | 10 | 253 | 300 | 340 | 340 | 170 |
| Near Coyote Rapids | 608 | 619 | 11 | 13 | 185 | 252 | 255 | 255 | 23 |
| Midway (China Bar) | 620 | 630 | 10 | 4 | 60 | 60 | 80 | 80 | 8 |
| Near Vernita Bar | 630 | 635 | 5 | 220 | 3,140 | 3,400 | 3,500 | 3,500 | 700 |
| Near Priest Rapids Dam | 635 | 638 | 3 | 0 | 7 | 10 | 10 | 10 | 3 |
| Total | -- | -- | -- | 474 | 10,160 | 12,934 | 13,268 | 13,268 |  |



Figure 8 Distribution of fall Chinook salmon redd counts by location for the 2016 aerial surveys in the Hanford Reach, Columbia River (Data provided by Mission Support Alliance).

### 11.3 Spawn Timing

Based on aerial redd counts and Vernita Bar spawning ground surveys, fall Chinook salmon spawning in the Hanford Reach during 2016 began in late October and ended in late November. River temperatures below Priest Rapids Dam varied from $15.4^{\circ} \mathrm{C}$ (October 20) to $12.3^{\circ} \mathrm{C}$ (November 20) during the spawning period which is typical to that of previous years.

### 11.4 Escapement

The estimated total escapement of fall Chinook salmon to the Hanford Reach for the 2016 return year was 116,388 fish (Table 24). This is lower than recent annual returns but still the fourth highest return since 1991 (Table 25). The historical mean and median escapement for 1991 through 2016 is 73,550 and 55,208 fish, respectively. The estimated adult Chinook salmon per redd is calculated by dividing the adult escapement to the Hanford Reach by peak number of redds reported in the redd survey. The estimated annual escapements to the Hanford Reach were not adjusted for pre-spawn mortality. For 2016, the estimated 9 fish per redd was the same as the historical mean.

Table 24 Calculation of escapement estimates for fall Chinook salmon in the Hanford Reach, Columbia River 2016.

|  | Return Year 2016 |  |  |
| :--- | ---: | ---: | ---: |
| Count Source | Adult | Jack | Total |
| McNary Ladder Counts | 239,791 | 24,870 | 264,661 |
| Adjusted Priest Rapids Adult Passage $^{1}$ | 35,464 | 2,759 | 38,223 |
| Ice Harbor Adult Passage | 36,713 | 13,066 | 49,779 |
| Prosser Adult Passage | 5,214 | 464 | 5,678 |
| Priest Rapids Hatchery | 27,904 | 882 | 28,786 |
| PRH discharge channel | 44 | 0 | 44 |


| Count Source | Return Year 2016 |  |  |
| :--- | ---: | ---: | ---: |
| Jack | Total |  |  |
| Wanapum Tribal Fishery | Adult | 1 | 36 |
| Ringold Springs Hatchery | 5,314 | 65 | 5,379 |
| Yakima River Escapement (Below Prosser) | 1,087 | 97 | 1,184 |
| Yakima River Sport Harvest | 922 | 31 | 953 |
| Hanford Sport Harvest | 16,859 | 1,068 | 17,927 |
| Angler Broodstock Collection | 284 | 0 | 284 |
| Total Non-Hanford Reach Escapement | $\mathbf{1 2 9 , 8 4 0}$ | $\mathbf{1 8 , 4 3 3}$ | $\mathbf{1 4 8 , 2 7 3}$ |
| Hanford Reach Escapement | $\mathbf{1 0 9 , 9 5 1}$ | $\mathbf{6 , 4 3 7}$ | $\mathbf{1 1 6 , 3 8 8}$ |

${ }^{1}$ Gross passage count reduced $13.53 \%$ to correct for estimated over counts resulting from fallbacks and reascension. The adjustments to adult fish passage were estimated by analysis of the PIT tag detections at PIT tag arrays located in the adult fish ways of the Priest Rapids Dam adult fishway and the discharge channel for Priest Rapids Hatchery.

Table 25 Escapement for fall Chinook salmon in the Hanford Reach, Return Years 1991-2016.

| Return Year | \# Fish per Redd | Redds | Total Escapement $^{\mathbf{1}}$ |
| :---: | :---: | :---: | :---: |
| 1991 | 11 | 4,939 | 52,196 |
| 1992 | 9 | 4,926 | 41,952 |
| 1993 | 13 | 2,863 | 37,347 |
| 1994 | 11 | 5,619 | 63,103 |
| 1995 | 18 | 3,136 | 55,208 |
| 1996 | 6 | 7,618 | 43,249 |
| 1997 | 6 | 7,600 | 43,493 |
| 1998 | 7 | 5,368 | 35,393 |
| 1999 | 5 | 6,068 | 29,812 |
| 2000 | 9 | 5,507 | 48,020 |
| 2001 | 10 | 6,248 | 59,848 |
| 2002 | 10 | 8,083 | 84,509 |
| 2003 | 9 | 9,465 | 100,508 |
| 2004 | 10 | 8,468 | 87,696 |
| 2005 | 9 | 7,891 | 71,967 |
| 2006 | 8 | 6,508 | 51,701 |
| 2007 | 6 | 4,018 | 22,272 |
| 2008 | 5 | 5,618 | 29,058 |
| 2009 | 7 | 4,996 | 36,720 |
| 2010 | 10 | 8,817 | 87,016 |
| 2011 | 8 | 8,915 | 75,256 |
| 2012 | 7 | 8,368 | 57,710 |
| 2013 | 10 | 17,398 | 174,651 |
| 2014 | 12 | 15,951 | 183,749 |
| 2015 | 9 | 20,678 | 266,327 |
| 2016 | $\mathbf{9}$ | 13,268 | 116,388 |
| Mean | $\mathbf{9}$ | $\mathbf{8 , 0 1 3}$ | $\mathbf{7 3 , 5 5 0}$ |
| Median |  |  | $\mathbf{5 5 , 2 0 8}$ |

[^3]
### 11.5 Hatchery Discharge Channel Redd Counts

The M\&E staff conducted redd counts in the PRH discharge channel on October 29, November 9, November 21, and December 2, 2016. Similar to historical observations, the majority of spawning activity was located in a 200 meter section of the discharge channel downstream adjacent to the volunteer trap. A peak count of 22 redds occurred on the December 5 survey. We observed superimposition occurring during multiple surveys; thus making it difficult to determine the total number of redds in a given survey. Viewing conditions during each survey were good to excellent.

### 12.0 Carcass Surveys

Prior to 2010, the carcass surveys in the Hanford Reach were generally performed by two boat crews of two staff operating seven days a week. Beginning in 2010, with support of the PRH M\&E Program, the effort was increased to three boats with a three-person crew operating seven days per week. The extra staffing was necessary to maintain the overall sampling efficiency given the additional effort required to pull otoliths from fish sampled and achieve hatchery M\&E objectives. The sampling goal for obtaining sufficient number of CWTs is $10 \%$ of the escapement. The recent record returns to the Hanford Reach have necessitated an increased level of effort to attain the $10 \%$ sampling goal.
Carcass surveys were performed from November 3 through December 9, 2016. All recovered carcasses were sampled for the presence of a CWT. Of those, $20 \%$ were sampled (i.e., random systematic 1:5 rate) for scales (age), otoliths, gender, length, and egg retention. All carcasses recovered were chopped in half after sampling to prevent the chance of double sampling.

Similar to methods used since 2010, the carcass survey crews recorded the sections in which carcasses were recovered in the Hanford Reach and adjacent areas. The Hanford Reach survey is divided into Sections 1 through 5 (Figure 9). The Priest Rapids Pool is designated as Section 6. The PRH discharge channel and the area of the Columbia River immediately below the discharge channel are designated as Sections 7 and 8, respectively. The fall Chinook salmon carcasses recovered in Section 8 were likely wash outs from the hatchery discharge channel.

- Section 1. Priest Rapids Dam to Vernita Bridge (14 km)
- Section 2. Vernita Bridge to Island 2 (19 km)
- Section 3. Island 2 to Powerline Towers at Hanford town site (21 km)
- Section 4. Power line Towers to Wooded Island (21 km)
- Section 5. Wooded Island to Interstate 182 Bridge (19 km)
- Section 6. Priest Rapids Pool (34 km)
- Section 7. Priest Rapids Hatchery discharge channel ( 0.5 km )
- Section 8. Columbia River at the mouth of the Hatchery discharge channel ( 0.5 km )


Figure 9 Location of aerial redd index areas and river surveys section in the Hanford Reach.

### 12.1 Hanford Reach Carcass Survey: Section 1 - 5

Staff recovered a record 8,886 fall Chinook salmon in the Hanford Reach in 2016; equating to $7.6 \%$ of the estimated fall Chinook salmon escapement (Table 26). The recovery rate was lower than expected due to unusually high river flows which regularly made it difficult to locate carcasses. The annual number of fall Chinook salmon carcass recovered in the Hanford Reach for the period of 1991 through 2016 is provided in Appendix F.
Table 26 Numbers and proportions of total escapement of fall Chinook salmon carcasses surveyed for coded-wire tags within each survey section on the Hanford Reach, Return Years, 2010-2016.

| Return Year | \# 1 |  | \# 2 |  | \# 3 |  | \# 4 |  | \# 5 |  | Total Sampled |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | $\mathbf{P}^{\wedge}$ | N | $\mathbf{P}^{\wedge}$ | N | $\mathbf{P}^{\wedge}$ | N | $\mathbf{P}^{\wedge}$ | N | $\mathbf{P}^{\wedge}$ | N | $\mathbf{P}^{\wedge}$ | Escapement |
| 2010 | 1,832 | 0.021 | 519 | 0.006 | 3,129 | 0.036 | 3,362 | 0.039 | 937 | 0.011 | 9,779 | 0.112 | 87,016 |
| 2011 | 1,581 | 0.021 | 160 | 0.002 | 2,606 | 0.035 | 2,622 | 0.035 | 1,422 | 0.019 | 8,391 | 0.111 | 75,256 |
| 2012 | 1,091 | 0.019 | 149 | 0.003 | 1,685 | 0.029 | 2,213 | 0.038 | 1,676 | 0.029 | 6,814 | 0.118 | 57,715 |
| 2013 | 2,182 | 0.012 | 1,973 | 0.011 | 2,844 | 0.016 | 3,774 | 0.022 | 2,298 | 0.013 | 13,071 | 0.075 | 174,651 |
| 2014 | 2,682 | 0.015 | 1,142 | 0.006 | 5,544 | 0.030 | 4,573 | 0.025 | 2,815 | 0.015 | 16,756 | 0.091 | 183,680 |
| 2015 | 2,913 | 0.011 | 823 | 0.003 | 6,187 | 0.023 | 5,868 | 0.022 | 1,947 | 0.007 | 17,738 | 0.067 | 266,346 |
| 2016 | 1,141 | 0.010 | 513 | 0.004 | 2,796 | 0.024 | 2,977 | 0.026 | 1,459 | 0.013 | 8,886 | 0.076 | 116,388 |
| Mean | 1,917 | 0.016 | 754 | 0.005 | 3,542 | 0.028 | 3,627 | 0.030 | 1,793 | 0.015 | 11,634 | 0.093 | 140,777 |

The survey effort was not equal for each section. Sections 3 and 4 were surveyed the most because these sections generally contain the largest number of carcasses (Table 27). As each
season progresses, crews focused their effort in sections which provided greater chances to recover carcasses.

Table 27 Number of carcass surveys conducted by section in the Hanford Reach, Return Years 2010-2016.

| Return Year | $\mathbf{\# 1}$ | $\mathbf{\# 2}$ | $\mathbf{\# 3}$ | $\boldsymbol{\# 4}$ | \# 5 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| 2010 | 21 | 6 | 26 | 26 | 11 | 90 |
| 2011 | 33 | 5 | 38 | 29 | 13 | 118 |
| 2012 | 19 | 4 | 26 | 28 | 24 | 101 |
| 2013 | 18 | 15 | 16 | 17 | 13 | 79 |
| 2014 | 23 | 17 | 30 | 31 | 24 | 125 |
| 2015 | 23 | 8 | 35 | 37 | 13 | 116 |
| 2016 | 18 | 11 | 29 | 27 | 15 | 100 |
| Mean | $\mathbf{2 2}$ | $\mathbf{9}$ | $\mathbf{2 9}$ | $\mathbf{2 8}$ | $\mathbf{1 6}$ | $\mathbf{1 0 4}$ |

12.1.1 Proportion of Escapement Sampled: Section 1-5

The spawning escapement for sections 1 through 5 was estimated by the proportion of redds counted in aerial surveys to the estimated escapement of natural spawners to the Hanford Reach (see Section 14 - Redd Surveys). The calculations for estimating the escapement to the Hanford Reach are given in Appendix G.
We recently identified through the carcass bias assessment that an unknown number of carcasses drift into downstream sections after spawning. The recovery of these carcasses may confound the estimate of the spawning escapement sampled by section as shown in Table 28. For example, there were no redds identified in Section 5 but 1,459 carcasses were recovered in that section. It is likely that sections 1 and 3 which have the greatest number of redds and largest spawning escapement end up with a net loss of carcasses to downstream sections. In 2016, we continued a pilot study to evaluate the magnitude and distribution of post spawn carcass drift. The preliminary results of this study are included in the Appendix H .
Table 28 Number of redds and carcasses, total spawning escapement, and proportion of escapement sampled for fall Chinook salmon in Section 1 through 5 of the Hanford Reach, Return Year 2016.

| Survey <br> Section | Total Number of <br> Redds | Total Number of <br> Carcasses | Spawning Escapement ${ }^{\mathbf{1}}$ | Proportion of <br> Escapement Sampled |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 3,590 | 1,141 | 31,492 | 0.036 |
| 2 | 595 | 513 | 5,219 | 0.098 |
| 3 | 8,222 | 2,796 | 72,124 | 0.039 |
| 4 | 861 | 2,977 | 7,553 | 0.394 |
| 5 | 0 | 1,459 | 0 | 0.000 |
| Total | $\mathbf{1 3 , 2 6 8}$ | $\mathbf{8 , 8 8 6}$ | $\mathbf{1 1 6 , 3 8 8}$ | 0.076 |

${ }^{1}$ Calculated based on proportion of redds by section

### 12.1.2 Carcass Distribution and Origin

Two methods were used to estimate the origin of carcasses recovered in the sections 1 through 5. The first method includes the expansion of pooled CWT recoveries using juvenile tag rates and survey sample rate. The second method includes calculating the proportion of combined hatchery marks (i.e., otolith mark, adipose clips, and CWTs) to non-marked carcasses. Estimates for both
methods are given for the 2012-2016 adult returns: these years include otolith marks for all ages of PRH origin fish.

The assumption was made that all Chinook salmon not accounted by hatchery origin CWT expansions were of natural origin. This assumption may underestimate the number of hatchery carcasses recovered in the annual surveys. We have compelling evidence to suggest this is the case with annual returns to PRH. The expansion of CWT recoveries suggest that $8.7 \%$ of the fall Chinook salmon carcasses recovered in the 2016 Hanford Reach stream surveys were hatchery origin (Table 29). This estimate is similar to those of previous years excluding 2013. The expanded CWT recovery data suggest the hatchery origin component of the escapement included $6.9 \%$ from PRH, $1.2 \%$ from RSH and $0.6 \%$ from other hatcheries. The highest proportions of hatchery origin carcasses recovered were in Sections 1, and 3, respectively.
Table 29 Numbers of natural and hatchery origin fall Chinook salmon carcasses sampled within Sections 1 through 5 of Hanford Reach based on expansions of coded-wire tag recoveries, Return Years 2010-2016.

| Return Year | Hanford Reach Sections |  |  |  |  |  |  | Proportio n of |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Origin | \# 1 | \# 2 | \# 3 | \# 4 | \# 5 | Total |  |
| 2010 | Natural | 1,751 | 473 | 3,020 | 3,242 | 909 | 9,395 | 0.960 |
|  | Hatchery | 81 | 46 | 116 | 125 | 28 | 396 | 0.040 |
|  | Proportion | 0.044 | 0.089 | 0.037 | 0.037 | 0.030 | 0.040 |  |
| 2011 | Natural | 1,350 | 155 | 2,520 | 2,475 | 1,347 | 7,847 | 0.935 |
|  | Hatchery | 231 | 5 | 86 | 147 | 75 | 544 | 0.065 |
|  | Proportion | 0.146 | 0.031 | 0.033 | 0.056 | 0.053 | 0.065 |  |
| 2012 | Natural | 1,142 | 149 | 1,526 | 2,081 | 1,510 | 6,408 | 0.927 |
|  | Hatchery | 49 | 0 | 159 | 132 | 166 | 506 | 0.073 |
|  | Proportion | 0.041 | 0.000 | 0.094 | 0.060 | 0.099 | 0.073 |  |
| 2013 | Natural | 1,572 | 1,587 | 2,433 | 2,895 | 1,748 | 10,235 | 0.783 |
|  | Hatchery | 610 | 386 | 411 | 879 | 550 | 2,836 | 0.217 |
|  | Proportion | 0.280 | 0.196 | 0.145 | 0.233 | 0.239 | 0.217 |  |
| 2014 | Natural | 2,469 | 1,072 | 5,264 | 4,329 | 2,703 | 15,838 | 0.945 |
|  | Hatchery | 213 | 70 | 280 | 244 | 112 | 918 | 0.055 |
|  | Proportion | 0.079 | 0.061 | 0.050 | 0.053 | 0.040 | 0.055 |  |
| 2015 | Natural | 2,654 | 709 | 5,745 | 5,490 | 1,858 | 16,456 | 0.928 |
|  | Hatchery | 259 | 114 | 442 | 378 | 89 | 1,282 | 0.072 |
|  | Proportion | 0.089 | 0.139 | 0.071 | 0.064 | 0.046 | 0.072 |  |
| 2016 | Natural | 1,108 | 256 | 2,585 | 2,866 | 684 | 8,111 | 0.913 |
|  | Hatchery | 162 | 33 | 257 | 211 | 111 | 775 | 0.087 |
|  | Proportion | 0.142 | 0.064 | 0.092 | 0.071 | 0.076 | 0.087 |  |
| Mean | Natural | 1,721 | 629 | 3,299 | 3,340 | 1,537 | 10,613 | 0.913 |
|  | Hatchery | 229 | 93 | 250 | 302 | 162 | 1,037 | 0.087 |
|  | Proportion | 0.115 | 0.090 | 0.074 | 0.082 | 0.092 | 0.089 |  |

The second estimate of origin of carcasses recovered is based on the proportion of hatchery marked to non-marked fish. For this method, we assume that all hatchery origin carcasses recovered are marked in some manner (e.g., otolith marks, CWT, and adipose clips) and that we are able to accurately detect these marks and tags.

PRH has marked their entire juvenile releases with annual marks on the otoliths beginning with progeny of brood year 2007. For the 2013-2016 returns, age-2 through 6 PRH origin carcasses recovered were otolith marked. The age-6 PRH origin fish were not otolith marked during return year 2012. However, since there were no age-6 fish recovered in the carcass surveys or at PRH, it is assumed that few, if any PRH origin age-6 fish spawned in the Hanford Reach.

Adipose clipped Chinook salmon without a CWT and without a thermal otolith mark were classified as strays from other hatcheries. The natural origin fish were identified by either a Hanford Reach origin CWT or by the presence of an adipose fin and the absence of an otolith mark. The demographic sample data suggests that $11.4 \%$ of fall Chinook salmon carcasses recovered in the 2016 Hanford Reach stream survey were hatchery origin (Table 30). For recent years, the hatchery proportions were generally higher in the upstream survey sections.
Table 30 Origin of Chinook salmon carcasses recovered in the Hanford Reach by section based on recoveries of marked and unmarked carcasses within the biological sample, Return Years 2012-2016.

| Year | Origin | \# 1 | \# 2 | \# 3 | \# 4 | \# 5 | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $2012$ <br> Biological sample Rate 1:4 $\mathrm{N}=1,609$ | $\mathrm{PRH}^{1}$ | 23 | 2 | 26 | 18 | 38 | 107 | 0.067 |
|  | Other Hatchery ${ }^{2}$ | 10 | 2 | 25 | 45 | 22 | 104 | 0.065 |
|  | Total Hatchery | 33 | 4 | 51 | 63 | 60 | 211 | 0.131 |
|  | Natural ${ }^{3}$ | 228 | 30 | 347 | 460 | 333 | 1,398 | 0.869 |
|  | Proportion Hatchery | 0.126 | 0.118 | 0.128 | 0.120 | 0.153 | 0.131 |  |
| $2013^{a}$ <br> Biological sample rate $=1: 5$ and then randomly subsampled, $\mathrm{N}=712$ | PRH ${ }^{1}$ | 32 | 19 | 34 | 30 | 32 | 147 | 0.206 |
|  | Other Hatchery ${ }^{2}$ | 6 | 3 | 16 | 21 | 6 | 52 | 0.073 |
|  | Total Hatchery | 38 | 22 | 50 | 51 | 38 | 199 | 0.279 |
|  | Natural ${ }^{3}$ | 76 | 84 | 113 | 155 | 85 | 513 | 0.721 |
|  | Proportion Hatchery | 0.333 | 0.208 | 0.307 | 0.248 | 0.309 | 0.279 |  |
| $2014^{a}$ <br> Biological sample rate $=1: 5$ and then randomly subsampled, $\mathrm{N}=$ 2,426 | PRH ${ }^{1}$ | 37 | 7 | 45 | 35 | 11 | 135 | 0.056 |
|  | Other Hatchery ${ }^{2}$ | 12 | 5 | 16 | 32 | 18 | 83 | 0.034 |
|  | Total Hatchery | 49 | 12 | 61 | 67 | 29 | 218 | 0.090 |
|  | Natural ${ }^{3}$ | 347 | 142 | 711 | 612 | 396 | 2208 | 0.910 |
|  | Proportion Hatchery | 0.124 | 0.078 | 0.079 | 0.099 | 0.068 | 0.090 |  |
| $2015$$\begin{gathered} \text { Biological sample } \\ \text { rate }=1: 7 \\ \mathrm{~N}=2,485 \end{gathered}$ | $\mathrm{PRH}^{1}$ | 47 | 12 | 61 | 55 | 13 | 188 | 0.076 |
|  | Other Hatchery ${ }^{2}$ | 6 | 2 | 17 | 20 | 7 | 52 | 0.021 |
|  | Total Hatchery | 53 | 14 | 78 | 75 | 20 | 240 | 0.097 |
|  | Natural ${ }^{3}$ | 346 | 101 | 792 | 752 | 254 | 2,245 | 0.903 |
|  | Proportion Hatchery | 0.133 | 0.122 | 0.090 | 0.091 | 0.073 | 0.097 |  |
| $2016$ <br> Biological sample $\begin{aligned} & \text { rate }=1: 5 \\ & \mathrm{~N}=1,743 \end{aligned}$ | PRH ${ }^{1}$ | 27 | 12 | 42 | 22 | 10 | 113 | 0.066 |
|  | Other Hatchery ${ }^{2}$ | 9 | 6 | 31 | 23 | 13 | 82 | 0.048 |
|  | Total Hatchery | 36 | 18 | 73 | 45 | 23 | 195 | 0.114 |
|  | Natural ${ }^{3}$ | 182 | 80 | 465 | 534 | 257 | 1,518 | 0.886 |
|  | Proportion Hatchery | 0.165 | 0.184 | 0.136 | 0.078 | 0.082 | 0.114 |  |
| Mean Proportion | PRH ${ }^{1}$ | 0.147 | 0.11 | 0.089 | 0.07 | 0.094 | 0.094 | 0.094 |
|  | Other Hatchery ${ }^{2}$ | 0.031 | 0.036 | 0.038 | 0.05 | 0.044 | 0.042 | 0.048 |


| Year | Origin | \# 1 | \# 2 | \# 3 | \# 4 | \# 5 | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | All Hatchery | 0.178 | 0.145 | 0.127 | 0.12 | 0.139 | 0.136 | 0.142 |
|  | Natural | 0.849 | 0.862 | 0.886 | 0.893 | 0.886 | 0.881 | 0.858 |
|  | Proportion | 0.176 | 0.142 | 0.148 | 0.127 | 0.137 | 0.142 |  |

${ }^{\text {a }}$ Estimate of origin based on random sub-sample of biological sample.
${ }^{1}$ Priest Rapids Hatchery fish were identified by either the presence of thermal otolith mark or by the presence of a PRH origin coded-wire tag
${ }^{2}$ Other hatchery strays were identified as adipose clipped Chinook salmon without a Priest Rapids Hatchery codedwire tag and without a thermal otolith mark or by the presence of other hatchery coded-wire tags.
${ }^{3}$ Natural origin fish were identified by either a Hanford Reach origin coded-wire tag or by the presence of an adipose fin and the absence of an otolith mark.

### 12.2 Priest Rapids Dam Pool Carcass Survey: Section 6

In total, eight carcass surveys were performed in Section 6 during return year 2016 (Table 31). Surveys were scheduled once or twice a week between early November and mid-December.
Table 31 Number of fall Chinook salmon carcasses sampled within Section 6 (Priest Rapids Dam Pool).

|  | Section 6 |  |
| :---: | :---: | :---: |
| Year | \# of Carcasses | \# of Surveys |
| 2010 | 123 | 8 |
| 2011 | 69 | 7 |
| 2012 | 72 | 4 |
| 2013 | 407 | 7 |
| 2014 | 237 | 7 |
| 2015 | 155 | 6 |
| 2016 | 139 | 8 |
| Mean | $\mathbf{1 7 2}$ | $\mathbf{7}$ |

### 12.2.1 Number sampled: Section 6

Survey crews recovered 139 Chinook salmon in Section 6 during return year 2016 (Table 31). All fish recovered were scanned for the presence of a CWT. Carcass recoveries in the lower portion of the pool suggest that carcasses drift downstream of the spawning areas below Wanapum Dam into deeper water where they are difficult to recover.

### 12.2.2 Proportion of Escapement Sampled: Section 6

The spawning escapement for Section 6 was calculated by subtracting from the Priest Rapids Dam fall Chinook salmon passage count, the fall Chinook salmon passage at Wanapum Dam, tribal and sport harvest of fall Chinook salmon in the Priest Rapids Dam pool, and the estimated fallback of fall Chinook salmon at Priest Rapids Dam (Appendix G).

The 2016 fall Chinook salmon spawning escapement estimate for Section 6 is 13,162 fish. Overall, roughly $1 \%$ of the total estimated spawning escapement in Section 6 was sampled in 2016 (Table 32).

Table 32 Carcasses sampled, total spawning escapement and proportion of escapement for fall Chinook salmon in Section 6 (Priest Rapids Dam Pool), return years 2010-2016.

| Return Year | \# of Surveys | \# of Carcasses | Spawning Escapement | Escapement Sampled |
| :---: | :---: | :---: | :---: | :---: |
| 2010 | 8 | 123 | 11,121 | 0.011 |
| 2011 | 7 | 69 | 11,362 | 0.006 |
| 2012 | 4 | 72 | 21,919 | 0.003 |
| 2013 | 7 | 407 | 62,237 | 0.007 |
| 2014 | 7 | 237 | 25,179 | 0.009 |
| 2015 | 6 | 155 | 38,313 | 0.004 |
| 2016 | 8 | 139 | 13,162 | 0.011 |
| Mean | $\mathbf{7}$ | $\mathbf{1 7 2}$ | $\mathbf{2 6 , 1 8 5}$ | $\mathbf{0 . 0 0 7}$ |

### 12.2.3 Carcass Origin: Section 6

Similar to those methods described in detail in the previous section, the carcasses included in the 1:1 demographic sample were identified as hatchery origin based on a combination of hatchery marks and tags (i.e., otoliths marks, adipose clips, and CWTs). Natural origin carcasses were identified by the absence of any hatchery mark or the presence of a natural origin CWT.
An estimated $49.6 \%$ of the fall Chinook salmon spawning in Section 6 were hatchery origin of which $95.7 \%$ were PRH origin (Table 33).
Table 33 Origin of fall Chinook salmon spawning in Section 6 (Priest Rapids Dam Pool), Return Years 2012-2016.

| Year | Origin | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: |
| $\begin{gathered} 2012 \\ \mathrm{~N}=70 \end{gathered}$ | PRH ${ }^{1}$ | 18 | 0.257 |
|  | Other Hatchery ${ }^{2}$ | 2 | 0.029 |
|  | Total Hatchery | 20 | 0.286 |
|  | Natural ${ }^{3}$ | 50 | 0.714 |
| $\begin{gathered} 2013 \\ \mathrm{~N}=98 \end{gathered}$ | PRH ${ }^{1}$ | 62 | 0.633 |
|  | Other Hatchery ${ }^{2}$ | 5 | 0.051 |
|  | Total Hatchery | 67 | 0.684 |
|  | Natural ${ }^{3}$ | 31 | 0.316 |
| $\begin{array}{r} 2014 \\ \mathrm{~N}=229 \end{array}$ | $\mathrm{PRH}^{1}$ | 81 | 0.354 |
|  | Other Hatchery ${ }^{2}$ | 5 | 0.022 |
|  | Total Hatchery | 86 | 0.376 |
|  | Natural ${ }^{3}$ | 143 | 0.624 |
| $\begin{gathered} 2015 \\ \mathrm{~N}=155 \end{gathered}$ | PRH ${ }^{1}$ | 83 | 0.535 |
|  | Other Hatchery ${ }^{2}$ | 3 | 0.019 |
|  | Total Hatchery | 155 | 0.555 |
|  | Natural ${ }^{3}$ | 69 | 0.445 |
| $\begin{gathered} 2016 \\ \mathrm{~N}=134 \end{gathered}$ | PRH ${ }^{1}$ | 66 | 0.475 |
|  | Other Hatchery ${ }^{2}$ | 3 | 0.022 |
|  | Total Hatchery | 69 | 0.496 |
|  | Natural ${ }^{3}$ | 65 | 0.468 |
|  | PRH ${ }^{1}$ | 62 | 0.411 |


| Year | Origin | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: |
| Means | Other Hatchery $^{2}$ | $\mathbf{4}$ | $\mathbf{0 . 0 2 6}$ |
|  | Total Hatchery $^{2}$ | 79 | $\mathbf{0 . 5 2 3}$ |
|  | Natural $^{3}$ | $\mathbf{7 2}$ | $\mathbf{0 . 4 7 7}$ |

${ }^{1}$ Priest Rapids Hatchery fish were identified by either the presence of thermal otolith mark or by the presence of a PRH origin coded-wire tag
${ }^{2}$ Other hatchery strays were identified as adipose clipped Chinook salmon without a Priest Rapids Hatchery codedwire tag and without a thermal otolith mark.
${ }^{3}$ Natural origin fish were identified by either a Hanford Reach origin coded-wire tag or by the presence of an adipose fin and the absence of an otolith mark.

### 12.3 Hatchery Discharge Channel: Sections 7 and 8 Carcass Survey

During return year 2016, crews performed one carcass surveys in Section 8 by boat and one carcass survey in Section 7 by foot. It has been observed that many carcasses drift out of the discharge channel under full flow conditions. Performing carcass surveys in the discharge channel when it is at full flow is difficult and dangerous due to poor footing and high velocities. Staff performed the one survey in Section 7 on November 28 when discharge levels in the channel were still high.

### 12.3.1 Number sampled: Sections 7 and 8

Survey crews recovered 33 carcasses in Section 7 and 26 in Section 8 (Table 34). All fish recovered were scanned for the presence of a CWT.

Table 34 The number of fall Chinook salmon carcass surveys within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at the confluence of the hatchery discharge channel), Return Years 2010-2016.

| Return Year | Section 7 |  | Section 8 |  | Total |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \# of Carcasses | \# of Surveys | \# of Carcasses | \# of Surveys | \# of Carcasses | \# of Surveys |
| 2010 | 87 | 1 | 123 | 9 | 210 | 10 |
| 2011 | 123 | 2 | 80 | 8 | 203 | 10 |
| 2012 | 99 | 3 | 108 | 10 | 207 | 13 |
| 2013 | 105 | 3 | 159 | 4 | 264 | 7 |
| 2014 | 9 | 1 | 52 | 7 | 61 | 8 |
| 2015 | 33 | 1 | 26 | 2 | 59 | 3 |
| 2016 | 3 | 1 | 7 | 1 | 10 | 2 |
| Mean | 66 | 2 | 79 | 6 | 145 | 8 |

### 12.3.2 Proportion of Escapement Sampled: Section 7 and 8

The 2016 fall Chinook salmon spawning escapement index for Sections 7 and 8 is 44 fish (Table 35). The spawning escapement for these Sections was calculated using the expansion factor of two fish/redd, based on a 0.9 male/female sex ratio including jacks, as estimated from the Hanford Reach 2016 escapement. Therefore, the assumption is made that each of the 22 redds represents one female and one male. We assume that most of the carcasses recovered in Section 8 drifted downstream from Section 7. In addition, it is likely a portion of carcasses from Sections 7 and 8 drift downstream into Sections 1 and 2.

Table 35 Number of carcasses sampled, total spawning escapement and proportion of escapement sampled for fall Chinook salmon within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at confluence of the hatchery discharge channel), Return Year 2016.

| Section | Total Number of Carcasses | Spawning Escapement | Escapement Sampled |
| :---: | :---: | :---: | :---: |
| $\# 7$ | 3 | 44 | 0.227 |
| $\# 8$ | 7 | 0 |  |
| Total | $\mathbf{1 0}$ | $\mathbf{4 4}$ |  |

12.3.3 Carcass Distribution and Origin: Section 7 and 8

The demographic sample rate was set at 1:1 to account for the low numbers of carcasses recovered. As described in detail previously, the carcasses included the demographic sample were identified as hatchery origin based on a combination of hatchery marks and tags (i.e., otoliths marks, adipose clips, and CWTs). Natural origin carcasses were identified by the absence of any hatchery mark or the presence of a natural origin CWT.
It is estimated that $66.7 \%$ of fall Chinook salmon recovered in Sections 7 and 8 were hatchery origin of which most all were PRH origin (Table 36).
Table 36 The origin of Chinook salmon carcasses recovered within Section 7 (Priest Rapids Hatchery Discharge Channel) and Section 8 (Columbia River at the confluence of the hatchery discharge channel), Return Years 2012-2016.

| Return Year | Origin | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: |
| $\begin{gathered} 2012 \\ \mathrm{~N}=70 \end{gathered}$ | PRH | 18 | 0.257 |
|  | Other Hatchery | 2 | 0.029 |
|  | Total Hatchery | 20 | 0.286 |
|  | Natural | 50 | 0.714 |
| $\begin{gathered} 2013 \\ \mathrm{~N}=33 \end{gathered}$ | PRH | 28 | 0.848 |
|  | Other Hatchery | 2 | 0.061 |
|  | Total Hatchery | 30 | 0.909 |
|  | Natural | 3 | 0.091 |
| $\begin{aligned} & 2014 \\ & \mathrm{~N}=5 \end{aligned}$ | PRH | 3 | 0.600 |
|  | Other Hatchery | 0 | 0.000 |
|  | Total Hatchery | 3 | 0.600 |
|  | Natural | 2 | 0.400 |
| $\begin{aligned} & 2015 \\ & \mathrm{~N}=59 \end{aligned}$ | PRH | 19 | 0.322 |
|  | Other Hatchery | 2 | 0.034 |
|  | Total Hatchery | 21 | 0.356 |
|  | Natural | 38 | 0.644 |
| $\begin{aligned} & 2016 \\ & \mathrm{~N}=6 \end{aligned}$ | PRH | 4 | 0.667 |
|  | Other Hatchery | 1 | 0.167 |
|  | Total Hatchery | 5 | 0.833 |
|  | Natural | 1 | 0.167 |
| Means | PRH | 14 | 0.416 |
|  | Other Hatchery | 1 | 0.058 |


| Return <br> Year | Origin | Total | Proportion of Sample |
| :---: | :---: | :---: | :---: |
| $=35$ | Total Hatchery | 16 | 0.597 |
|  | Natural | 19 | 0.403 |

### 13.0 Life History Monitoring

Migration timing of hatchery and natural origin Hanford Reach fall Chinook salmon is estimated from arrival timing at Bonneville Dam based on PIT tag observations at the adult fish ladder for both PRH and Hanford Reach origin fall Chinook salmon.

Life history characteristics of Hanford Reach fall Chinook salmon were assessed by examining carcasses on spawning grounds, fish collected or examined at broodstock collection sites, and by reviewing tagging data and fisheries statistics.
For the 2012-2016 returns, the origin of fall Chinook salmon for the comparison of age and length at maturity is based on a combination of hatchery marks and tags (i.e., otolith, adipose clips, and CWT). PRH origin fall Chinook Salmon were identified by either the presence of an otolith mark specific to PRH or by the presence of a PRH origin CWT. Adipose clipped Chinook salmon without a CWT and without an otolith mark were classified as fish from other hatcheries. The natural origin fish were identified by either a Hanford Reach origin CWT or by the presence of an adipose fin combined with the absence of any hatchery marks. The age composition for both the natural and hatchery origin fall Chinook salmon recovered in return years 2012-2016 were assembled from the carcass recoveries in sections 1-8 of the Hanford Reach.

In order to make coarse comparisons between hatchery and natural origin fish prior to return year 2012, the designation of origin required the assumption that all fish collected in the Hanford Reach, except for those that were of known hatchery origin (e.g., adipose clipped or possessed a CWT), were natural origin. We know this was not the case, but we were not able to identify all of the hatchery origin fish in the demographic samples and it was assumed that the majority of the fish sampled in the stream surveys were natural origin.

### 13.1 Migration Timing

PIT tag observations for both PRH and Hanford Reach natural origin adult fall Chinook salmon at the PIT tag arrays in the Bonneville Dam adult fish ladders were used to assess arrival timing. The PIT tag observation data was obtained from the PTAGIS website. Arrival date for each unique tagged adult was based on its first observation date and time at Bonneville Dam. Annually, the sample sizes have been relatively small due to the low numbers of both hatchery and natural origin fall Chinook salmon PIT tagged. Beginning with the 2011 brood, the number of juveniles PIT tagged at PRH increased from 3,000 to roughly 43,000 annually
The adult PIT tag detections at Bonneville Dam are useful to compare migration timing between Hanford Reach natural origin and PRH origin fall Chinook salmon because harvest and other losses upstream of Bonneville Dam reduce the number of potential detections at upstream sites.
The $10^{\text {th }}, 50^{\text {th }}$, and $90^{\text {th }}$ percentiles of the annual migration timing to Bonneville Dam are given in (Table 37). The observation sample size of both groups of PIT tagged fish at Bonneville Dam can be small and therefore, may not be representative of the populations. However this may be the best migration information currently available.

Table 37 The week that $10 \%, 50 \%$ (median), and $90 \%$ of the natural and hatchery origin fall Chinook salmon passed Bonneville Dam, 2010-2016. Migration timing is based on PIT tag passage of Hanford natural origin and Priest Rapids Hatchery in the adult fish ladder at Bonneville Dam.

| Return Year | Origin | Priest Rapids Origin |  |  |  | Hanford Reach Natural Origin |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age 2 | Age 3 | Age 4 | Age 5 | Age 2 | Age 3 | Age 4 | Age 5 |
| 2010 | $\mathbf{1 0}^{\text {th }}$ Percentile | 28-Aug | 26-Aug |  | 24-Aug | 31-Aug | 5-Sep | 25-Aug |  |
|  | $50^{\text {th }}$ Percentile | 9-Sep | 17-Sep |  | 4-Sep | 21-Sep | 17-Sep | 9-Sep |  |
|  | ${ }^{90}{ }^{\text {th }}$ Percentile | 15-Sep | 24-Sep |  | 6-Sep | 4-Oct | 6-Oct | 15-Sep |  |
|  | N | 5 | 20 | 0 | 3 | 8 | 22 | 18 | 0 |
| 2011 | $10^{\text {th }}$ Percentile | 8-Aug | 3-Sep | 23-Aug |  |  | 4-Sep | 24-Aug | 4-Aug |
|  | $50{ }^{\text {th }}$ Percentile | 8-Sep | 20-Sep | 8-Sep |  |  | 4-Sep | 10-Sep | 30-Aug |
|  | $9^{\text {90 }}$ Percentile | 21-Sep | 25-Sep | 21-Sep |  |  | 10-Sep | 2-Oct | 1-Sep |
|  | N | 6 | 7 | 10 | 0 | 0 | 2 | 65 | 3 |
| 2012 | $10^{\text {th }}$ Percentile | 31-Aug | 6-Sep | 13-Sep | 7-Sep | 14-Sep | 4-Sep | 28-Aug | 27-Aug |
|  | $50{ }^{\text {th }}$ Percentile | 16-Sep | 11-Sep | 13-Sep | 7-Sep | 23-Sep | 16-Sep | 5-Sep | 8-Sep |
|  | $9^{\text {90 }}$ Percentile | 27-Sep | 21-Sep | 19-Sep | 7-Sep | 10-Oct | 26-Sep | 21-Sep | 19-Sep |
|  | N | 7 | 13 | 2 | 1 | 10 | 11 | 19 | 26 |
| 2013 | $10^{\text {th }}$ Percentile | 24-Aug | 28-Aug | 25-Aug |  | 11-Sep | 2-Sep | 2-Sep | 9-Aug |
|  | $50^{\text {th }}$ Percentile | 8-Sep | 9-Sep | 3-Sep |  | 11-Sep | 22-Sep | 9-Sep | 27-Aug |
|  | $9^{\text {90 }}$ Percentile | 18-Sep | 22-Sep | 15-Sep |  | 11-Sep | 10-Oct | 19-Sep | 2-Oct |
|  | N | 40 | 55 | 16 | 0 | 1 | 29 | 22 | 10 |
| 2014 | $10^{\text {th }}$ Percentile | 6-Sep | 4-Sep | 5-Sep |  | 24-Sep | 10-Sep | 3-Sep | 29-Aug |
|  | $50{ }^{\text {th }}$ Percentile | 16-Sep | 13-Sep | 12-Sep |  | 25-Sep | 11-Sep | 12-Sep | 1-Sep |
|  | ${ }^{90}{ }^{\text {th }}$ Percentile | 28-Sep | 25-Sep | 23-Sep |  | 1-Oct | 28-Sep | 26-Sep | 15-Sep |
|  | N | 175 | 228 | 50 | 0 | 3 | 4 | 62 | 5 |
| 2015 | 10 ${ }^{\text {th }}$ Percentile | 16-Oct | 8-Sep | 25-Aug | 14-Sep |  | 10-Sep | 30-Aug | 29-Aug |
|  | $50{ }^{\text {th }}$ Percentile | 16-Oct | 21-Sep | 6-Sep | 26-Sep |  | 20-Sep | 10-Sep | 09-Sep |
|  | $90^{\text {th }}$ Percentile | 16-Oct | 9 -Oct | 18-Sep | 26-Sep |  | 1-Oct | 25-Sep | 25-Sep |
|  | N | 1 | 345 | 323 | 2 | 0 | 5 | 13 | 32 |
| 2016 | 10 ${ }^{\text {th }}$ Percentile |  | 30-Aug | 8-Aug | 14-Aug |  | 21-Sep | 28-Aug | 31-Aug |
|  | $50^{\text {th }}$ Percentile |  | 13-Sep | 7-Sep | 1-Sep |  | 21-Sep | 10-Sep | 7-Sep |
|  | $90^{\text {² }}$ Percentile |  | 6-Oct | 19-Sep | 15-Sep |  | 14-Oct | 19-Sep | 14-Sep |
|  | N | 0 | 41 | 182 | 41 | 0 | 2 | 10 | 5 |

### 13.2 Age at Maturity

Prior to return year 2012, the age composition for hatchery origin returns to PRH was generated by pooling all of the sub-samples from the volunteer trap and ponded fish after expanding for differing demographic sample rates and sub-sample rates. Only one demographic sample rate was used annually in the Hanford Reach stream survey; precluding the need to expand and pool samples. In addition, the fish origin was assigned by location of survey due to the lack of identifiable hatchery marks and low CWT recoveries that were not representative of natural origin fish. Hence, the age composition for natural origin returns was generated from all the samples collected within the carcass survey regardless of true origin. Likewise, the age
composition for hatchery origin fish was generated from all samples collected at PRH regardless of true origin.

The age compositions of the Hanford Reach escapement and the PRH returns are not directly comparable between locations without some adjustment. There is likely a recovery bias against smaller/younger fish in the stream surveys (Zhou 2002; Murdoch et al. 2010; Richards and Pearsons, 2013). Hence, the age composition for the Hanford Reach escapement is likely biased towards larger/older fish. Results and brief discussion for the pilot carcass bias assessments are given in Appendix I. All fish recovered from the PRH volunteer trap are available for systematic sampling; reducing the potential bias of the age composition data. Although this dataset is imperfect, the dataset is maintained for future reference should a method be established to correct the data for associated age and origin bias (Table 38).

The availability of otolith data combined with other hatchery mark data from the Hanford Reach carcass recoveries for the 2012 through 2016 return years provide the ability to estimate age compositions for both hatchery and natural origin fish within the Hanford Reach escapement (Table 40). However, the hatchery origin age composition may be influenced by the low number of hatchery origin fish present in the demographic samples which is further reduced by subsampling the demographic origin. In addition, the age composition for both groups may be biased towards larger fish due to potential size recovery biases in the carcass surveys. Larger demographic samples per return year are required to better represent the age composition data before conclusions can be made. Beginning with return year 2014, the sub-sample size to determine origin was increased substantially to roughly 2,500 fish in order to capture more hatchery origin fish in the sub-sample. Regardless of the methodology, it appears that natural origin fish return at older ages than hatchery origin fish (Table 39 and Table 40). More specifically, the proportion of hatchery origin fish was higher than natural origin fish at ages 2 and 3, and the opposite was true for ages 4, 5, and 6 during brood years 2007-2011.
Table 38 Age compositions for fall Chinook salmon sampled in the Hanford Reach escapement compared to fall Chinook salmon sampled at Priest Rapids Hatchery (genders combined), Brood Years 1998-2011.

| Brood Year | Source ${ }^{1}$ | Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 1998 | Escapement | 0.119 | 0.097 | 0.420 | 0.346 | 0.018 |
|  | PRH Returns | 0.034 | 0.575 | 0.353 | 0.038 | 0.000 |
| 1999 | Escapement | 0.123 | 0.089 | 0.390 | 0.392 | 0.005 |
|  | PRH Returns | 0.061 | 0.366 | 0.432 | 0.140 | 0.001 |
| 2000 | Escapement | 0.262 | 0.081 | 0.290 | 0.359 | 0.009 |
|  | PRH Returns | 0.070 | 0.303 | 0.467 | 0.152 | 0.007 |
| 2001 | Escapement | 0.152 | 0.149 | 0.488 | 0.206 | 0.005 |
|  | PRH Returns | 0.061 | 0.506 | 0.309 | 0.122 | 0.002 |
| 2002 | Escapement | 0.178 | 0.154 | 0.568 | 0.099 | 0.001 |
|  | PRH Returns | 0.103 | 0.386 | 0.466 | 0.043 | 0.001 |
| 2003 | Escapement | 0.249 | 0.170 | 0.248 | 0.331 | 0.000 |
|  | PRH Returns | 0.041 | 0.443 | 0.355 | 0.160 | 0.000 |
| 2004 | Escapement | 0.216 | 0.064 | 0.406 | 0.311 | 0.003 |


| Brood Year | Source ${ }^{1}$ | Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
|  | PRH Returns | 0.133 | 0.398 | 0.406 | 0.063 | 0.000 |
| 2005 | Escapement | 0.151 | 0.082 | 0.306 | 0.458 | 0.003 |
|  | PRH Returns | 0.116 | 0.572 | 0.284 | 0.028 | 0.000 |
| 2006 | Escapement | 0.109 | 0.052 | 0.632 | 0.206 | 0.000 |
|  | PRH Returns | 0.331 | 0.325 | 0.314 | 0.030 | 0.000 |
| 2007 | Escapement | 0.109 | 0.230 | 0.490 | 0.171 | 0.001 |
|  | PRH Returns | 0.103 | 0.483 | 0.381 | 0.033 | 0.000 |
| 2008 | Escapement | 0.159 | 0.193 | 0.511 | 0.137 | 0.000 |
|  | PRH Returns | 0.221 | 0.497 | 0.279 | 0.002 | 0.000 |
| 2009 | Escapement | 0.091 | 0.136 | 0.688 | 0.083 | 0.001 |
|  | PRH Returns | 0.125 | 0.564 | 0.2410 | 0.071 | 0.000 |
| 2010 | Escapement | 0.020 | 0.269 | 0.441 | 0.265 | 0.006 |
|  | PRH Returns | 0.108 | 0.386 | 0.468 | 0.038 | 0.000 |
| $2011^{\text {a }}$ | Escapement | 0.102 | 0.075 | 0.642 | 0.181 | 0.000 |
|  | PRH Returns | 0.065 | 0.430 | 0.449 | 0.056 | 0.000 |
| Mean 1998-2011 | Escapement | 0.146 | 0.132 | 0.466 | 0.253 | 0.004 |
|  | PRH Returns | 0.112 | 0.445 | 0.372 | 0.070 | 0.001 |
| Mean 2007-2011 | Escapement | 0.096 | 0.181 | 0.554 | 0.167 | 0.002 |
|  | PRH Returns | 0.124 | 0.472 | 0.364 | 0.040 | 0.000 |

${ }^{1}$ The origin is assigned by survey
${ }^{\text {a }}$ Does not include age- 6 returns

Table 39 Age compositions for natural and hatchery origin fall Chinook salmon sampled in the Hanford Reach escapement, Brood Years 2007-2011.

| Brood Year | Origin ${ }^{1}$ | $\mathrm{N}^{2}$ | Male Age Composition |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural | 1,093 | No otolith data | 0.377 | 0.483 | 0.139 | 0.002 |
|  | Hatchery | 121 |  | 0.801 | 0.116 | 0.083 | 0.000 |
| 2008 | Natural | 1,234 | 0.044 | 0.336 | 0.502 | 0.118 | 0.000 |
|  | Hatchery | 49 | 0.255 | 0.299 | 0.353 | 0.092 | 0.000 |
| 2009 | Natural | 816 | 0.034 | 0.231 | 0.660 | 0.076 | 0.000 |
|  | Hatchery | 139 | 0.033 | 0.270 | 0.678 | 0.019 | 0.000 |
| 2010 | Natural | 2,093 | 0.008 | 0.361 | 0.454 | 0.176 | 0.000 |
|  | Hatchery | 333 | 0.043 | 0.814 | 0.108 | 0.034 | 0.000 |
| $2011^{\text {a }}$ | Natural | 835 | 0.057 | 0.180 | 0.628 | 0.135 | 0.000 |
|  | Hatchery | 72 | 0.123 | 0.252 | 0.626 | 0.000 | 0.000 |
| Mean | Natural | 1,214 | 0.036 | 0.297 | 0.545 | 0.129 | 0.000 |
|  | Hatchery | 143 | 0.091 | 0.487 | 0.376 | 0.046 | 0.000 |
|  |  |  | Female Age Composition |  |  |  |  |
| Brood Year | Origin ${ }^{1}$ | $\mathrm{N}^{2}$ | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural | 1,299 | No otolith data | 0.047 | 0.706 | 0.247 | 0.000 |
|  | Hatchery | 167 |  | 0.532 | 0.317 | 0.151 | 0.000 |
| 2008 | Natural | 426 | 0.000 | 0.117 | 0.679 | 0.204 | 0.000 |
|  | Hatchery | 74 | 0.000 | 0.176 | 0.651 | 0.172 | 0.000 |
| 2009 | Natural | 486 | 0.000 | 0.033 | 0.789 | 0.175 | 0.003 |
|  | Hatchery | 188 | 0.000 | 0.060 | 0.918 | 0.021 | 0.000 |
| 2010 | Natural | 1,934 | 0.000 | 0.026 | 0.542 | 0.432 | 0.000 |
|  | Hatchery | 353 | 0.000 | 0.418 | 0.448 | 0.133 | 0.000 |
| $2011^{\text {a }}$ | Natural | 921 | 0.000 | 0.005 | 0.769 | 0.226 | 0.000 |
|  | Hatchery | 118 | 0.000 | 0.028 | 0.972 | 0.000 | 0.000 |
| Mean | Natural | 1,013 | 0.000 | 0.046 | 0.697 | 0.257 | 0.001 |
|  | Hatchery | 180 | 0.000 | 0.243 | 0.661 | 0.095 | 0.000 |
|  |  |  | Gender Combined Age Composition |  |  |  |  |
| Brood Year | Origin ${ }^{1}$ | $\mathrm{N}^{2}$ | Age-2 | Age-3 | Age-4 | Age-5 | Age-6 |
| 2007 | Natural | 2,392 | No Otolith Data | 0.201 | 0.602 | 0.196 | 0.001 |
|  | Hatchery | 288 |  | 0.656 | 0.225 | 0.119 | 0.000 |
| 2008 | Natural | 1,660 | 0.022 | 0.230 | 0.587 | 0.160 | 0.002 |
|  | Hatchery | 123 | 0.100 | 0.224 | 0.535 | 0.141 | 0.000 |
| 2009 | Natural | 1,302 | 0.019 | 0.147 | 0.715 | 0.118 | 0.001 |
|  | Hatchery | 327 | 0.012 | 0.136 | 0.831 | 0.021 | 0.000 |
| 2010 | Natural | 4,027 | 0.004 | 0.198 | 0.497 | 0.301 | 0.000 |
|  | Hatchery | 686 | 0.022 | 0.617 | 0.278 | 0.084 | 0.000 |
| $2011^{\text {a }}$ | Natural | 1,756 | 0.025 | 0.084 | 0.706 | 0.185 | 0.000 |
|  | Hatchery | 190 | 0.045 | 0.110 | 0.845 | 0.000 | 0.000 |
| Mean | Natural | 2,227 | 0.018 | 0.172 | 0.621 | 0.192 | 0.001 |
|  | Hatchery | 323 | 0.036 | 0.349 | 0.543 | 0.073 | 0.000 |

[^4]
### 13.3 Size at Maturity

Prior to return year 2012, the size (fork length) at maturity comparisons between fall Chinook salmon recovered at PRH and the Hanford Reach stream survey were calculated in a similar manner as the age composition data for the same time period. Likewise, the assignment of origin was based on the survey (i.e., stream or hatchery). The estimates based on this method may not be representative of natural and hatchery origin fish due to possible size bias during recovery of carcasses.

The availability of otolith marks in addition to other hatchery marks (i.e., otoliths, adipose clips, and CWTs) for the 2012 through 2016 return years provide the ability to estimate size at maturity for both hatchery and natural origin fish within the Hanford Reach escapement. Sub-sample sizes were determined as described in Section 7.

The size at maturity data is essentially complete for brood years 2007 through 2011. It appears that age- 2 and 3 hatchery origin fish tend to be larger than natural origin fish. Likewise, age-4 and 5 natural origin fish tend to be larger than their hatchery origin counterparts (Table 42). This pattern is also seen in Table 41 with the exception that age-3 fish are similar size in the PRH and stream sample.

### 13.4 Gender Composition for Adult Escapement

Prior to return year 2012, the gender ratio comparisons between fall Chinook salmon recovered at PRH and the Hanford Reach stream survey were based on the survey type (i.e., stream or hatchery). Although the estimates based on this method may are imperfect, we continue to present this information to maintain the longest data set available (Table 43).

Gender ratios (male/females) by brood year and origin of adult fall Chinook salmon sampled in the Hanford Reach carcass survey are given in (Table 43). Annually, higher male to female ratios have been observed in the natural origin fish than that of the hatchery origin fish. This may be the result of earlier age of maturity of hatchery origin fish and a size related bias of recovering carcasses in the Hanford Reach.

Table 40 Mean fork length (cm) at age (total age) of fall Chinook salmon sampled in the Hanford Reach escapement compared to fall Chinook salmon sampled at Priest Rapids Hatchery, Brood Years 1999-2011. $\mathrm{N}=$ sample size and SD = 1 standard deviation.

| Brood Year | Origin | Fall Chinook fork length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 1999 | Escapement | 83 | 44 | 4 | 227 | 70 | 6 | 1,423 | 86 | 7 | 1,085 | 93 | 7 | 22 | 103 | 10 |
|  | PRH Returns | 85 | 46 | 5 | 488 | 70 | 5 | 762 | 84 | 6 | 170 | 92 | 6 | 2 | 94 | 11 |
| 2000 | Escapement | 17 | 44 | 4 | 118 | 65 | 7 | 428 | 82 | 6 | 669 | 94 | 8 | 6 | 96 | 9 |
|  | PRH Returns | 25 | 44 | 5 | 136 | 69 | 6 | 196 | 82 | 6 | 58 | 93 | 7 | 2 | 103 | 10 |
| 2001 | Escapement | 32 | 44 | 5 | 251 | 69 | 6 | 1,157 | 84 | 6 | 288 | 93 | 7 | 18 | 97 | 5 |
|  | PRH Returns | 121 | 48 | 4 | 1,040 | 69 | 5 | 628 | 81 | 6 | 183 | 91 | 6 | 9 | 94 | 9 |
| 2002 | Escapement | 31 | 46 | 4 | 229 | 70 | 6 | 194 | 86 | 8 | 239 | 95 | 8 | 2 | 99 | 6 |
|  | PRH Returns | 80 | 52 | 4 | 281 | 70 | 5 | 246 | 84 | 6 | 61 | 91 | 6 | 1 | 73 | 0 |
| 2003 | Escapement | 19 | 48 | 5 | 42 | 69 | 7 | 395 | 85 | 6 | 450 | 96 | 8 |  |  |  |
|  | PRH Returns | 12 | 49 | 6 | 93 | 70 | 6 | 215 | 83 | 6 | 20 | 91 | 4 |  |  |  |
| 2004 | Escapement | 34 | 47 | 4 | 71 | 68 | 6 | 386 | 84 | 6 | 208 | 94 | 8 | 2 | 91 | 1 |


| Brood Year | Origin | Fall Chinook fork length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
|  | PRH Returns | 19 | 55 | 4 | 115 | 69 | 5 | 51 | 84 | 5 | 9 | 95 | 7 |  |  |  |
| 2005 | Escapement | 25 | 50 | 5 | 202 | 70 | 6 | 532 | 84 | 7 | 744 | 96 | 8 | 5 | 96 | 6 |
|  | PRH Returns | 31 | 49 | 4 | 429 | 73 | 4 | 428 | 84 | 6 | 180 | 91 | 6 |  |  |  |
| 2006 | Escapement | 20 | 48 | 4 | 85 | 69 | 6 | 962 | 86 | 6 | 340 | 92 | 7 |  |  |  |
|  | PRH Returns | 3 | 45 | 3 | 42 | 71 | 4 | 170 | 84 | 6 | 13 | 92 | 7 |  |  |  |
| 2007 | Escapement | 24 | 46 | 5 | 642 | 72 | 6 | 1,468 | 84 | 7 | 482 | 92 | 7 | 1 | 105 | 0 |
|  | PRH Returns | 5 | 50 | 4 | 1,149 | 71 | 4 | 1,419 | 80 | 5 | 179 | 87 | 6 |  |  |  |
| 2008 | Escapement | 34 | 50 | 4 | 243 | 70 | 5 | 620 | 84 | 7 | 72 | 92 | 8 | 1 | 84 | 0 |
|  | PRH Returns | 22 | 52 | 5 | 652 | 69 | 4 | 573 | 81 | 6 | 1 | 84 | 0 |  |  |  |
| 2009 | Escapement | 50 | 48 | 4 | 421 | 69 | 6 | 931 | 81 | 6 | 183 | 92 | 10 |  |  |  |
|  | PRH Returns | 308 | 48 | 4 | 1,690 | 68 | 5 | 218 | 77 | 5 | 70 | 86 | 7 |  |  |  |
| 2010 | Escapement | 63 | 47 | 7 | 1,040 | 68 | 5 | 2,754 | 82 | 7 | 826 | 88 | 7 | 29 | 90 | 6 |
|  | PRH Returns | 883 | 48 | 4 | 1,375 | 69 | 4 | 1,528 | 78 | 5 | 55 | 84 | 4 |  |  |  |
| $2011^{\text {a }}$ | Escapement | 58 | 46 | 4 | 226 | 67 | 5 | 1,151 | 80 | 6 | 465 | 88 | 7 |  |  |  |
|  | PRH Returns | 111 | 47 | 3 | 694 | 67 | 4 | 355 | 77 | 5 | 201 | 84 | 6 | 1 | 87 | 0 |
| $\begin{gathered} \text { Mean } 99 \\ -11 \\ \hline \end{gathered}$ | Escapement | 38 | 47 | 5 | 292 | 69 | 6 | 954 | 84 | 7 | 465 | 93 | 8 | 7 | 66 | 3 |
|  | PRH Returns | 131 | 49 | 4 | 630 | 70 | 5 | 522 | 81 | 6 | 92 | 89 | 6 | 1 | 35 | 2 |
| $\begin{aligned} & \text { Mean } \\ & \text { 07-11 } \end{aligned}$ | Escapement | 46 | 47 | 5 | 514 | 69 | 5 | 1,385 | 82 | 7 | 406 | 90 | 8 | 6 | 56 | 1 |
|  | PRH Returns | 266 | 49 | 4 | 1,112 | 69 | 4 | 819 | 79 | 5 | 101 | 85 | 5 | 0 | 17 | 0 |

${ }^{a}$ Does not include age-6 returns

Table 41 Mean fork length (cm) at age (total age) of natural and hatchery origin fall Chinook salmon that spawned naturally in the Hanford Reach, Brood Years 2007-2011. $\mathrm{N}=$ sample size and $\mathrm{SD}=1$ standard deviation.

| Brood Year | Origin | Male Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2007 | Natural | No otolith Data |  |  | 362 | 70 | 5 | 206 | 84 | 8 | 154 | 98 | 8 | 1 | 105 | 0 |
|  | Hatchery |  |  |  | 44 | 72 | 4 | 16 | 82 | 5 | 6 | 93 | 7 | 0 |  |  |
| 2008 | Natural | 22 | 49 | 4 | 134 | 69 | 5 | 260 | 85 | 8 | 25 | 99 | 7 | 0 |  |  |
|  | Hatchery | 8 | 52 | 3 | 20 | 69 | 5 | 7 | 86 | 4 | 2 | 91 | 15 | 0 |  |  |
| 2009 | Natural | 3 | 48 | 3 | 325 | 68 | 6 | 123 | 82 | 6 | 40 | 99 | 7 | 0 |  |  |
|  | Hatchery | 2 | 55 | 5 | 34 | 71 | 6 | 21 | 79 | 10 | 2 | 96 | 6 | 0 |  |  |
| 2010 | Natural | 33 | 45 | 4 | 325 | 68 | 6 | 855 | 83 | 8 | 238 | 94 | 8 | 4 | 97 | 8 |
|  | Hatchery | 25 | 50 | 4 | 34 | 71 | 6 | 34 | 79 | 7 | 7 | 92 | 7 | 0 |  |  |
| 2011 ${ }^{\text {a }}$ | Natural | 33 | 45 | 4 | 175 | 66 | 5 | 413 | 67 | 8 | 137 | 94 | 7 |  |  |  |
|  | Hatchery | 25 | 50 | 4 | 19 | 68 | 4 | 31 | 69 | 6 | 7 | 88 | 7 |  |  |  |
| Mean | Natural | 23 | 47 | 4 | 264 | 68 | 5 | 371 | 80 | 8 | 119 | 97 | 7 | 3 | 101 | 4 |
|  | Hatchery | 15 | 52 | 4 | 30 | 70 | 5 | 22 | 79 | 6 | 5 | 92 | 8 | 0 | 0 | 0 |
|  |  | Female Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brood |  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |
| Year | Origin | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2007 | Natural | 0 |  |  | 83 | 72 | 5 | 376 | 83 | 5 | 326 | 89 | 5 | 0 |  |  |
|  | Hatchery | 0 |  |  | 48 | 72 | 4 | 48 | 80 | 4 | 8 | 86 | 6 | 0 |  |  |
| 2008 | Natural | 0 |  |  | 36 | 70 | 3 | 344 | 83 | 5 | 49 | 88 | 5 | 1 | 84 | 0 |
|  | Hatchery | 0 |  |  | 23 | 70 | 5 | 21 | 82 | 4 | 7 | 85 | 6 | 0 |  |  |
| 2009 | Natural | 0 |  |  | 44 | 71 | 5 | 105 | 80 | 4 | 82 | 87 | 11 | 1 | 73 | 0 |
|  | Hatchery | 0 |  |  | 12 | 68 | 4 | 49 | 78 | 6 | 4 | 85 | 4 | 0 |  |  |
| 2010 | Natural | 0 |  |  | 44 | 71 | 5 | 82 | 87 | 5 | 550 | 85 | 4 | 20 | 89 | 5 |
|  | Hatchery | 0 |  |  | 10 | 69 | 4 | 4 | 87 | 5 | 29 | 82 | 4 | 0 |  |  |
| 2011 ${ }^{\text {a }}$ | Natural | 0 |  |  | 7 | 67 | 5 | 626 | 80 | 5 | 282 | 85 | 5 |  |  |  |
|  | Hatchery | 0 |  |  | 4 | 65 | 2 | 76 | 77 | 4 | 35 | 84 | 4 |  |  |  |
| Mean | Natural | 0 |  |  | 43 | 70 | 5 | 307 | 83 | 5 | 258 | 87 | 6 | 7 | 82 | 2 |
|  | Hatchery | 0 |  |  | 19 | 69 | 4 | 40 | 81 | 5 | 17 | 84 | 5 | 0 | 0 | 0 |
|  |  | Gender Combined Fork Length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Brood |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
| Year | Origin | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2007 | Natural | No otolith Data |  |  | 445 | 70 | 5 | 582 | 83 | 6 | 480 | 92 | 6 | 1 | 105 | 0 |
|  | Hatchery |  |  |  | 92 | 72 | 4 | 64 | 81 | 4 | 14 | 89 | 6 | 0 |  |  |
| 2008 | Natural | 22 | 49 | 4 | 170 | 69 | 5 | 604 | 84 | 6 | 74 | 92 | 6 | 1 | 84 | 0 |
|  | Hatchery | 8 | 52 | 3 | 43 | 70 | 5 | 28 | 83 | 4 | 9 | 86 | 8 | 0 |  |  |
| 2009 | Natural | 3 | 48 | 3 | 369 | 68 | 6 | 228 | 81 | 5 | 122 | 91 | 10 | 1 | 73 | 0 |
|  | Hatchery | 2 | 55 | 5 | 46 | 70 | 5 | 70 | 78 | 7 | 6 | 89 | 5 | 0 |  |  |
| 2010 | Natural | 33 | 45 | 4 | 369 | 68 | 6 | 937 | 83 | 8 | 788 | 88 | 5 | 24 | 90 | 7 |
|  | Hatchery | 25 | 50 | 4 | 44 | 71 | 6 | 38 | 80 | 7 | 36 | 84 | 5 | 0 |  |  |
| 2011 ${ }^{\text {a }}$ | Natural | 33 | 45 | 4 | 182 | 66 | 5 | 1039 | 80 | 6 | 419 | 88 | 7 |  |  |  |
|  | Hatchery | 25 | 50 | 4 | 23 | 67 | 4 | 107 | 78 | 5 | 42 | 84 | 5 |  |  |  |
| Mean | Natural | 23 | 47 | 4 | 307 | 68 | 5 | 678 | 82 | 6 | 377 | 90 | 7 | 7 | 88 | 2 |
|  | Hatchery | 15 | 52 | 4 | 50 | 70 | 5 | 61 | 80 | 5 | 21 | 86 | 6 | 0 | 0 | 0 |

${ }^{\text {a }}$ Brood year does not include age-6 returns

Table 42 Comparison male to female ratio of fall Chinook salmon sampled at Priest Rapids Hatchery and in the Hanford Reach stream surveys, Brood Years 2007-2011.

| Brood Year | Origin | Male ${ }^{1}$ : Female Ratio |
| :---: | :---: | :---: |
| 1996 | Stream | 0.94:1 |
|  | Hatchery | 1.98:1 |
| 1997 | Stream | 0.48:1 |
|  | Hatchery | 1.88:1 |
| 1998 | Stream | 0.66:1 |
|  | Hatchery | 1.38:1 |
| 1999 | Stream | 0.71:1 |
|  | Hatchery | 2.15:1 |
| 2000 | Stream | 1.51:1 |
|  | Hatchery | 2.40:1 |
| 2001 | Stream | 0.67:1 |
|  | Hatchery | 2.31:1 |
| 2002 | Stream | 1.40:1 |
|  | Hatchery | 1.94:1 |
| 2003 | Stream | 1.25:1 |
|  | Hatchery | 1.64:1 |
| 2004 | Stream | 1.17:1 |
|  | Hatchery | 1.63:1 |
| 2005 | Stream | 0.87:1 |
|  | Hatchery | 2.15:1 |
| 2006 | Stream | 0.75:1 |
|  | Hatchery | 2.57:1 |
| 2007 | Stream | 0.78:1 |
|  | Hatchery | 1.60:1 |
| 2008 | Stream | 0.82:1 |
|  | Hatchery | 1.89:1 |
| 2009 | Stream | 1.07:1 |
|  | Hatchery | 2.57:1 |
| 2010 | Stream | 0.70:1 |
|  | Hatchery | 1.47:1 |
| $2011^{\text {a }}$ | Stream | 0.71:1 |
|  | Hatchery | 2.00:1 |
| Mean | Stream | 0.90:1 |
|  | Hatchery | 1.97:1 |

${ }^{1}$ Includes both adult males and jacks.
${ }^{\text {a }}$ Includes age- 2 through 5 .

Table 43 Comparison male to female ratio of fall Chinook salmon sampled in the Hanford Reach stream surveys, Brood Years 2007-2011.

| Brood Year | Origin | Male ${ }^{1}$ : Female Ratio |
| :---: | :---: | :---: |
| $2007{ }^{\text {a }}$ | Natural | 0.86:1.00 |
|  | Hatchery | 0.74:1.00 |
| 2008 | Natural | 1.07:1.00 |
|  | Hatchery | 0.64:1.00 |
| 2009 | Natural | 1.37:1.00 |
|  | Hatchery | 0.56:1.00 |
| 2010 | Natural | 1.02:1.00 |
|  | Hatchery | 1.01:1.00 |
| $2011{ }^{\text {b }}$ | Natural | 0.94:1.00 |
|  | Hatchery | 0.51:1:00 |
| Mean | Natural | 1.05:1.00 |
|  | Hatchery | 0.69:1:00 |

${ }^{1}$ Includes both adult males and jacks. ${ }^{\text {a }}$ Does not include age-2. ${ }^{\mathrm{b}}$ Includes age-2 through 5.

### 13.5 Egg Retention

All female Chinook included in the demographic sample for the Hanford Reach stream surveys were examined for egg retention to assess spawn success. The females sampled were partitioned into the egg retention categories of $0 \%, 25 \%, 50 \%, 75 \%$ and $100 \%$. The assignment of origin for each female for years 2010 and 2011 were based on the presence or absence of an adipose fin. The adipose intact group may include non-adipose clipped fish from PRH. A combination of hatchery marks (i.e., adipose clips, CWTs, and otolith marks were used to identify hatchery origin fish in years 2013-2016. For all years, we assume that fish not possessing any hatchery marks are natural origin fish.

The assessment of egg retention is compromised by the loss of eggs during the collection and transport of carcasses prior to sampling. In addition, the methods for quantifying egg retention and assignment of origin for each female have varied between years. The amount of egg retention for years 2010 through 2013 were determined by visual estimates; whereas, during 2014 through 2016, the amount of retention was based on egg counts when the gametes were not completely intact. For these recent data sets, the percent of egg retention was calculated by dividing the amount of egg retained by an estimated fecundity based on length versus fecundity regressions by origin (Hatchery or Natural). An explanation of these regressions is provided in the fecundity section of this report.
During the 2015 and 2016 surveys, staff recorded visual observations of egg retention based on the standard egg retention categories to make comparisons with egg retention based on egg counts. The data from the egg counts were categorized into the standard egg retention categories based on the following ranges: $1=100-88 \%, 2=87-63 \%, 3=62-38 \%, 4=37-11 \%$, and $5=10-$ $0 \%$. This comparison may allow us to assess the egg retention estimates based on methods used prior to 2015. The difference between two methods was less than 1 percentage point by category for each year (Table 44), which provides some confidence that the visual methods of the past may provide reasonable indices of spawning success.

Table 44 Comparison of egg retention of natural and hatchery origin fall Chinook sampled in the Hanford Reach stream survey, Return Years 2015-2016.

| Survey Year | Egg Retention Categories | \% by Category based on Egg Counts | \% by Category based on Visual Observations | Difference between Actual and Observed (\%) |
| :---: | :---: | :---: | :---: | :---: |
| 2015 | 1-100\% | 0.0 | 0.4 | -0.4 |
|  | 2-75\% | 0.3 | 0.5 | -0.2 |
|  | 3-50\% | 0.6 | 0.6 | 0.0 |
|  | 4-25\% | 1.5 | 1.6 | -0.1 |
|  | 5-0\% | 97.7 | 96.9 | 0.8 |
|  | $\mathrm{N}=1,405$ |  |  |  |
| 2016 | 1-100\% | 0.0 | 0.2 | -0.2 |
|  | $2-75 \%$ | 0.2 | 0.3 | -0.1 |
|  | 3-50\% | 0.6 | 1.1 | -0.5 |
|  | 4-25\% | 1.8 | 1.2 | 0.6 |
|  | 5-0\% | 97.4 | 97.2 | 0.2 |
|  | $\mathrm{N}=995$ |  |  |  |

The adjusted spawn successes for the escapement between years 2010 and 2016 were generally greater than $97 \%$ (Table 45). The spawn success was lower for both hatchery and natural origin females in return year 2013 compared to other years. These observations coincide with an elevated pHOS which most likely resulted from high hatchery fish escapement and restrictive operations of the PRH volunteer trap during 2013. It is possible that a portion of these PRH origin females which were unable to enter the trap, died without spawning, and ended up being surveyed in the Hanford Reach carcass survey. The spawn success was high during 2014 and 2015 despite historically high record escapements to the Hanford Reach.

Table 45 Comparison of egg retention of natural and hatchery origin fall Chinook sampled in the Hanford Reach stream survey, Return Years 2010-2016.

| Return Year | Origin | Females Sampled | Egg Retention Categories |  |  |  |  | No Egg <br> Retention (\%) | Adj Spawn Success for Escapement (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 0 \% | $\mathbf{2 5 \%}$ | $\mathbf{5 0 \%}$ | $\mathbf{7 5 \%}$ | 100\% |  |  |
| 2004 | Combined | 1,176 | 1,151 | NA | 21 | NA | 4 | 97.9 | 98.8 |
| 2005 | Combined | 1,323 | 1,310 | NA | 6 | NA | 7 | 99.0 | 99.2 |
| 2006 | Combined | 352 | 343 | NA | 8 | NA | 1 | 97.4 | 98.6 |
| 2007 | Combined | 454 | 443 | NA | 8 | NA | 3 | 97.6 | 98.5 |
| 2008 | Combined | No spawn success data collected |  |  |  |  |  |  |  |
| 2009 | Combined | 499 | 484 | NA | 5 | NA | 10 | 97.0 | 97.5 |
| 2010 | Combined | 1,173 | 1,147 | 6 | 13 | 1 | 6 | 97.8 | 98.7 |
| 2011 | Combined | 1,264 | 1,203 | 1 | 52 | 5 | 3 | 95.2 | 97.4 |
| $2012{ }^{\text {b }}$ | Natural | 681 | 658 | 14 | 5 | 1 | 3 | 96.6 | 98.6 |
|  | Hatchery | 90 | 89 | 0 | 0 | 0 | 1 | 98.9 | 98.9 |
|  | Total | 771 | 747 | 14 | 5 | 1 | 4 | 96.9 | 98.6 |
| $2013{ }^{\text {b }}$ | Natural | 461 | 392 | 51 | 9 | 3 | 6 | 85.0 | 94.5 |
|  | Hatchery | 224 | 144 | 39 | 11 | 13 | 17 | 64.3 | 81.3 |
|  | Total | 685 | 536 | 90 | 20 | 16 | 23 | 78.2 | 90.1 |
| $2014{ }^{\text {b }}$ | Natural | 1,082 | 1,074 | 1 | 0 | 0 | 7 | 99.3 | 99.3 |
|  | Hatchery | 153 | 141 | 3 | 0 | 0 | 9 | 92.2 | 93.6 |
|  | Total | 1,235 | 1,215 | 4 | 0 | 0 | 16 | 98.4 | 98.6 |
| $2015{ }^{\text {b }}$ | Natural | 1256 | 1237 | 14 | 3 | 2 | 0 | 98.5 | 99.5 |
|  | Hatchery | 149 | 135 | 7 | 5 | 2 | 0 | 90.6 | 96.1 |
|  | Total | 1,405 | 1,372 | 21 | 8 | 4 | 0 | 97.7 | 99.1 |
| $2016{ }^{\text {b }}$ | Natural | 857 | 842 | 7 | 3 | 1 | 0 | 98.2 | 99.5 |
|  | Hatchery | 138 | 127 | 11 | 3 | 1 | 0 | 92 | 96.4 |
|  | Total | 995 | 969 | 18 | 6 | 2 | 0 | 97.4 | 99.1 |
| Mean Natural Spawn Success (RY 2012-2016) |  |  |  |  |  |  |  | 95.5 | 98.3 |
| Mean Hatchery Spawn Success (RY 2012 - 2016) |  |  |  |  |  |  |  | 87.6 | 93.3 |
| Mean Combined Spawn Success (RY 2010-2016) |  |  |  |  |  |  |  | 94.5 | 97.4 |

The measure for reporting egg retention changed from that used for previous years beginning in 2010
${ }^{\mathrm{b}}$ Origins were determined the presence or absence of otolith marks, adipose clips and CWTs

### 14.0 Contribution to Fisheries

The contribution of fish produced at PRH to fisheries was estimated by querying the Regional Mark Processing Center (RMPC) database. This is central repository for all CWT and otherwise associated release, catch, sample, and recovery data of anadromous salmonids in the greater Pacific Coast Region of the United States of America. The Regional Mark Information System database (RMIS) within the RMPC provides specific recovery data for individual tag codes, along with the sample rate used to derive the estimated total number of recoveries by fishery type.

The CWT data reported to RMPC are expanded by sample rates generated by the agency reporting the data. In some cases, the estimated number of tags reported is less than the number actually observed. This typically occurs when the sample rate is unknown, not reported, or biased (Gilbert Lensegrav, WDFW, personal communication). In these instances, the observed number was used instead of the estimated number to calculate the numbers of PRH origin fish recovered by location.

The RMIS database was queried for tag recoveries on January 16, 2017 to provide CWT recoveries of PRH origin fish. The database for the 2009 brood should be complete for age- 2 through age-5. The age-6 recovered during RY2016 may not be included until January 1, 2018 due to the lag in reporting field data to RMPC.

Beginning with the 2010 release year, portions of the non-adipose clipped smolts released from PRH received a CWT as part of a double index tag (DIT) study to evaluate the effect of various mark-selective fisheries occurring in Oregon, Washington, and British Columbia waters (PSC 2013). We are currently reviewing the data reported to the RMPC database to evaluate the results of the double index tagging for the PRH origin fish. Data for brood years 2009 and 2010 show that adipose clipped fish from the DIT groups are being recovered in mark selective fisheries occurring in ocean, marine, and freshwater zones. Comparisons of the demographics between the DIT groups recovered at PRH are very similar (Appendix J). Therefore, mark selective fisheries do not appear to markedly influence the demographic data collected at PRH.
Fall Chinook salmon released from PRH supplement Pacific Ocean harvest for both commercial and sport fisheries from Washington to Southeast Alaska as well as Columbia River commercial, sport, and treaty tribal harvest. The Hanford Reach sport fishery for fall Chinook salmon is an extremely popular fishery. This fishery typically runs annually from August 1 to late October. In 2016, an estimated 17,927 fall Chinook salmon were harvested during this fishery; 16,859 adults and 1,068 jacks. Estimates generated from CWT recoveries from the Hanford Reach sport fishery suggest that $12.1 \%$ ( 2,177 fish) of the total sport harvest in the Hanford Reach was comprised of fall Chinook salmon released from PRH (Table 46). Likewise, fall Chinook salmon released from Ringold Springs Hatchery comprised 2.8\% (992 fish) of the sport fishery. Strays from other hatcheries combined represent $0.9 \%$ ( 319 fish) of the harvest. Sport harvest monitoring in the Hanford Reach and lower Yakima includes surveying both adipose intact and adipose clipped fish for CWTs. Recent data from otolith sampling indicates that CWT expansions may underestimate the number of PRH origin fall Chinook salmon annually returning to PRH. A similar situation may occur when evaluating hatchery contributions to the sport fishery.

The CWT data for PRH origin fall Chinook salmon that were marked with an adipose clip were reviewed to assess contributions to marine and freshwater, commercial, tribal, and sport fisheries. The largest proportion of the harvest of PRH origin fall Chinook salmon occurred in ocean fisheries followed by Zone-6 tribal harvest. For brood years 1997 through 2010, 49\% of the reported harvest was taken in ocean fisheries and the other $51 \%$ in the Columbia River fisheries (Table 47). The adipose clip CWT rate for the 2009 and 2010 broods notably increased from previous brood years. Not all CWT surveys locations check harvested adipose intact fish for the presence of a CWT. Therefore, the data presented in Table 47 includes harvest estimates based on recoveries of adipose clipped CWT tagged fish.

Table 46 Hatchery fall Chinook salmon contributions to harvest in the Hanford Reach fall Chinook salmon fishery. Coded-wire tag recoveries provided from RMIS database were expanded by sample rate and juvenile tag rate, Return Years 2003-2016.

| Return <br> Year | Harvest \& Sampling |  |  | CWT Expansions |  |  | \% of Harvest |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Harvest | Sampled | \% | PRH | RSH | Other Hatcheries | PRH | RSH | Other Hatcheries |
| 2003 | 7,190 | 1,848 | 25.7 | 510 | 424 | 43 | 7.1 | 5.9 | 0.6 |
| 2004 | 8,787 | 2,255 | 25.7 | 276 | 62 | 23 | 3.1 | 0.7 | 0.3 |
| 2005 | 7,974 | 1,834 | 23.0 | 1,200 | 265 | 35 | 15.0 | 3.3 | 0.4 |
| 2006 | 4,508 | 1,296 | 28.7 | 683 | 66 | 10 | 15.1 | 1.5 | 0.2 |
| 2007 | 6,466 | 1,812 | 28.0 | 929 | 50 | 89 | 14.4 | 0.8 | 1.4 |
| 2008 | 7,013 | 1,593 | 22.7 | 304 | 66 | 22 | 4.3 | 0.9 | 0.3 |
| 2009 | 8,806 | 1,741 | 19.8 | 520 | 0 | 10 | 5.9 | 0.0 | 0.1 |
| 2010 | 12,499 | 2,475 | 19.8 | 1,157 | 399 | 10 | 9.3 | 3.2 | 0.1 |
| 2011 | 14,262 | 2,715 | 19.0 | 1,558 | 663 | 121 | 10.9 | 4.6 | 0.8 |
| 2012 | 18,854 | 3,615 | 19.2 | 3,974 | 1,974 | 237 | 21.1 | 10.5 | 1.3 |
| 2013 | 27,630 | 5,555 | 20.2 | 6,570 | 3,947 | 537 | 23.8 | 14.3 | 1.9 |
| 2014 | 32,417 | 8,319 | 25.7 | 3,987 | 1,419 | 332 | 12.3 | 4.4 | 1.0 |
| 2015 | 35,419 | 10,327 | 29.2 | 4,144 | 992 | 319 | 11.7 | 2.8 | 0.9 |
| 2016 | 17,927 | 5,544 | 30.9 | 2,177 | 822 | 339 | 12.1 | 4.6 | 1.9 |
| Mean | 14,982 | 3,638 | 24.1 | 1,999 | 796 | 152 | 11.9 | 4.1 | 0.8 |

Table 47 Priest Rapids Hatchery coded-wire tag recoveries provided from RMIS by brood year and harvest type expanded by sample rate and juvenile tag rate, Brood Years 1997-2010. Data only includes coded-wire tag recoveries from adipose clipped fish expanded by the juvenile tag rate.

| Brood Year | Ocean Fisheries |  | Columbia River Fisheries |  |  |  |  |  | Total Harvest | P^ CWT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Tribal |  | Commercial |  | Recreational |  |  |  |
|  | \# | \% | \# | \% | \# | \% | \# | \% |  |  |
| 1997 | 1,100 | 36.7 | 1,506 | 50.2 | 304 | 10.1 | 91 | 3.0 | 3,001 | 0.03 |
| 1998 | 6,580 | 48.4 | 3,956 | 29.1 | 1,066 | 7.8 | 1,981 | 14.6 | 13,583 | 0.03 |
| 1999 | 14,190 | 54.6 | 5,908 | 22.8 | 2,410 | 9.3 | 3,458 | 13.3 | 25,966 | 0.029 |
| 2000 | 4,938 | 61.5 | 1,583 | 19.7 | 1,099 | 13.7 | 412 | 5.1 | 8,032 | 0.032 |
| 2001 | 17,758 | 56.5 | 6,612 | 21.1 | 1,554 | 4.9 | 5,484 | 17.5 | 31,408 | 0.052 |
| 2002 | 3,779 | 50.6 | 1,240 | 16.6 | 576 | 7.7 | 1,869 | 25.0 | 7,464 | 0.052 |
| 2003 | 1,871 | 54.6 | 570 | 16.6 | 226 | 6.6 | 757 | 22.1 | 3,424 | 0.059 |
| 2004 | 562 | 49.3 | 364 | 31.9 | 214 | 18.8 | 0 | 0.0 | 1,140 | 0.059 |
| 2005 | 10,699 | 52.1 | 5,975 | 29.1 | 998 | 4.9 | 2,871 | 14.0 | 20,543 | 0.03 |
| 2006 | 1,023 | 44.1 | 713 | 30.7 | 288 | 12.4 | 298 | 12.8 | 2,322 | 0.029 |
| 2007 | 13,838 | 44.4 | 10,620 | 34.1 | 2,160 | 6.9 | 4,523 | 14.5 | 31,141 | 0.03 |
| 2008 | 5,763 | 43.7 | 4,447 | 33.7 | 887 | 6.7 | 2,080 | 15.8 | 13,177 | 0.032 |
| 2009 | 24,872 | 43.4 | 21,121 | 36.8 | 2,581 | 4.5 | 8,761 | 15.3 | 57,335 | 0.091 |
| 2010 | 46,584 | 43.5 | 34275 | 32.0 | 7886 | 7.4 | 18299 | 17.1 | 107,044 | 0.089 |
| Mean | 10,968 | 48.8 | 7,064 | 28.9 | 1,589 | 8.7 | 3,635 | 13.6 | 23,256 | 0.046 |

### 15.0 Straying

The distribution of PRH origin fish spawning in areas outside of the target stream is presented to assess the level of straying and potential impacts on other populations. The presumptive target spawning location for PRH origin fish includes the section of Columbia River from McNary Dam to Wanapum Dam as well as the lower Yakima River below Prosser Dam.

The spawning escapement of PRH origin fish by brood year is determined from CWT recoveries collected during spawning surveys. The CWT recoveries are expanded by the juvenile mark rates and survey sampling rates to estimate the number of PRH origin fish recovered on spawning grounds.
The stray rates (i.e., fish that spawned outside of the presumptive target area / total escapement) for each brood year were calculated from the estimated recoveries of PRH origin fish from spawning grounds within and outside of the presumptive target area. CWT recoveries at nontarget hatcheries and adult fish traps are not included. These fish were not considered strays because these fish were not able to leave the facilities on their own volition.

There are three target rates for straying given in the Monitoring and Evaluation Plan for PUD Hatchery Programs (Hillman et al. 2013):
1). Stray rate for PRH origin fall Chinook salmon should be less than $5 \%$ of total brood return.
2). Stray rate for PRH origin fall Chinook salmon should be less than $5 \%$ of the spawning escapement for other non-target independent populations based on run year.
3). Stray rate for PRH origin fall Chinook salmon should be less than $10 \%$ of the spawning escapement of any non-target streams within the independent population based on run year.

With one exception, less than $5 \%$ of the PRH origin returns for each brood year were estimated to have spawned outside of the presumptive target spawning area (Table 48). The 2006 brood is the only cohort found at rates greater than 5\% outside of the presumptive target area. For this cohort, $37 \%$ of the estimated strays occurred in the Chelan River. This estimate is based on the expansion of one PRH origin CWT fish recovered in the Chelan River escapement. The Chelan River spawning population is a mix of both summer and fall Chinook salmon strays and is not considered an independent population. This location was included to show contributions of PRH strays to this group of fish.

Examination of CWT recoveries by return year for presumptive non-target streams or areas suggest that PRH fall Chinook salmon seldom exceeded more than $5 \%$ of the spawning escapement for other independent populations of fall Chinook salmon. However, for multiple return years, greater than $5 \%$ of the spawning escapement for the Chelan River consisted of PRH origin fall Chinook salmon (

Table 49).

Table 48 Estimated number and proportions of Priest Rapids Hatchery fall Chinook salmon spawning escapement to Priest Rapids Hatchery and streams within and outside of the presumptive target stream by brood year (1992-2010). Coded-wire tag recoveries are expanded by juvenile mark rate and survey sample rate for each brood year.

| Brood Year | Number of PRH Origin Recoveries | Homing |  |  |  | Straying Outside of Target Stream |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Target Hatchery |  | Target Stream ${ }^{1}$ |  |  |  |
|  |  | Number | Proportion | Number | Proportion | Number | Proportion |
| 1992 | 9,037 | 7,630 | 0.844 | 1,037 | 0.115 | 370 | 0.041 |
| 1993 | 25,966 | 21,144 | 0.814 | 4,821 | 0.186 | 0 | 0.000 |
| 1994 | 1,692 | 1,385 | 0.818 | 308 | 0.182 | 0 | 0.000 |
| 1995 | 30,655 | 23,414 | 0.764 | 7,207 | 0.235 | 34 | 0.001 |
| 1996 | 13,552 | 10,034 | 0.740 | 3,517 | 0.260 | 0 | 0.000 |
| 1997 | 3,172 | 2,690 | 0.848 | 483 | 0.152 | 0 | 0.000 |
| 1998 | 18,167 | 11,833 | 0.651 | 5,867 | 0.323 | 467 | 0.026 |
| 1999 | 27,333 | 15,467 | 0.566 | 11,867 | 0.434 | 0 | 0.000 |
| 2000 | 4,759 | 3,690 | 0.775 | 1,069 | 0.225 | 0 | 0.000 |
| 2001 | 25,375 | 15,875 | 0.626 | 9,469 | 0.373 | 31 | 0.001 |
| 2002 | 5,288 | 3,769 | 0.713 | 1,519 | 0.287 | 0 | 0.000 |
| 2003 | 3,034 | 2,034 | 0.670 | 949 | 0.313 | 51 | 0.017 |
| 2004 | 1,133 | 1,133 | 1.000 | 0 | 0.000 | 0 | 0.000 |
| 2005 | 21,379 | 17,103 | 0.800 | 4,241 | 0.198 | 34 | 0.002 |
| 2006 | 1,001 | 634 | 0.633 | 0 | 0.000 | 367 | 0.367 |
| 2007 | 22,206 | 19,220 | 0.866 | 2,964 | 0.133 | 22 | 0.001 |
| 2008 | 11,867 | 9,002 | 0.759 | 2,864 | 0.241 | 0 | 0.000 |
| 2009 | 27,968 | 17,760 | 0.635 | 10,132 | 0.362 | 36 | 0.001 |
| 2010 | 99,571 | 68,231 | 0.685 | 40,574 | 0.407 | 63 | 0.001 |
| Mean | 18,585 | 13,266 | 0.748 | 5,731 | 0.233 | 78 | 0.024 |

${ }^{1}$ Target stream includes the Columbia River between McNary and Wanapum dams as well as the Yakima River below Prosser Dam.

Table 49 Proportion of fall/summer Chinook spawning populations by return year comprised of Priest Rapids Hatchery fall Chinook from 1998-2010 brood releases based on coded-wire tag recoveries.

| Return Year | Presumptive Non-Target Stream |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Snake Fall Chinook |  | Yakima Fall Chinook |  | Wenatchee Summer Chinook |  | Entiat River ${ }^{1}$ |  | Chelan River ${ }^{1}$ |  | Methow Summer Chinook |  | Okanogan Summer Chinook |  |
|  | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| 2000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2001 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 334 | 0.343 | 0 | 0.000 | 0 | 0.000 |
| 2002 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 274 | 0.468 | 0 | 0.000 | 0 | 0.000 |
| 2003 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2004 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2005 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2006 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2007 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2008 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 57 | 0.031 |
| 2009 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 117 | 0.091 | 0 | 0.000 | 0 | 0.000 |
| 2010 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 359 | 0.328 | 0 | 0.000 | 0 | 0.000 |
| 2011 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2012 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 19 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2013 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 37 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2014 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 30 | 0.000 | 0 | 0.000 | 0 | 0.000 |
| 2015 | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 22 | 0.019 | 135 | 0.093 | 0 | 0.000 | 0 | 0.000 |
| Mean | 0 | 0.000 | 0 | 0.000 | 0 | 0.000 | 1 | 0.001 | 82 | 0.083 | 0 | 0.000 | 4 | 0.002 |

${ }^{1}$ The Chelan and Entiat River spawning populations are a mix of both summer and fall Chinook salmon strays and are not considered independent populations. These locations were included to show contributions of PRH strays to these groups of fish.
As previously described in Section 4, approximately 3,000 smolts at PRH have been annually PIT tagged at PRH from brood years 1995 through 2010. The annual release of PIT tagged smolts was increased to $\sim 43,000$ beginning with brood year 2011. The last known observations of individual PIT tag adult fall Chinook salmon originating from PRH at detection locations above McNary Dam are given in Table 50 for brood years 1999 through 2013. The number of observed PRH PIT tagged adults should dramatically increase in the forthcoming years.
The majority of the PIT tagged PRH adults observed at McNary Dam have been observed at PRD and/or PRH. Very few fish have been detected in the Snake River, which is possibly the area of most concern for straying. In addition, notable proportions of the returns for several brood years have been observed at sites upstream of PRD. It is unclear whether fish spawned outside of the target areas because fish could return to a target location after being detected at a PIT tag array outside of the target stream without being detected again. Observations for PIT tagged presumptive Hanford Reach natural origin adults show very few detections above PRD.

Table 50 Last observations of unique PIT tagged adult fall Chinook from Priest Rapids Hatchery at detection sites upstream of McNary Dam, Brood Years 1999-2013.

|  | IT | Number of last known detections of unique Priest Rapids Origin PIT tags by site |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Brood Year | tagged | $\mathbf{M C N}$ | ICH | PRO | PRH | PRA | RIA | LWE | RRF | EBO | ENL | WEA | LMR | Total |
| 1999 | 3000 | 9 |  |  |  | 7 | 1 |  |  |  |  | 1 |  | 18 |
| 2000 | 3000 | 3 |  |  |  | 4 |  |  |  |  |  |  |  | 7 |
| 2001 | 3000 | 5 |  |  |  | 6 |  |  |  |  |  |  |  | 11 |
| 2002 | 3000 | 7 |  |  |  | 1 |  |  |  |  |  |  |  | 8 |
| 2003 | 3000 |  |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 2004 | 3000 |  |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 2005 | 3000 | 9 |  |  |  | 4 | 1 |  |  |  |  |  |  | 14 |
| 2006 | 3000 |  |  |  |  |  |  |  |  |  |  |  |  | 0 |
| 2007 | 3,000 | 20 |  |  | 1 | 12 | 2 |  | 2 |  |  | 1 | 1 | 39 |
| 2008 | 2,994 | 5 |  |  |  | 6 |  |  | 1 |  |  |  |  | 12 |
| 2009 | 1,995 | 4 |  |  | 8 | 8 | 2 |  |  |  |  |  |  | 22 |
| 2010 | 3,000 | 8 |  |  | 34 | 23 | 5 | 1 | 3 |  |  | 3 |  | 77 |
| 2011 (age 2-5) | 42,844 | 73 |  |  | 289 | 155 | 8 | 3 | 27 | 1 |  | 22 | 5 | 583 |
| 2012 (age 2-4) | 42,908 | 97 | 1 |  | 441 | 120 | 6 | 1 | 18 | 1 | 1 | 14 | 2 | 702 |
| 2013 (age 2-3) | 42,988 | 9 |  | 1 | 38 | 10 |  |  |  | 3 | 1 |  | 1 | 63 |
| MCN | McNary Da | $m$ Adult F | Fishway | s RKM |  |  |  | LWE | Lower W | enatchee | River | KM 754 |  |  |
| ICH | Ice Harbor D | Dam Adult | lt Fishw | ays RKM | - 522 |  |  | RRF | Rocky R | each Dan | Adult | ishway R | M 763 |  |
| PRO | Prosser Dive | ersion Da | m RKM | 539 |  |  |  | EBO | East Ban | $k$ Hatche | ry Outfa | 1 RKM 7 |  |  |
| PRH | Priest Rapid | s Hatcher | ry Outf | all RKM |  |  |  | ENL L | Lower E | ntiat Rive | er RKM |  |  |  |
| PRA | Priest Rapid | s Dam A | dult Fis | hway R | KM 639 |  |  | WEA | Well Da | m Adult | Fishway | RKM 83 |  |  |
| RIA | Rock Island | Dam Adut | ult Fish | ways RK | M 730 |  |  | LMR | Lower M | ethow R | iver at P | teros RK | M 843 |  |

### 16.0 Genetics

Genetic tissue was collected from each Chinook salmon spawned at PRH during 2016 by staff from the Columbia River Inter-Tribal Fish Commission (CRITFC). In total 4,938 specimens were collected to support their work associated with genetic stock identification and parentagebased tagging. Tissue samples were numbered consistent with PRH M\&E data so that biological information could be associated with genetic data. The tissue samples collected from return years 2011 through 2016 is currently being archived by CRITFC. During 2010, WDFW staff collected 100 genetic tissue samples from both the Priest Rapids Hatchery broodstock and naturally spawning broodstock from the Hanford Reach. WDFW has not collected genetic samples since the 2010 return because of the large sampling and archiving effort by CRITFC.

### 17.0 Proportion of Natural Influence

The intent of integrated hatchery programs is to achieve management objectives while having hatchery and natural origin fish share a common gene pool. Gene flow and the associated risks within and between the hatchery and natural environments can be estimated using a simple ratio estimator using the proportion of natural origin fish in the hatchery broodstock ( pNOB ) and the proportion of hatchery origin fish in the natural spawning escapement (pHOS). This ratio of $\mathrm{pNOB} /(\mathrm{pHOS}+\mathrm{pNOB})$ is termed the Proportionate Natural Influence (PNI). The larger the PNI
ratio, the greater selection that the natural environment has on the population relative to that of the hatchery environment. Alternatively, PNI estimates addressing gene flow from multiple sources/hatchery programs can be calculated from a multiple population gene flow model based on the Ford model which has been extended to three or more populations (Busack 2015, 2016).
In order for the natural environment to dominate selection, PNI for either calculation should be greater than 0.5 and for integrated hatchery programs the Hatchery Scientific Review Group (HSRG) recommends a PNI $\geq 0.67$ (HSRG/WDFW/NWIFC 2004). The HSRG recommends a minimum target of 0.15 for the proportion of natural origin Chinook salmon to be incorporated into the hatchery broodstock ( pNOB ) as well as a maximum target of 0.30 for the proportion of hatchery origin Chinook allowed to spawn in the natural environment (pHOS) for the Hanford Reach if it is to be managed as an integrated hatchery program.

Several estimates of PNI have been calculated to show the contributions of multiple programs on the overall PNI for the Hanford Reach. These programs include the hatchery production associated with the Grant PUD and USACE mitigation and the influence of strays. The different PNI estimates are based on pNOB and pHOS estimates specific to each source of spawning adults. The methods used to allocate pNOB and pHOS are described in the following sections.

### 17.1 Estimate of pNOB

Estimates of pNOB based on otolith samples are limited to return years 2012 through 2016. Otolith marking began with the 2007 brood. Therefore, otolith marks are only available for specific age classes of PRH origin fish during return years 2010 and 2011 and do not provide representative samples for estimating pNOB for the PRH broodstock.

The annual pNOB for fish spawned at PRH and used for Grant PUD and USACE smolt releases into the Hanford Reach during return years 2012 through 2016 is provided in Table 51.

Table 51 Origin of broodstock and pNOB apportioned to program for fall Chinook salmon spawned at Priest Rapids Hatchery, Return Years 2012-2016.

| Return Year | $\mathbf{N}$ | GCPUD pNOB | USACE pNOB | PRH and RSH <br> Combined pNOB | Other Programs $^{\text {pNOB }^{\mathbf{1}}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathbf{2 0 1 2}$ | 4,974 | 0.182 | 0.057 | 0.119 | N/A |
| $\mathbf{2 0 1 3}$ | 5,442 | 0.225 | 0.026 | 0.127 | N/A |
| $\mathbf{2 0 1 4}$ | 5,443 | 0.343 | 0.076 | 0.206 | 0.000 |
| $\mathbf{2 0 1 5}$ | 5,524 | 0.313 | 0.045 | 0.179 | 0.000 |
| $\mathbf{2 0 1 6}$ | 4,938 | 0.259 | 0.073 | 0.163 | 0.000 |
| Mean | 5,264 | 0.264 | 0.055 | 0.159 | 0.000 |

${ }^{1}$ Represents pHOB associated with egg-takes utilized outside of the Hanford Reach.
The 2016 broodstock included 4,938 adults which were comprised of 4,325 fish from the volunteer trap, 366 from the OLAFT and 247 from the ABC. In general, broodstock from ABC and OLAFT are held in a specific holding pond (Pond 4) and mated with fish from this pond or with high-graded fish collected from the PRH volunteer trap and held in another specific holding pond (Pond 1). The fish culturists segregate the progeny resulting from these matings for release from PRH. Brood stock utilized for non-Grant PUD programs are collected from the PRH volunteer trap and held in a specific pond (Pond 2). Large portions of the progeny from the Pond 2 broodstock are shipped to other facilities for use by other programs.

Grant PUD funds the collection of non-marked or tagged broodstock from the ABC and OLAFT with the intent of improving the pNOB associated with the production of their 5.6 million smolt
mitigation requirement. The inclusion of these fish contributed greatly to the Grant PUD program's egg-take goal and the resulting pNOB (Table 52).
Table 52 Origin of broodstock and pNOB apportioned to program for fall Chinook salmon spawned at Priest Rapids Hatchery, Brood Year 2016.

| Program | Egg-Take | Facility <br> Mean <br> Fecundity | Natural Females | Hatchery <br> Females | Natural Males | Hatchery Males | Total Natural | Total Hatchery | pNOB |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| GCPUD | 4,050,782 | 3,921 | 224 | 809 | 80 | 437 | 304 | 1246 | 0.196 |
| GCPUD Alt Mating ${ }^{1}$ | 2,220,002 | 3,369 | 162 | 497 | 153 | 24 | 315 | 521 | 0.377 |
| GCPUD Combined | 6,270,784 | 3,706 | 386 | 1,306 | 233 | 461 | 619 | 1,767 | 0.259 |
| USACE - PRH | 1,903,631 | 3,640 | 18 | 505 | 19 | 224 | 37 | 729 | 0.048 |
| USACE - RSH | 4,204,716 | 3,572 | 96 | 1,081 | 51 | 545 | 147 | 1,626 | 0.083 |
| USACE Combined | 6,108,347 | 3,593 | 114 | 1,586 | 70 | 769 | 184 | 2,355 | 0.072 |
| Combined PRH and RSH Programs | 12,379,131 | 3,392 | 500 | 2,892 | 303 | 1,230 | 803 | 4,122 | 0.163 |
| Other Programs ${ }^{2}$ | 32,399 | 3,600 | 0 | 9 | 0 | 5 | 0 | 14 | 0.000 |

${ }^{1}$ Alternative mating strategy incorporates 1 natural origin male x 4 females.
${ }^{2}$ Includes eggs from presumed hatchery $x$ hatchery crosses shipped to educational organizations.
An alternative pNOB for calculating PNI was developed to account for the genetic influence on pNOB resulting from the PRH spawning protocol of spawning one male with one, two, or four females. It is intended to represent actual gene flow to the progeny instead of strictly the origin and number of parents. This information is presented in Appendix K for comparison to other conventional pNOB calculations.

### 17.2 Estimates of pHOS

Estimates of pHOS based on otolith samples are limited to return years 2012 through 2016. Otolith marking began with the 2007 brood. Hence, otolith marks are only available for specific age classes of PRH origin fish during return years 2010 and 2011 and do not provide representative samples for estimating population level pHOS . The population level pHOS estimates for recent annual Hanford Reach spawning escapements are presented Table 53.
Table 53 Proportion of hatchery Chinook salmon on the spawning grounds ( $\mathbf{p H O S \text { ) in }}$ the Hanford Reach, Brood Years 2012-2016.

| Return <br> Year | N | Total Escapement | Hatchery Origin Spawners (pHOS) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1,609 | 57,631 | 0.062 | 0.066 | 0.005 | 0.135 |
| $\mathbf{2 0 1 3}$ | 927 | 126,744 | 0.203 | 0.054 | 0.018 | 0.275 |
| $\mathbf{2 0 1 4}$ | 2,426 | 183,750 | 0.052 | 0.015 | 0.028 | 0.096 |
| $\mathbf{2 0 1 5}$ | 2,485 | 266,347 | 0.076 | 0.017 | 0.004 | 0.097 |
| $\mathbf{2 0 1 6}$ | 1,648 | 116,421 | 0.066 | 0.022 | 0.027 | 0.115 |
| Mean | 1,819 | 150,179 | 0.092 | 0.035 | 0.016 | 0.144 |

${ }^{1}$ Includes fish from other hatcheries based on presence of a coded-wire tag or adipose clip fish without an otolith mark

Estimates for pHOS were calculated for contributing sources of hatchery origin fall Chinook salmon spawning naturally in the Hanford Reach. The primary source of pHOS originates from fish released from PRH. This source of PRH-pHOS was apportioned to the Grant PUD and USACE programs at PRH based on the annual mitigation requirement for the number of
juveniles released by each program between brood year 2008 and 2012 (Table 54). An estimated 10,245 PRH origin fish spawned naturally in the Hanford Reach during the 2016 return year. Of these, $74.6 \%$ and $25.4 \%$ were allocated respectively to Grant PUD and USACE production at PRH. The USACE's $25.4 \%$ portion of PRH origin pHOS was combined with the pHOS associated with the USACE's RSH production to estimate the total pHOS associated with the USACE programs in the Hanford Reach.

The calculation of pHOS specific to each program includes proportions which are based on the entire population of natural origin fish in the denominators. Therefore this method of calculating program specific pHOS results in lower values than the population level pHOS and may only be useful for assessing the individual program's contribution of hatchery origin fish to the spawning population in the natural environment.

Table $54 \quad \begin{aligned} & \text { Origin of pHOS apportioned by program source for fall Chinook salmon } \\ & \text { spawning naturally in the Hanford Reach, Return Years 2012-2016. }\end{aligned}$

| Return Year | Natural Origin | Hatchery Origin Spawners |  |  |  | pHOS by Source |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{aligned} & \text { Grant } \\ & \text { PUD }^{1} \end{aligned}$ | USACE ${ }^{\mathbf{1 , 2}}$ | Other ${ }^{3}$ | Total | $\begin{aligned} & \text { Grant } \\ & \text { PUD }{ }^{1} \end{aligned}$ | $\text { USACE }{ }^{1,2}$ | Other ${ }^{3}$ | Combined |
| 2012 | 50,072 | 3,943 | 3,598 | 261 | 7,803 | 0.068 | 0.062 | 0.005 | 0.135 |
| 2013 | 126,782 | 26,507 | 18,427 | 3,123 | 48,057 | 0.152 | 0.105 | 0.018 | 0.275 |
| 2014 | 166,183 | 7,185 | 5,262 | 5,120 | 17,567 | 0.039 | 0.029 | 0.028 | 0.096 |
| 2015 | 240,511 | 15,101 | 9,669 | 1,065 | 25,835 | 0.057 | 0.036 | 0.004 | 0.097 |
| 2016 | 103,033 | 5,732 | 4,513 | 3,143 | 13,388 | 0.049 | 0.039 | 0.027 | 0.115 |

${ }^{1}$ Estimated number of PRH origin fish that spawned naturally in the Hanford Reach. Of these, $74.6 \%$ and $25.4 \%$ were apportioned to Grant PUD-PRH and USACE-PRH, respectively. The allocation of pHOS was based on the proportion of annual juvenile mitigation goals for each agency for brood years 2008 through 2012.
${ }^{2}$ Includes hatchery origin fish released from Ringold Springs Hatchery.
${ }^{3}$ Includes hatchery origin fish released from other hatcheries based on the presence of a hatchery mark without an otolith mark.

### 17.3 Estimates of PNI

We present a hierarchy of PNI estimates based on pNOB and pHOS values calculated to reflect differing methodologies driven by the type of data available to assign origin of adult Chinook salmon returns. The population level PNI for the Hanford Reach includes all hatchery origin fish regardless of hatchery program or funding source.
Prior to return year 2012, $\mathrm{pHOS}, \mathrm{pNOB}$ and PNI rates are based on CWT recoveries from the adult returns. Historically, we used juvenile mark rate expansions of CWT recoveries in the hatchery and stream surveys for these calculations. The pNOB estimated from CWT requires the assumption that fish unaccounted for by the juvenile mark rate expansions are natural origin fish. As discussed in Section 10 of this report, this assumption significantly over estimates pNOB and PNI. This method of estimated pNOB for the 2015 and 2016 broodstock was not calculated due to high-grading to remove fish possessing a CWT and or an adipose clip. Hence, the broodstock origin is poorly represented by CWT.
The pHOS estimates based on juvenile mark rate expansions of CWT recoveries also likely underestimate the presences of PRH and RSH origin fish as explained in Section 10. For comparison, we present CWT based estimates of PNI derived from CWT adult-to-adult expansions for PRH and RSH origin adult recoveries at their respective hatcheries. An
explanation of methods is given in Appendix L. Estimates of pNOB, pHOS, and PNI based on both methods of CWT expansions are presented in Table 55.

The pHOS and pNOB estimates from limited otolith datasets for recent complete brood years are more similar to the estimates produced by adult-to-adult CWT expansions versus juvenile mark rate expansions of CWT recoveries of returning adults.
Table 55 PNI of the Hanford Reach fall Chinook salmon supplementation program based on expanded coded wire-tag recoveries of all fish surveyed, Return Year 2001-2016.

| Return Year | pNOB ${ }^{1}$ | pHOS ${ }^{1}$ | pNOB ${ }^{2}$ | pHOS ${ }^{2}$ | PNI based on $\mathrm{pNOB}^{1}$ and $\mathrm{pHOS}^{1}$ | PNI based on $\mathrm{pNOB}^{2}$ and $\mathrm{pHOS}^{2}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0.155 | 0.094 | 0.046 | 0.066 | 0.622 | 0.411 |
| 2002 | 0.145 | 0.101 | 0.046 | 0.125 | 0.589 | 0.269 |
| 2003 | 0.132 | 0.099 | 0.046 | 0.117 | 0.571 | 0.282 |
| 2004 | 0.229 | 0.081 | 0.046 | 0.099 | 0.739 | 0.317 |
| 2005 | 0.370 | 0.106 | 0.046 | 0.156 | 0.777 | 0.229 |
| 2006 | 0.507 | 0.057 | 0.046 | 0.124 | 0.899 | 0.271 |
| 2007 | 0.326 | 0.041 | 0.046 | 0.065 | 0.888 | 0.414 |
| 2008 | 0.501 | 0.046 | 0.046 | 0.087 | 0.916 | 0.346 |
| 2009 | 0.568 | 0.077 | 0.046 | 0.174 | 0.881 | 0.209 |
| 2010 | 0.392 | 0.040 | 0.046 | 0.076 | 0.907 | 0.377 |
| 2011 | 0.381 | 0.076 | 0.046 | 0.154 | 0.836 | 0.230 |
| 2012 | 0.304 | 0.074 | $0.119^{\text {a }}$ | 0.118 | 0.871 | 0.529 |
| 2013 | 0.252 | 0.217 | $0.127^{\text {a }}$ | 0.287 | 0.537 | 0.300 |
| 2014 | 0.443 | 0.054 | $0.206^{\text {a }}$ | 0.069 | 0.888 | 0.760 |
| 2015 | N/A ${ }^{3}$ | 0.072 | $0.179^{\text {a }}$ | 0.075 | N/A ${ }^{3}$ | 0.691 |
| 2016 | N/A ${ }^{3}$ | 0.092 | $0.163^{\text {a }}$ | 0.097 | N/A ${ }^{3}$ | 0.627 |

pNOB ${ }^{1}$ Assumes that all fish not accounted for by juvenile coded-wire tag expansions are natural origin.
pHOS ${ }^{1}$ based on hatchery origin coded-wire recoveries expanded by juvenile mark rate and survey sample rate.
$\mathrm{pNOB}^{2}$ is assigned to years $2001-2011$ based on an average proportion of natural origin returns to PRH for return years 2012-2014 as determined by otolith and other hatchery marks.
$\mathrm{pHOS}^{2}$ is based on an adult coded-wire tag expansion rate for PRH and RSH origin adults recovered in the Hanford Reach escapement combined with juveniles coded-wire tag mark rate expansions for other hatchery strays. Both groups were expanded by the survey sample rate.
${ }^{3}$ Brood stock was generally high-graded to remove coded-wire tagged fish during ponding.
${ }^{a}$ pNOB of broodstock used for production of PRH and RSH programs as determined from otoliths and other hatchery marks.

For return years 2012-2016 we present PNI estimates calculated from the multiple population gene flow model (Table 56). The output from this model suggests that the PNI values for return years 2014 and 2016 have exceeded the goal of 0.670 .

Table 56 PNI estimates for the Hanford Reach fall Chinook salmon supplementation programs, Return Years 2012-2016. Calculated from multiple population gene flow model based on the Ford model which has been extended to three or more populations.

| Return <br> Year | pNOB |  |  | pHOS |  |  | pHOS <br> Reach ${ }^{7}$ | $\begin{gathered} \text { PNI } \\ \text { Population }^{8} \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | GCPUD ${ }^{1}$ | USACE ${ }^{2}$ | Facility ${ }^{3}$ | GCPUD ${ }^{4}$ | USACE ${ }^{5}$ | Other ${ }^{6}$ |  |  |
| 2012 | 0.182 | 0.057 | 0.119 | 0.068 | 0.062 | 0.005 | 0.135 | 0.599 |
| 2013 | 0.225 | 0.027 | 0.127 | 0.152 | 0.105 | 0.018 | 0.275 | 0.463 |
| 2014 | 0.343 | 0.076 | 0.206 | 0.039 | 0.029 | 0.028 | 0.096 | 0.775 |
| 2015 | 0.313 | 0.045 | 0.179 | 0.057 | 0.036 | 0.004 | 0.097 | 0.762 |
| 2016 | 0.259 | 0.072 | 0.163 | 0.049 | 0.039 | 0.027 | 0.115 | 0.700 |
| Mean | 0.265 | 0.055 | 0.159 | 0.073 | 0.054 | 0.016 | 0.144 | 0.660 |

${ }^{1}$ Includes broodstock associated with GCPUD production at PRH.
${ }^{2}$ Includes broodstock associated with USACE production at PRH and RSH.
${ }^{3}$ Includes broodstock spawned at PRH for all production
${ }^{4}$ Includes pHOS associated with GCPUD mitigation smolt releases at PRH
${ }^{5}$ Includes pHOS associated with USACE mitigation smolt releases at PRH and RSH
${ }^{6}$ Includes pHOS associated with strays from hatcheries outside of the Hanford Reach
${ }^{7}$ Population level pHOS in the Hanford Reach
${ }^{8}$ Population level PNI for the Hanford Reach. Assumes strays from hatcheries outside of the Hanford Reach have an associated pNOB of zero.

### 18.0 Natural and Hatchery Replacement Rates

The numbers of hatchery origin recruits (HOR) were estimated from CWT recoveries for brood year returns to the PRH and the Hanford Reach of the Columbia River. The recovered CWTs are expanded by sample rate of the survey and then by the juvenile tag rate. CWTs recovered from natural origin recruits (NOR) originating from the Hanford Reach are difficult to expand accurately because the juvenile tag rates are unknown. Therefore, the assumption was made that returns not accounted for by HOR CWT recoveries are NOR. Recent data indicates that that CWT data likely underestimates the true number of HORs; Hence, our assumption likely overestimates the number of NOR.

Hatchery replacement rates (HRR) were calculated as the ratio of HOR to the parent broodstock at PRH. This broodstock is an estimate of the number of fish spawned at PRH to produce the target release of subyearling fall Chinook salmon. Similarly, natural replacement rates (NRR) for the Hanford Reach URB fall Chinook salmon were calculated as the ratio of NOR to the parent population spawning naturally in the Hanford Reach natural environment. This spawning population is based on the escapement estimate to the Hanford Reach without adjustments for spawn success.
Harvest estimates for HOR were calculated from the proportion of the expanded CWT recoveries in the fisheries to the total number of the expanded CWTs recoveries included in fisheries, stream surveys, and hatchery racks. The CWT recoveries are expanded by sample rate of the survey and juvenile mark rate for the CWT group. Since there is not a CWT mark rate for NOR, the harvest rates for PRH origin returns (HOR) were used as an indicator for similar brood years of NOR.

For brood years 1996 through 2010, the HRR (14.24) has been consistently higher than the NRR (3.16) (Table 57). The HRR for BY 2010 including harvest was the highest that has been observed (61.69) and was substantially higher than the NRR (6.21). The HRR should be greater than or equal to 5.30 (the target value in Murdoch and Peven 2005).
Table 57 Broodstock spawned at Priest Rapids Hatchery, estimated escapement to the Hanford Reach, natural and hatchery origin recruits (NOR and HOR), and natural and hatchery replacement rates (NRR and HRR, with and without harvest) for natural origin fall Chinook salmon in the Hanford Reach, Brood Years 1996-2010.

| Brood Year | Broodstock Spawned | Hanford Reach Escapement ${ }^{1}$ | Harvest not included |  |  |  | Harvest included ${ }^{2}$ |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | HOR | NOR | HRR | NRR | HOR | NOR | HRR | NRR |
| 1996 | 2,859 | 43,249 | 13,584 | 28,849 | 4.75 | 0.67 | 26,205 | 59,899 | 9.17 | 1.38 |
| 1997 | 2,726 | 43,493 | 3,002 | 44,416 | 1.10 | 1.02 | 6,037 | 88,349 | 2.21 | 2.03 |
| 1998 | 3,027 | 35,393 | 18,464 | 93,999 | 6.10 | 2.66 | 31,932 | 222,865 | 10.55 | 6.30 |
| 1999 | 2,619 | 29,812 | 27,093 | 115,237 | 10.34 | 3.87 | 52,099 | 240,090 | 19.89 | 8.05 |
| 2000 | 2,619 | 48,020 | 4,665 | 56,422 | 1.78 | 1.17 | 12,508 | 89,983 | 4.78 | 1.87 |
| 2001 | 3,621 | 59,848 | 25,059 | 71,359 | 6.92 | 1.19 | 55,789 | 129,548 | 15.41 | 2.16 |
| 2002 | 3,630 | 84,509 | 5,277 | 47,813 | 1.45 | 0.57 | 12,744 | 81,600 | 3.51 | 0.97 |
| 2003 | 3,003 | 100,508 | 3,021 | 31,788 | 1.01 | 0.32 | 5,974 | 64,307 | 1.99 | 0.64 |
| 2004 | 3,014 | 87,696 | 1,109 | 22,747 | 0.37 | 0.26 | 3,262 | 34,465 | 1.08 | 0.39 |
| 2005 | 2,898 | 71,967 | 21,107 | 64,011 | 7.28 | 0.89 | 61,122 | 97,777 | 21.09 | 1.36 |
| 2006 | 2,911 | 51,701 | 998 | 54,288 | 0.34 | 1.05 | 3,347 | 77,344 | 1.15 | 1.50 |
| 2007 | 2,096 | 22,274 | 22,453 | 101,753 | 10.71 | 4.57 | 53,685 | 174,905 | 25.61 | 7.85 |
| 2008 | 2,959 | 29,058 | 11,935 | 41,809 | 4.03 | 1.44 | 25,234 | 79,330 | 8.53 | 2.73 |
| 2009 | 3,177 | 36,720 | 28,197 | 97,834 | 8.88 | 2.66 | 85,533 | 145,949 | 26.92 | 3.97 |
| 2010 | 3,348 | 86,982 | 99,402 | 280,205 | 29.69 | 3.22 | 206,551 | 540,151 | 61.69 | 6.21 |
| Mean | 2,967 | 55,415 | 19,024 | 76,835 | 6.32 | 1.70 | 42,801 | 141,771 | 14.24 | 3.16 |
| Median | 2,959 | 48,020 | 13,584 | 56,422 | 4.75 | 1.17 | 26,205 | 89,983 | 9.17 | 2.03 |

${ }^{1}$ Includes estimated adult and jack escapement to the Hanford Reach natural environment.
${ }^{2}$ Harvest rates for NORs was estimated using the HRRs harvest rates for similar brood years as an indicator stock.

### 19.0 Smolt-to-Adult Survivals

Smolt-to-adult survival ratios (SAR) were calculated by dividing the expanded number of adults possessing a CWT recovered by the number of CWT smolts released. This estimate could be biased low for both hatchery and natural origin fish because of some of CWT bias identified previously in this report. The following data was obtained from the RMPC's RMIS online database: http://www.rmpc.org/. The 2010 brood year data was queried on January 16, 2017. This query should account for age 2 through 5 fall Chinook salmon sampled through December 2015. The lag in reporting field data for the 2016 return year likely excludes recoveries of a limited number of age-6 fish from the 2010 brood.

Annual SAR for hatchery fall Chinook salmon released from PRH for brood years 1992 through 2010 have a mean of 0.0062 with a median of 0.0040 (

Table 58). The SAR for the PRH origin 2010 brood is 0.0304 which is the highest SAR on record for PRH releases.

Table 58 Smolt-to-adult-Survival Ratios (SAR) for Priest Rapids Hatchery fall Chinook salmon, Brood Years 1992-2010. Data includes coded-wire tag recoveries from adipose clipped fish.

| Brood Year | Number of Tagged <br> Smolts Released | Estimated Adult Captures | SAR |
| :---: | :---: | :---: | :---: |
| 1992 | 194,622 | 448 | 0.0023 |
| 1993 | 185,683 | 1,479 | 0.0080 |
| 1994 | 175,880 | 108 | 0.0006 |
| 1995 | 196,189 | 1,786 | 0.0091 |
| 1996 | 193,215 | 762 | 0.0040 |
| 1997 | 196,249 | 183 | 0.0009 |
| 1998 | 193,660 | 946 | 0.0049 |
| 1999 | 204,346 | 1,573 | 0.0077 |
| 2000 | 200,779 | 370 | 0.0018 |
| 2001 | 219,926 | 1,810 | 0.0082 |
| 2002 | 355,373 | 669 | 0.0019 |
| 2003 | 399,116 | 352 | 0.0009 |
| 2004 | 200,072 | 100 | 0.0005 |
| 2005 | 199,445 | 1,718 | 0.0086 |
| 2006 | 202,000 | 100 | 0.0005 |
| 2007 | 202,568 | 2,391 | 0.0118 |
| 2008 | 218,082 | 740 | 0.0034 |
| 2009 | 619,568 | 7,820 | 0.0126 |
| 2010 | 605,000 | 18,620 | 0.0308 |
| Mean | $\mathbf{2 6 1 , 1 4 6}$ | $\mathbf{2 , 2 0 9}$ | $\mathbf{0 . 0 0 6 2}$ |
| Median | $\mathbf{2 0 0 , 7 7 9}$ | $\mathbf{7 6 2}$ | 0.0040 |

Annual SAR for Hanford Reach natural origin fall Chinook salmon for brood years 1992 through 2010 had a mean of 0.0042 with a median of 0.0023 (Table 59). The SAR for the Hanford Reach natural origin 2010 brood is 0.0164 which is the highest SAR on record for the Hanford Reach natural origin stock.

Table 59 Smolt-to-adult-Survival Ratios (SAR) for Hanford Reach natural origin fall Chinook salmon, Brood Years 1992-2010. Data includes coded-wire tag recoveries from adipose clipped fish.

| Brood Year | Number of Tagged <br> Smolts Released | Estimated Adult Captures | SAR |
| :---: | :---: | :---: | :---: |
| 1992 | 203,591 | 829 | 0.0041 |
| 1993 | 95,897 | 485 | 0.0051 |
| 1994 | 148,585 | 74 | 0.0005 |
| 1995 | 146,887 | 340 | 0.0023 |
| 1996 | 92,262 | 111 | 0.0012 |
| 1997 | 199,896 | 365 | 0.0018 |
| 1998 | 129,850 | 784 | 0.0060 |
| 1999 | 213,259 | 2,378 | 0.0112 |
| 2000 | 204,925 | 362 | 0.0018 |
| 2001 | 127,758 | 519 | 0.0041 |
| 2002 | 203,557 | 338 | 0.0017 |
| 2003 | 207,168 | 199 | 0.0010 |
| 2004 | 163,884 | 147 | 0.0009 |
| 2005 | 203,929 | 301 | 0.0015 |
| 2006 | 263,478 | 356 | 0.0007 |
| 2007 | 53,618 | 456 | 0.0085 |
| 2008 | 203,947 | 520 | 0.0025 |
| 2009 | 201,606 | 1,597 | 0.0079 |
| 2010 | 179,727 | $\mathbf{1 7 0 , 7 2 8}$ | $\mathbf{6 9 0}$ |
| Mean | $\mathbf{1 9 9 , 8 9 6}$ | $\mathbf{3 6 5}$ | 0.0164 |
| Median |  |  | $\mathbf{0 . 0 0 4 2}$ |

### 20.0 ESA/HCP Compliance

### 20.1 Broodstock Collection

Section 10(a)(1)(B) Permit 1347 authorizes collection of fall Chinook broodstock at the OLAFT for the Priest Rapids hatchery program with an incidental take limit of 10 steelhead (an aggregate of hatchery or wild). Due to the absence of an identified steelhead take limit for operation of the PRH volunteer trap in permit 1347 and through ongoing coordination with NOAA Fisheries, the 10 fish take limit for broodstock collection at the OLAFT, on an interim basis (until a new permit is issued), has been re-conceptualized to include broodstock collection at the PRH volunteer trap, and in the ABC fishery. During the 2016 fall Chinook broodstock collection activities, a total of 5 steelhead were encountered at the PRH volunteer trap with no incidental mortality reported. No steelhead mortalities were associated with broodstock collection at the OLAFT or in the ABC fishery (Table 60).

Table 60 Recoveries and disposition of steelhead at the Priest Rapids Hatchery volunteer trap, Return Year 2016.

|  |  | No Mark | Ad Only | Ad-RV | Total |
| :---: | :--- | :---: | :---: | :---: | :---: |
| Released | Males | 1 | 0 | 0 | 1 |
|  | Females | 1 | 1 | 2 | 4 |
|  | Sub Total | 2 | 1 | 2 | 5 |
| Killed | Males | 0 | 0 | 0 | 0 |
|  | Females | 0 | 0 | 0 | 0 |
|  | Sub Total | 0 | 0 | 0 | 0 |
| Total |  | 2 | 1 | 2 | 5 |

### 20.2 Hatchery Rearing and Release

The juvenile fall Chinook salmon from the 2016 brood year reared throughout their life-stages at PRH without incident. The 2017 smolt release totaled 7,022,271 URB fall Chinook salmon, representing $96 \%$ of the production objective and was compliant with the $10 \%$ overage allowable in ESA Section 10 Permit 1347.

### 20.3 Distribution of Surpluses, Mortalities, and Spawned, Adult fall Chinook Salmon from Priest Rapids Hatchery

All adult Chinook salmon recovered at PRH are eventually distributed to multiple organizations depending on the condition and treatment of the individual fish while at the hatchery. A large majority of these fish are suitable for consumption and transported to Foodbanks (Table 61). In $2016,72 \%$ of the surplus fish were used for human consumption.

## Table 61 Disposition of Chinook salmon removed from Priest Rapids Hatchery

 volunteer trap, Return Years 2001-2016.| Return Year | Dispo <br> Mortal Treate <br> Pet <br> Food | al of ies and Fish <br> Landfill | WDFW <br> Nutrient Enhancement Projects | Donations to Educational Programs \& Research | $\begin{array}{\|c} \text { Donations } \\ \text { to } \\ \text { Foodbanks } \\ \hline \end{array}$ | Donations to Tribes | Sold to Fish Buyers | Fish Removed from Priest Rapids Hatchery |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 6,597 | 2,054 | 0 |  | 525 | 6,139 | 15,315 |
| 2002 | 0 | 6,572 | 2,192 | 0 | 3,130 | 502 | 0 | 12,396 |
| 2003 | 0 | 5,144 | 3,211 | 9 | 881 | 98 | 0 | 9,343 |
| 2004 | 350 | 2,661 | 2,756 | 88 | 9,371 |  | 595 | 15,821 |
| 2005 | 153 | 5,635 | 318 | 2 | 0 |  | 4,503 | 10,611 |
| 2006 | 0 | 5,467 | 0 | 250 | 0 | 340 | 2,146 | 8,203 |
| 2007 | 2,595 | 0 | 0 | 0 | 0 | 159 | 3,345 | 6,099 |
| 2008 | 5,384 | 90 | 0 | 340 | 0 | 375 | 13,42 | 19,617 |
| 2009 | 5,846 | 0 | 0 | 310 | 0 | 201 | 6,502 | 12,859 |
| 2010 | 5,412 | 1,937 | 1,937 | 452 | 3,548 | 8 | 8,259 | 21,553 |
| 2011 | 6,951 | 0 | 1,500 | 412 | 11,217 | 588 | 0 | 20,668 |
| 2012 | 7,554 | 0 | 0 | 460 | 20,628 |  | 0 | 28,642 |
| 2013 | 10,108 | 0 | 0 | 489 | 31,647 | 626 | 0 | 42,870 |
| 2014 | 10,805 | 0 | 0 | 237 | 67,684 | 783 | 0 | 79,509 |
| 2015 | 7,402 | 0 | 0 | 398 | 52,987 | 4,228 | 0 | 65,015 |
| 2016 | 7,833 | 0 | 0 | 411 | 19,424 | 1,948 | 0 | 29,616 |
| Mean | 4,400 | 2,131 | 873 | 241 | 14,701 | 799 | 2,807 | 24,884 |
| Median | 5,398 | 45 | 0 | 280 | 3,548 | 502 | 298 | 17,719 |

### 20.4 Hatchery Effluent Monitoring

Per ESA Permits 1196, 1347, and 1395, permit holders shall monitor and report hatchery effluents in compliance with applicable National Pollution Discharge Elimination Systems (NPDES) (EPA 1999) permit limitations. There were no NPDES violations reported at Grant PUD Hatchery facilities during the September 2016 through June 2017 collection and rearing periods.

### 20.5 Ecological Risk Assessment

One of the regional objectives in the Grant PUD M\&E plan is to conduct an ecological risk assessment on non-target taxa of concern to determine if additional M\&E is necessary (Pearsons and Langshaw 2009). The methodology that was used to assess risks was presented in Pearsons et al. (2012) and Pearsons and Busack (2012). This objective was completed through an approved report that summarized the methods and results of the risk assessment (Mackey et al. 2014).

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## Appendix A Evaluation of Coded-Wire Tag Bias

We annually evaluate bias associated with estimates of the number of hatchery origin returns to PRH generated using coded-wire tags (CWT). Results from demographic sampling of the fall Chinook returns for 2010 through 2014 indicate that estimates of hatchery contributions to broodstock, the terminal sport fishery, and to escapement of the Hanford Reach and to the PRH trap calculated from otolith marks were substantially different from estimates generated using CWTs expanded by sampling rates and juvenile mark rates. This was of significant concern because many estimates such as stray rate, survival, origin, and harvest are dependent upon estimates generated from CWTs.

To assess the level of CWT recovery bias for any brood year, we used the following equation:

$$
\text { CWT Recovery Bias }=\frac{(\# \text { of PRH Origin CWT Fish Recovered } / \# \text { of PRH Origin Fish Collected })}{\text { CWT Mark Rate for Brood Year }}
$$

Where:
\# of PRH origin fish collected = Estimate of the number of PRH origin fish for a specific age/brood year as determined by otoliths, scale aging, and expansion and pooling of age samples to represent total returns by age
\# of PRH Origin CWT Fish Recovered = Number of PRH origin CWT fish for a specific age/brood recovered at the hatchery ( $100 \%$ sample rate)
CWT Mark Rate $=$ CWT marking rate for the specific brood year which is the number of CWT placed in fish divided by the estimated total number of fish at the time of marking.
If no CWT bias exists, the proportion of PRH CWT returns to the PRH CWT mark rate should equal 1.000. The values for CWT Recovery bias ranged from 0.573 to 2.026 for the different age/broods examined (Table 1). CWT Recovery bias appears to be reduced for brood years 2012 and 2013 relative to that of earlier years. The source of any bias is unknown but several factors may contribute to the variation in CWT Recovery bias shown in Table 1: 1) tag loss, 2) CWT detection efficiency, 3) inappropriate expansion estimates, or 4) differential survival of tagged fish. The estimate of bias may be influenced by the level of precision of the estimated \# of PRH origin fish collected which varies for each age class of a given brood year due to size of the otolith sub-sample pulled from the demographic sample. In some cases, there are relatively few samples for age- 2 and 5 fish for a given brood year for this estimate. Verification of the juvenile CWT rate at time of release is necessary to determine level of potential bias associated with reported juvenile CWT rates. Sampling for CWT rates at time of release has occurred at PRH since brood year 2014. On the day prior to release, roughly 1,000 subyearlings from each of the five rearing ponds were captured and scanned with a V-detector to determine the proportions of adipose clipped CWT fish and adipose intact CWT fish within the sample. These proportions at release were compared to the proportions reported as ponded. In general, these two groups of proportions are similar for each brood year (Table 2).

Table A. 1 Estimate of coded-wire tags bias for Priest Rapids origin returns to the hatchery, Brood Years 2007-2014.
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| Brood | Age | Proporti on CWT Marked | \# of PRH Origin <br> CWT Fish <br> Recovered | Estimate d \# of PRH origin Fish Collected | Proportion of PRH Origin Brood Return CWT | Proportion of PRH CWT Returns to the PRH CWT Mark Rate (CWT <br> Recovery Bias) | Primary Detector Type |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 5 | 0.045 | 48 | 928 | 0.052 | 1.161 | Blue Wand |
| 2007 | 4 | 0.045 | 280 | 10,977 | 0.026 | 0.573 | Blue Wand |
| 2007 | 3 | 0.045 | 410 | 14,073 | 0.029 | 0.654 | Blue Wand |
| 2007 | 2 | No otolith data collected during return year 2009 |  |  |  |  |  |
| 2008 | 5 | 0.032 | 2 | 31 | 0.065 | 2.026 | Blue Wand |
| 2008 | 4 | 0.032 | 81 | 3,029 | 0.027 | 0.840 | Blue Wand |
| 2008 | 3 | 0.032 | 124 | 5,606 | 0.022 | 0.695 | Blue Wand |
| 2008 | 2 | 0.032 | 57 | 2,578 | 0.022 | 0.694 | Blue Wand |
| 2009 | 5 | 0.243 | 407 | 1,980 | 0.206 | 0.846 | R9500 |
| 2009 | 4 | 0.243 | 1,081 | 6,025 | 0.179 | 0.739 | Blue Wand |
| 2009 | 3 | 0.243 | 2,309 | 13,713 | 0.168 | 0.693 | Blue Wand |
| 2009 | 2 | 0.243 | 628 | 3,083 | 0.204 | 0.839 | Blue Wand |
| 2010 | 6 | 0.237 | 23 | 21 | 1.095 | 4.620 | R9500 |
| 2010 | 5 | 0.237 | 999 | 2,375 | 0.421 | 1.774 | R9500 |
| 2010 | 4 | 0.237 | 8,719 | 39,621 | 0.220 | 0.928 | R9500 |
| 2010 | 3 | 0.237 | 5,828 | 32,014 | 0.182 | 0.768 | Blue Wand |
| 2010 | 2 | 0.237 | 1,498 | 8,932 | 0.168 | 0.707 | Blue Wand |
| 2011 | 5 | 0.169 | 395 | 2,561 | 0.154 | 0.912 | R9500 |
| 2011 | 4 | 0.169 | 2,988 | 19,909 | 0.150 | 0.887 | R9500 |
| 2011 | 3 | 0.169 | 2,596 | 19,692 | 0.132 | 0.779 | R9500 |
| 2011 | 2 | 0.169 | 349 | 3,008 | 0.116 | 0.686 | R9500 |
| 2012 | 4 | 0.177 | 2206 | 13,821 | 0.160 | 0.904 | R9500 |
| 2012 | 3 | 0.177 | 5,933 | 34,082 | 0.174 | 0.986 | R9500 |
| 2012 | 2 | 0.177 | 1,910 | 11,259 | 0.170 | 0.961 | R9500 |
| 2013 | 3 | 0.166 | 1,805 | 10,967 | 0.165 | 0.991 | R9500 |
| 2013 | 2 | 0.166 | 545 | 3,327 | 0.164 | 0.986 | R9500 |
| 2014 | 2 | 0.172 | 78 | 486 | 0.160 | 0.935 | R9500 |
| CWT <br> Recovery Bias |  | Ages |  |  |  |  |  |
|  |  | Mean | 0.830 | 0.795 | 0.812 | 1.344 | 4.620 |
|  |  | Median | 0.839 | 0.768 | 0.864 | 1.161 | 4.620 |

Table A.2. Proportions of coded-wire tagged fish reported ponded and the proportions of coded-wire tagged fish sampled at time of release, Brood Years 2014-15.

| Coded-wire sampling at release, Brood Year 2014 |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| \# of Fish | Pond E | Pond D | Pond C | Pond B | Pond A | Total |  |
| Fish Released | $1,425,371$ | $1,457,198$ | $1,400,956$ | $1,444,918$ | $1,311,100$ | $7,039,543$ |  |
| N $=$ | 1,040 | 1,024 | 1,018 | 1,023 | 1,565 | 5,670 |  |
| CWT Only Sampled | 98 | 85 | 79 | 67 | 220 | 549 |  |
| Ad-CWT Sampled | 102 | 69 | 73 | 86 | 165 | 495 |  |
| Proportion of Release Tagged |  |  |  |  |  |  |  |
| CWT Only | $8.5 \%$ | $8.3 \%$ | $8.6 \%$ | $8.2 \%$ | $9.0 \%$ | $8.5 \%$ |  |
| Ad-CWT | $8.5 \%$ | $8.2 \%$ | $8.6 \%$ | $8.7 \%$ | $8.7 \%$ | $8.5 \%$ |  |
| Proportion of Sample Tagged |  |  |  |  |  |  |  |
| CWT Only | $9.4 \%$ | $8.3 \%$ | $7.8 \%$ | $6.5 \%$ | $14.1 \%$ | $9.7 \%$ |  |
| Ad-CWT | $9.8 \%$ | $6.7 \%$ | $7.2 \%$ | $8.4 \%$ | $10.5 \%$ | $8.7 \%$ |  |


| Coded-wire sampling at release, Brood Year 2015 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| \# of Fish | Pond E | Pond D | Pond C | Pond $B$ | Pond A | Total |
| Fish Released | 1,445,733 | 1,448,510 | 1,507,753 | 1,512,437 | 1,327,621 | 7,242,054 |
| $\mathrm{N}=$ | 1,015 | 995 | 991 | 1,048 | 1,021 | 5,070 |
| CWT Only Sampled | 91 | 86 | 77 | 62 | 76 | 392 |
| Ad-CWT Sampled | 71 | 87 | 79 | 71 | 80 | 388 |
| Proportion of Release Tagged |  |  |  |  |  |  |
| CWT Only | 8.1\% | 8.6\% | 8.3\% | 7.5\% | 9.1\% | 8.3\% |
| Ad-CWT | 8.3\% | 8.6\% | 7.7\% | 8.0\% | 9.1\% | 8.3\% |
| Proportion of Sample Tagged |  |  |  |  |  |  |
| CWT Only | 9.0\% | 8.6\% | 7.8\% | 5.9\% | 7.4\% | 7.7\% |
| Ad-CWT | 7.0\% | 8.7\% | 8.0\% | 6.8\% | 7.8\% | 7.7\% |


| Coded-wire sampling at release, Brood Year 2016 |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| \# of Fish | Pond E | Pond D | Pond C | Pond B | Pond A | Total |  |
| Fish Released | $1,401,157$ | $1,455,960$ | $1,450,776$ | $1,487,339$ | $1,211,019$ | $7,006,251$ |  |
| $\mathrm{~N}=$ | 1,031 | 1,317 | 2,228 | 1,117 | 1,181 | 6,874 |  |
| CWT Only Sampled | 119 | 103 | 205 | 116 | 120 | 663 |  |
| Ad-CWT Sampled | 101 | 96 | 224 | 112 | 117 | 650 |  |
| Proportion of Release Tagged |  |  |  |  |  |  |  |
| CWT Only | $8.6 \%$ | $8.3 \%$ | $8.3 \%$ | $8.1 \%$ | $10.0 \%$ | $8.6 \%$ |  |
| Ad-CWT | $8.6 \%$ | $8.3 \%$ | $8.3 \%$ | $8.1 \%$ | $10.0 \%$ | $8.6 \%$ |  |
| CWT Only | $11.5 \%$ | $7.8 \%$ | $9.2 \%$ | $10.4 \%$ | $10.2 \%$ | $9.6 \%$ |  |
| Ad-CWT | $9.8 \%$ | $7.3 \%$ | $10.1 \%$ | $10.0 \%$ | $9.9 \%$ | $9.5 \%$ |  |

Assessment of CWT detection efficiency has been conducted annually at PRH since 2010 during adult fish sampling with enhancement to these procedures developed over time. In 2013, M\&E staff randomly selected a total of 1,063 quality control fish being surplused with no CWT
detected using the T-wand (Table 3). These fish were then re-scanned with the older blue-wand. If CWT was detected using a blue wand the fish was again scanned using the T-wand. In such a manner the missed CWT could be inferred as a result of operator error or the inability of the Twand to detect the CWT. On a few occasions the T-wand did not detect a CWT identified by the blue-wand. In these instances, the snouts were removed from the fish to increase the likelihood of detection and then passed through a V-detector. Similar to quality control results for previous years, there were only a few ( 4 tags; $0.4 \%$ of the sample CWT detections observed in the quality control fish sampled that were not detected initially by the T-wands.

Table A. 3 Quality control results for coded-wire tag detection at Priest Rapids Hatchery, Brood Years 2013- 2016.

| Brood Year | Initial Device | QC Device | \# Sampled | \# Missed CWT | P $^{\wedge}$ CWT Missed |
| ---: | :---: | :---: | :---: | :---: | :---: |
| 2013 | T-Wand | Blue Wand | 1,063 | 4 | 0.004 |
| 2014 | R9500 | T-Wand | 2,000 | 3 | 0.002 |
| 2015 | R9500 | T-Wand | 4,596 | 2 | 0.000 |
| 2016 | R9500 | T-Wand | 5,943 | 3 | 0.001 |

During 2013 and 2014, we found the T-wands to be overly sensitive leading to false positive detections and additional work related to processing snouts to extract CWTs. On October 2, 2014 we setup two series R9500 detectors to expedite scanning for CWTs (Figure 1). The detectors were checked for proper operation each day prior to scanning any fish. Informal quality control checks occurred daily during the first two weeks of operation in order to identify the detection efficiency of each detector. These checks involved running 100 fish through each machine and then re-scanning the fish with the T-wands. A total of 2,000 fish were passed through the R9500 units of which 422 were identified to possess a CWT. Of these fish, 419 signaled positive for a CWT during the initial scanning. The three fish possessing a CWT that were not identified by the R9500 during the initial scanning were correctly detected when re-ran though the detectors. The missed fish were likely the result of passing fish through the detectors too rapidly which can interfere with the operation of the flip gates.
R9500 detectors were used to scan the vast majority of fish surplused at PRH during 2015 and 2016. During both years, the first group of fish handled each day was used to test the CWT detection of each R9500 detector. The test fish that a CWT was not detected were re-scanned with a T-wand to assess the performance of the R9500 detectors. In 2015, 4,596 fish were sampled in this fashion and only $2(0.04 \%)$ were found to possess a CWT that was not initially detected by the R9500 and diverted to the tote containing CWT fish. In 2016, only 3 of 5,943 ( $0.05 \%$ ) were found to possess a CWT that was not detected by the R9500 and diverted to the tote containing CWT fish. Similar to observations in 2014, when these fish were rescanned by the R9500 they were correctly diverted to the tote containing CWT fish.

The methods describe here do not provide a definitive estimate of undetected CWTs for fish sampled at PRH. We make an assumption, that if the CWT detection wands and R9500 units do not detect a CWT in a given fish, then it did not possess a tag. Based on this assumption, the CWT detection efficiency at PRH is likely greater than $99 \%$. Therefore, the magnitude of the CWT recovery bias expressed in Table 1 is not likely due to poor CWT detection efficiency.


Figure 1. Series R9500 Coded-wire tag detectors used at Priest Rapids Hatchery, 2014

## Appendix B

## Recovery of coded-wire tags collected from adult Chinook salmon broodstock spawned at Priest Rapids Hatchery during

 Return Year 2016.| Code | Tag \# | BY | Run | Age | Stock | Release Location | CWT Release |  |  | Expansion |  | Escapement |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Date | $\begin{gathered} \hline \text { AD } \\ \text { CWT } \end{gathered}$ | $\begin{aligned} & \hline \text { CWT } \\ & \text { Only } \end{aligned}$ | $\begin{gathered} \text { All } \\ \text { CWT } \end{gathered}$ | $\begin{gathered} \text { AD } \\ \text { CWT } \end{gathered}$ | \# | \% |
| 635971 | 22 | 2010 | Fall | 6 | Priest Rapids | Priest Rapids | 2011 | 0 | 204,590 | 4.0 | 11.3 | N/A | N/A |
| 635972 | 14 | 2010 | Fall | 6 | Priest Rapids | Priest Rapids | 2011 | 0 | 199,600 | 4.0 | 11.3 | N/A | N/A |
| 636507 | 43 | 2012 | Fall | 4 | Priest Rapids | Priest Rapids | 2013 | 603,930 | 0 | 5.662 | 11.297 | N/A | N/A |
| 636508 | 176 | 2012 | Fall | 4 | Priest Rapids | Priest Rapids | 2013 | 0 | 601,009 | 5.662 | 11.297 | N/A | N/A |
| 636681 | 12 | 2013 | Fall | 3 | Priest Rapids | Priest Rapids | 2014 | 600,883 | 2,914 | 6.017 | 12.093 | N/A | N/A |
| 636682 | 27 | 2013 | Fall | 3 | Priest Rapids | Priest Rapids | 2014 | 0 | 603,819 | 6.017 | 12.093 | N/A | N/A |
| 090681 | 1 | 2012 | Fall | 4 | Priest Rapids | Ringold | 2013 | 214,873 | 5,943 | 14.706 | 15.113 | N/A | N/A |
| 090863 | 1 | 2013 | Fall | 3 | Priest Rapids | Ringold | 2011 | 222,740 | 2,784 | 15.096 | 15.287 | N/A | N/A |
| 090704 | 1 | 2012 | Fall | 4 | Umatilla R | Umatilla R | 2013 | 140,915 | 120 | 1.986 | 1.987 | N/A | N/A |
| Total | 297 | 4,325 Volunteer Trap Broodstock Spawned |  |  |  |  |  |  |  |  |  | N/A | N/A |

## Appendix C

Recovery of coded-wire tags collected from adult Chinook salmon surplus or mortalities from Priest Rapids Hatchery during Return Year 2016.

| Code | $\begin{gathered} \text { Tag } \\ \# \end{gathered}$ | BY | Race | Age | Stock | Release Location | CWT Release |  |  | Expansion |  | Escapement |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Date | $\begin{gathered} \text { AD } \\ \text { CWT } \\ \hline \end{gathered}$ | $\begin{aligned} & \text { CWT } \\ & \text { Only } \\ & \hline \end{aligned}$ | $\begin{gathered} \text { All } \\ \text { CWT } \\ \hline \end{gathered}$ | $\begin{aligned} & \hline \text { ADC } \\ & \text { WT } \\ & \hline \end{aligned}$ | \# | \% |
| 220238 | 1 | 2013 | Fall | 3 | Lyons Ferry | Clearwater@lapwai | 2014 | 100,911 | 0 | 1.370 | 4.118 | 1 | 0.0\% |
| 220240 | 5 | 2013 | Fall | 3 | Lyons Ferry | Clearwater@lapwai | 2015 | 0 | 202,383 | 1.370 | 4.118 | 7 | 0.0\% |
| 635274 | 1 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 0 | 99,800 | 4.0 | 11.3 | 4 | 0.0\% |
| 635699 | 7 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 203,682 | 409 | 3.972 | 11.282 | 28 | 0.1\% |
| 635764 | 1 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 199,698 | 401 | 4.0 | 11.3 | 4 | 0.0\% |
| 635766 | 2 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 0 | 204,091 | 4.0 | 11.3 | 8 | 0.0\% |
| 635970 | 1 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 199,200 | 400 | 4.0 | 11.3 | 4 | 0.0\% |
| 635971 | 6 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 0 | 204,590 | 4.0 | 11.3 | 24 | 0.1\% |
| 635972 | 1 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 0 | 199,600 | 4.0 | 11.3 | 4 | 0.0\% |
| 635973 | 1 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 0 | 200,099 | 4.0 | 11.3 | 4 | 0.0\% |
| 635974 | 3 | 2010 | Fall | 6 | Priest Rapids | CR@Priest Rapids | 2011 | 0 | 199,600 | 3.972 | 11.282 | 12 | 0.0\% |
| 636371 | 187 | 2011 | Fall | 5 | Priest Rapids | CR@Priest Rapids | 2012 | 0 | 598,031 | 5.912 | 11.848 | 1106 | 3.8\% |
| 636372 | 208 | 2011 | Fall | 5 | Priest Rapids | CR@Priest Rapids | 2012 | 595,608 | 0 | 5.912 | 11.848 | 1230 | 4.3\% |
| 636507 | 1060 | 2012 | Fall | 4 | Priest Rapids | CR@Priest Rapids | 2013 | 603,930 | 0 | 5.662 | 11.297 | 6002 | 20.9\% |
| 636508 | 1146 | 2012 | Fall | 4 | Priest Rapids | CR@Priest Rapids | 2013 | 0 | 601,009 | 5.662 | 11.297 | 6489 | 22.5\% |
| 636681 | 836 | 2013 | Fall | 3 | Priest Rapids | CR@Priest Rapids | 2014 | 600,883 | 2,914 | 6.017 | 12.093 | 5031 | 17.5\% |
| 636682 | 969 | 2013 | Fall | 3 | Priest Rapids | CR@Priest Rapids | 2014 | 0 | 603,819 | 6.017 | 12.093 | 5831 | 20.3\% |
| 636836 | 40 | 2014 | Fall | 2 | Priest Rapids | CR@Priest Rapids | 2015 | 604,425 | 425 | 5.826 | 11.660 | 233 | 0.8\% |
| 636837 | 38 | 2014 | Fall | 2 | Priest Rapids | CR@Priest Rapids | 2015 | 0 | 604,861 | 5.826 | 11.660 | 221 | 0.8\% |
| 090488 | 1 | 2010 | Fall | 6 | Priest Rapids | CR@Ringold | 2011 | 221,389 | 1,527 | 15.598 | 15.705 | 16 | 0.1\% |
| 090570 | 2 | 2011 | Fall | 5 | Priest Rapids | CR@Ringold | 2012 | 194,871 | 0 | 17.083 | 17.083 | 34 | 0.1\% |
| 090681 | 17 | 2012 | Fall | 4 | Priest Rapids | CR@Ringold | 2013 | 214,873 | 5,943 | 14.706 | 15.113 | 250 | 0.9\% |
| 90863 | 13 | 2013 | Fall | 3 | Priest Rapids | CR@Ringold | 2011 | 222,740 | 2,784 | 15.096 | 15.287 | 196 | 0.7\% |
| 635773 | 1 | 2011 | Summer | 5 | Wells | CR@Wells | 2013 | 289,476 | 522 | 1.0 | 1.0 | 1 | 0.0\% |
| 636505 | 2 | 2012 | Summer | 4 | Wells | CR@Wells | 2014 | 434,716 | 1,767 | 1.000 | 1.000 | 2 | 0.0\% |
| 636739 | 4 | 2013 | Fall | 3 | Lyons Ferry | Grande Ronde R | 2014 | 202,273 | 0 | 1.997 | 1.997 | 8 | 0.0\% |
| 610439 | 1 | 2011 | Fall | 5 | Hanford URB Wild | Hanford Reach | 2012 | 27,173 | 0 |  |  | 0 | 0.0\% |
| 610448 | 1 | 2012 | Fall | 4 | Hanford URB Wild | Hanford Reach | 2013 | 46,496 | 0 |  |  | 0 | 0.0\% |
| 610449 | 1 | 2012 | Fall | 4 | Hanford URB Wild | Hanford Reach | 2013 | 26,771 | 0 |  |  | 0 | 0.0\% |
| 610451 | 1 | 2012 | Fall | 4 | Hanford URB Wild | Hanford Reach | 2013 | 22,763 | 0 |  |  | 0 | 0.0\% |


| Code | $\begin{gathered} \text { Tag } \\ \# \\ \hline \end{gathered}$ | BY | Race | Age | Stock | Release Location | CWT Release |  |  | Expansion |  | Escapement |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Date | $\begin{gathered} \text { AD } \\ \text { CWT } \end{gathered}$ | $\begin{aligned} & \hline \text { CWT } \\ & \text { Only } \end{aligned}$ | $\begin{gathered} \text { All } \\ \text { CWT } \end{gathered}$ | $\begin{aligned} & \hline \text { ADC } \\ & \text { WT } \\ & \hline \end{aligned}$ | \# | \% |
| 610454 | 2 | 2013 | Fall | 3 | Hanford URB Wild | Hanford Reach | 2014 | 49,354 | 0 |  |  | 0 | 0.0\% |
| 610456 | 4 | 2013 | Fall | 3 | Hanford URB Wild | Hanford Reach | 2014 | 29,452 | 0 |  |  | 0 | 0.0\% |
| 610464 | 1 | 2014 | Fall | 2 | Hanford URB Wild | Hanford Reach | 2015 | 11,895 | 0 |  |  | 0 | 0.0\% |
| 220234 | 1 | 2013 | Fall | 3 | Lyons Ferry | Lukes Gulch | 2014 | 100,870 | 0 | 1.250 | 2.531 | 1 | 0.0\% |
| 636583 | 1 | 2012 | Fall | 4 | Snake River | Lyons Ferry | 2014 | 246,702 | 2,685 | 1.007 | 2.040 | 1 | 0.0\% |
| 220233 | 3 | 2013 | Fall | 3 | Lyons Ferry | Magrudor Corridor | 2014 | 102,430 | 0 | 1.253 | 2.469 | 4 | 0.0\% |
| 636173 | 1 | 2011 | Summer | 5 | Methow | Similkameen R | 2013 |  | 209,118 | 0.000 | 1.004 | 1 | 0.0\% |
| 635680 | 1 | 2011 | Summer | 5 | Methow | Similkameen R | 2013 | 206,700 | 1,553 | 1.0 | 1.0 | 1 | 0.0\% |
| 220337 | 1 | 2012 | Fall | 4 | Lyons Ferry | Snake R | 2014 | 88,140 | 0 | 1.003 | 1.003 | 1 | 0.0\% |
| 220342 | 1 | 2013 | Fall | 3 | Lyons Ferry | Snake R | 2014 | 98,628 | 0 | 2.727 | 2.727 | 3 | 0.0\% |
| 636737 | 2 | 2013 | Fall | 3 | Lyons Ferry | Snake R | 2014 | 202,329 | 1,071 | 1.130 | 1.136 | 2 | 0.0\% |
| 220336 | 1 | 2012 | Fall | 4 | Snake R | Snake R @ Capt John | 2014 | 0 | 86,280 | 1.011 | 2.172 | 1 | 0.0\% |
| 220343 | 3 | 2013 | Fall | 3 | Lyons Ferry | Snake R @ Capt John | 2014 | 0 | 99,142 | 2.554 | 5.056 | 8 | 0.0\% |
| 90818 | 3 | 2013 | Fall | 3 | Snake R | Snake R@Hells Canyon | 2014 | 191,092 | 525 | 4.757 | 4.771 | 14 | 0.0\% |
| 90888 | 2 | 2014 | Fall | 2 | Snake R | Snake R@Hells Canyon | 2015 | 244,342 | 268 | 4.279 | 4.284 | 9 | 0.0\% |
| 220347 | 1 | 2013 | Fall | 3 | Lyons Ferry | Snake@Pittsburg | 2014 | 100,063 | 0 | 2.009 | 4.006 | 2 | 0.0\% |
| 220358 | 1 | 2014 | Fall | 2 | Lyons Ferry | Snake@Pittsburg | 2015 | 0 | 96,274 | 2.058 | 4.098 | 2 | 0.0\% |
| 090585 | 1 | 2011 | Fall | 5 | Umatilla Hatchery | Umatilla R | 2012 | 154,611 | 0 | 1.694 | 1.694 | 2 | 0.0\% |
| 090655 | 2 | 2011 | Fall | 5 | Umatilla Hatchery | Umatilla R | 2013 | 50,112 | 613 | 1.040 | 2.046 | 2 | 0.0\% |
| 090657 | 1 | 2011 | Fall | 5 | Umatilla Hatchery | Umatilla R | 2013 | 88,668 | 359 | 1.005 | 2.025 | 1 | 0.0\% |
| 090658 | 1 | 2011 | Fall | 5 | Umatilla Hatchery | Umatilla R | 2013 | 0 | 223,550 | 1.005 | 2.025 | 1 | 0.0\% |
| 090682 | 3 | 2012 | Fall | 4 | Umatilla R | Umatilla R | 2014 | 0 | 229,652 | 1.040 | 2.046 | 3 | 0.0\% |
| 090683 | 2 | 2012 | Fall | 4 | Umatilla R | Umatilla R | 2014 | 102,499 | 1,784 | 1.040 | 2.046 | 2 | 0.0\% |
| 090684 | 3 | 2012 | Fall | 4 | Umatilla R | Umatilla R | 2014 | 49,266 | 200 | 1.040 | 2.046 | 3 | 0.0\% |
| 090704 | 11 | 2012 | Fall | 4 | Umatilla R | Umatilla R | 2013 | 140,915 | 120 | 1.986 | 1.987 | 22 | 0.1\% |
| 090705 | 10 | 2012 | Fall | 4 | Umatilla R | Umatilla R | 2013 | 166,640 | 0 | 1.986 | 1.987 | 20 | 0.1\% |
| 090816 | 19 | 2013 | Fall | 3 | Umatilla River | Umatilla R | 2014 | 168,393 | 824 | 1.875 | 1.880 | 36 | 0.1\% |
| 090817 | 13 | 2013 | Fall | 3 | Umatilla River | Umatilla R | 2014 | 163,114 | 0 | 1.875 | 1.880 | 24 | 0.1\% |
| 90870 | 3 | 2013 | Fall | 3 | Umatilla R | Umatilla R | 2015 | 210,611 | 0 | 1.027 | 1.030 | 3 | 0.0\% |
| 90871 | 2 | 2013 | Fall | 3 | Umatilla R | Umatilla R | 2015 | 114,305 | 462 | 1.027 | 1.030 | 2 | 0.0\% |
| 90917 | 1 | 2014 | Fall | 2 | Umatilla R | Umatilla R | 2015 | 161,668 | 0 | 4.040 | 4.040 | 4 | 0.0\% |
| 90944 | 4 | 2014 | Fall | 2 | Umatilla R | Umatilla R | 2016 | 483,071 | 975 | 1.007 | 1.009 | 4 | 0.0\% |
| 90945 | 4 | 2014 | Fall | 2 | Umatilla R | Umatilla R | 2016 | 227,783 | 1,256 | 1.007 | 1.013 | 4 | 0.0\% |


| Code | $\begin{gathered} \text { Tag } \\ \# \end{gathered}$ | BY | Race | Age | Stock | Release Location | CWT Release |  |  | Exp |  | Escapement |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  | Date | $\begin{gathered} \hline \text { AD } \\ \text { CWT } \end{gathered}$ | $\begin{aligned} & \hline \text { CWT } \\ & \text { Only } \end{aligned}$ | $\begin{gathered} \hline \text { All } \\ \text { CWT } \end{gathered}$ | $\begin{gathered} \hline \text { ADC } \\ \text { WT } \end{gathered}$ | \# | \% |
| 90946 | 3 | 2014 | Fall | 2 | Umatilla R | Umatilla R | 2016 | 105,561 | 531 | 1.002 | 1.007 | 3 | 0.0\% |
| 90909 | 45 | NA | Fall | NA | ODFW | ODFW | NA | NA | NA | NA | NA | NA | 0.0\% |
| Total | 4714 |  | 28,786 |  | red at PRH |  |  |  |  |  |  | 26,935 | 93.6\% |

## Appendix D

Juvenile fish health inspections for Priest Rapids Hatchery fall Chinook salmon, Brood Years 1998-2008. The description in the Condition column indicates the presence of a certain condition within at least one of the fish examined.

| Hatchery | Date | Species | Stock | Brood Year | Condition |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Priest Rapids | 23-Feb-99 | CHF | Priest Rapids | 1998 | Healthy |
| Priest Rapids | 22-Mar-99 | CHF | Priest Rapids | 1998 | Healthy |
| Priest Rapids | 23-Apr-99 | CHF | Priest Rapids | 1998 | Healthy |
| Priest Rapids | 25-May-99 | CHF | Priest Rapids | 1998 | Dropout Syndrome \& Bacterial |
| Priest Rapids | 08-Jun-99 | CHF | Priest Rapids | 1998 | Bacterial Kidney Disease |
| Priest Rapids | 06-Mar-00 | CHF | Priest Rapids | 1999 | Healthy |
| Priest Rapids | 14-Apr-00 | CHF | Priest Rapids | 1999 | Healthy |
| Priest Rapids | 16-May-00 | CHF | Priest Rapids | 1999 | Healthy |
| Priest Rapids | 12-Jun-00 | CHF | Priest Rapids | 1999 | Healthy |
| Priest Rapids | 23-Feb-01 | CHF | Priest Rapids | 2000 | Healthy |
| Priest Rapids | 05-Apr-01 | CHF | Priest Rapids | 2000 | Healthy |
| Priest Rapids | 07-May-01 | CHF | Priest Rapids | 2000 | Healthy |
| Priest Rapids | 06-Jun-01 | CHF | Priest Rapids | 2000 | Healthy |
| Priest Rapids | 13-Feb-02 | CHF | Priest Rapids | 2001 | Healthy |
| Priest Rapids | 01-Mar-02 | CHF | Priest Rapids | 2001 | Coagulated Yolk Syndrome |
| Priest Rapids | 22-Apr-02 | CHF | Priest Rapids | 2001 | Healthy |
| Priest Rapids | 10-Jun-02 | CHF | Priest Rapids | 2001 | Healthy |
| Priest Rapids | 07-Mar-03 | CHF | Priest Rapids | 2002 | Healthy |
| Priest Rapids | 15-Apr-03 | CHF | Priest Rapids | 2002 | Healthy |
| Priest Rapids | 02-Jun-03 | CHF | Priest Rapids | 2002 | Healthy |
| Priest Rapids | 01-Apr-04 | CHF | Priest Rapids | 2003 | Healthy |
| Priest Rapids | 06-May-04 | CHF | Priest Rapids | 2003 | Healthy |
| Priest Rapids | 07-Jun-04 | CHF | Priest Rapids | 2003 | Healthy |
| Priest Rapids | 11-Mar-05 | CHF | Priest Rapids | 2004 | Healthy |
| Priest Rapids | 14-Apr-05 | CHF | Priest Rapids | 2004 | Healthy |
| Priest Rapids | 1-Jun-05 | CHF | Priest Rapids | 2004 | Healthy |
| Priest Rapids | 6-Mar-06 | CHF | Priest Rapids | 2005 | Healthy |
| Priest Rapids | 25-Apr-06 | CHF | Priest Rapids | 2005 | Healthy |
| Priest Rapids | 13-Jun-06 | CHF | Priest Rapids | 2005 | Healthy |
| Priest Rapids | 9-Mar-07 | CHF | Priest Rapids | 2006 | Healthy |
| Priest Rapids | 19-Apr-07 | CHF | Priest Rapids | 2006 | Healthy |
| Priest Rapids | 1-Jun-07 | CHF | Priest Rapids | 2006 | Healthy |
| Priest Rapids | 12-Feb-08 | CHF | Priest Rapids | 2007 | Coagulated Yolk Syndrome |
| Priest Rapids | 23-Apr-08 | CHF | Priest Rapids | 2007 | Healthy |
| Priest Rapids | 4-Jun-08 | CHF | Priest Rapids | 2007 | Healthy |
| Priest Rapids | 12-Feb-09 | CHF | Priest Rapids | 2008 | Coagulated Yolk Syndrome |
| Priest Rapids | 22-Apr-09 | CHF | Priest Rapids | 2008 | Healthy |
| Priest Rapids | 8-Jun-09 | CHF | Priest Rapids | 2008 | Healthy |

## Appendix E

## Summary of aerial fall Chinook salmon redd counts in the Hanford Reach, Columbia River, Washington.

Number and percent of fall Chinook salmon redds counted in different reaches of the Columbia River, 2001-
2016. Data for years 2001-2010 was collected by staff with Pacific Northwest National Laboratory. Data for
years 2011-2016 was collected by staff with Environmental Assessment Services, LLC.

| Location | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Islands 11-21 | 297 | 509 | 554 | 337 | 708 | 36 | 302 | 371 | 176 | 562 |
| Islands 8-10 | 480 | 865 | 1,133 | 867 | 1,067 | 435 | 338 | 416 | 722 | 870 |
| Near Island 7 | 350 | 280 | 455 | 415 | 500 | 873 | 311 | 360 | 380 | 457 |
| Island 6 (lower) | 750 | 940 | 1,241 | 1,084 | 1,229 | 289 | 615 | 753 | 878 | 1,135 |
| Island 4, 5,6 | 1,130 | 1,165 | 1,242 | 1,655 | 1,130 | 934 | 655 | 960 | 796 | 1,562 |
| Near Island 3 | 460 | 249 | 475 | 325 | 345 | 1,305 | 152 | 230 | 285 | 244 |
| Near Island 2 | 780 | 955 | 850 | 960 | 895 | 523 | 455 | 555 | 459 | 657 |
| Near Island 1 | 35 | 235 | 270 | 330 | 255 | 253 | 47 | 148 | 160 | 324 |
| Coyote Rapids | 16 | 63 | 354 | 180 | 304 | 150 | 10 | 29 | 34 | 49 |
| China Bar | 20 | 25 | 85 | 75 | 28 | 52 | 3 | 35 | 1,090 | 299 |
| Vernita Bar | 1,930 | 2,755 | 2,806 | 2,240 | 1,430 | 1,658 | 1,135 | 1,731 | 16 | 2,658 |
| Total | 6,248 | 8,041 | 9,465 | 8,468 | 7,891 | 6,508 | 4,023 | 5,588 | 4,996 | 8,817 |
| Location | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| Islands 11-21 | 5\% | 6\% | 6\% | 4\% | 9\% | 1\% | 8\% | 7\% | 4\% | 6\% |
| Islands 8-10 | 8\% | 11\% | 12\% | 10\% | 14\% | 7\% | 8\% | 7\% | 14\% | 10\% |
| Near Island 7 | 6\% | 3\% | 5\% | 5\% | 6\% | 13\% | 8\% | 6\% | 8\% | 5\% |
| Island 6 (lower) | 12\% | 12\% | 13\% | 13\% | 16\% | 4\% | 15\% | 13\% | 18\% | 13\% |
| Island 4, 5, 6 | 18\% | 14\% | 13\% | 20\% | 14\% | 14\% | 16\% | 17\% | 16\% | 18\% |
| Near Island 3 | 7\% | 3\% | 5\% | 4\% | 4\% | 20\% | 4\% | 4\% | 6\% | 3\% |
| Near Island 2 | 12\% | 12\% | 9\% | 11\% | 11\% | 8\% | 11\% | 10\% | 9\% | 7\% |
| Near Island 1 | 1\% | 3\% | 3\% | 4\% | 3\% | 4\% | 1\% | 3\% | 3\% | 4\% |
| Coyote Rapids | $>1 \%$ | 1\% | 4\% | 2\% | 4\% | 2\% | $>1 \%$ | 1\% | 1\% | 1\% |
| China Bar | $>1 \%$ | $>1 \%$ | 1\% | 1\% | $>1 \%$ | 1\% | $>1 \%$ | 1\% | 22\% | 3\% |
| Vernita Bar | 31\% | 34\% | 30\% | 26\% | 18\% | 25\% | 28\% | $31 \%$ | $>1 \%$ | 30\% |
| Location | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |  |  |  | (07-16) Mean |
| Islands 11-21 | 676 | 533 | 798 | 906 | 1,193 | 861 |  |  |  | 638 |
| Islands 8-10 | 814 | 807 | 2,200 | 1,565 | 3,145 | 1,735 |  |  |  | 1,261 |
| Near Island 7 | 670 | 700 | 655 | 1,100 | 800 | 670 |  |  |  | 610 |
| Island 6 (lower) | 1,181 | 1,375 | 3,340 | 2,530 | 2,315 | 1,807 |  |  |  | 1,593 |
| Island 4, 5,6 | 1,524 | 1,195 | 2,650 | 2,080 | 2,540 | 2,270 |  |  |  | 1,623 |
| Near Island 3 | 525 | 475 | 1,000 | 1,000 | 1,100 | 600 |  |  |  | 561 |
| Near Island 2 | 653 | 528 | 1,700 | 2,050 | 1,900 | 1,140 |  |  |  | 1,010 |
| Near Island 1 | 295 | 340 | 900 | 500 | 1,000 | 340 |  |  |  | 405 |
| Coyote Rapids | 44 | 29 | 520 | 500 | 765 | 255 |  |  |  | 224 |
| China Bar | 67 | 68 | 100 | 60 | 1,730 | 80 |  |  |  | 353 |
| Vernita Bar | 2,466 | 2,318 | 3,535 | 3,650 | 4,190 | 3,510 |  |  |  | 2,521 |
| Total | 8,915 | 8,368 | 17,398 | 15,951 | 20,678 | 13,268 |  |  |  | 10,799 |
| Location | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |  |  |  | (07-16) Mean |
| Islands 11-21 | 8\% | 6\% | 5\% | 6\% | 6\% | 6\% |  |  |  | 6\% |
| Islands 8-10 | 9\% | 10\% | 13\% | 10\% | 15\% | 13\% |  |  |  | 12\% |
| Near Island 7 | 8\% | 8\% | 4\% | 7\% | 4\% | 5\% |  |  |  | 6\% |
| Island 6 (lower) | 13\% | 16\% | 19\% | 16\% | 11\% | 14\% |  |  |  | 15\% |
| Island 4, 5, 6 | 17\% | 14\% | 15\% | 13\% | 12\% | 17\% |  |  |  | 15\% |
| Near Island 3 | 6\% | 6\% | 6\% | 6\% | 5\% | 5\% |  |  |  | 5\% |
| Near Island 2 | 7\% | 6\% | 10\% | 13\% | 9\% | 9\% |  |  |  | 9\% |
| Near Island 1 | 3\% | 4\% | 5\% | 3\% | 5\% | 3\% |  |  |  | 4\% |
| Coyote Rapids | $>1 \%$ | $>1 \%$ | 3\% | 3\% | 4\% | 2\% |  |  |  | 2\% |
| China Bar | 1\% | 1\% | 1\% | 0\% | 8\% | 1\% |  |  |  | 3\% |
| Vernita Bar | 28\% | 28\% | 20\% | 23\% | 20\% | 26\% |  |  |  | 23\% |

## Appendix F

Historical numbers of Chinook salmon carcasses recovered during the annual Hanford Reach fall Chinook salmon carcass survey, Return Years 1991-2016.

| Return Year | Total Recoveries | Total Escapement | Proportion of Escapement <br> Recovered |
| :---: | :---: | :---: | :---: |
| 1991 | 2,519 | 52,196 | 0.048 |
| 1992 | 2,221 | 41,952 | 0.053 |
| 1993 | 3,340 | 37,347 | 0.089 |
| 1994 | 5,739 | 63,103 | 0.091 |
| 1995 | 3,914 | 55,208 | 0.071 |
| 1996 | 4,529 | 43,249 | 0.105 |
| 1997 | 5,053 | 43,493 | 0.116 |
| 1998 | 4,456 | 35,393 | 0.126 |
| 1999 | 4,412 | 29,812 | 0.148 |
| 2000 | 10,556 | 48,020 | 0.220 |
| 2001 | 6,072 | 59,848 | 0.101 |
| 2002 | 8,402 | 84,509 | 0.099 |
| 2003 | 13,573 | 100,840 | 0.135 |
| 2004 | 11,030 | 87,696 | 0.126 |
| 2005 | 8,491 | 71,967 | 0.118 |
| 2006 | 5,972 | 51,701 | 0.116 |
| 2007 | 3,115 | 22,272 | 0.140 |
| 2008 | 5,455 | 29,058 | 0.188 |
| 2009 | 5,318 | 36,720 | 0.145 |
| 2010 | 9,779 | 87,016 | 0.112 |
| 2011 | 8,391 | 75,256 | 0.111 |
| 2012 | 6,814 | 57,710 | 0.118 |
| 2013 | 13,071 | 174,651 | 0.075 |
| 2014 | 16,756 | 183,749 | 0.091 |
| 2015 | 17,738 | 266,346 | 0.086 |
| 2016 | 8,886 | 116,421 | 0.076 |
| $\mathbf{M e a n}$ |  | $\mathbf{8 5 , 2 1 0}$ | $\mathbf{0 . 1 2 3}$ |
|  |  |  |  |
|  |  |  |  |

## Appendix G

## Estimated escapements for fall Chinook spawning in Hanford Reach and Priest Rapids Dam pool, Return Year 2016.

2016 Hanford Reach Fall Chinook Escapement Estimate

| Count Source |  | 2016 |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Adult | Jack | Total |
|  | McNary ${ }^{1}$ | 239,791 | 24,870 | 264,661 |
|  | Wanapum ${ }^{2}$ | 24,088 | 2,015 | 26,103 |
|  | Priest Rapids ${ }^{3}$ | 41,013 | 3,191 | 44,204 |
|  | Fallback Adjustment ${ }^{4}$ | 5,549 | 432 | 5,981 |
|  | Ice Harbor ${ }^{5}$ | 36,713 | 13,066 | 49,779 |
|  | Prosser ${ }^{6}$ | 5,214 | 464 | 5,678 |
|  | Priest Rapids Hatchery | 27,904 | 882 | 28,786 |
|  | Priest Rapids Hatchery Channel | 10 | 0 | 10 |
|  | Angler Broodstock Collection | 284 | 0 | 284 |
|  | Ringold Springs Hatchery | 5,314 | 65 | 5,379 |
|  | Hanford Sport Harvest | 16,859 | 1,068 | 17,927 |
|  | Yakima River Sport Harvest | 922 | 31 | 953 |
|  | Wanapum Tribal Fishery | 35 | 1 | 36 |
|  | Yakima River (Lower) ${ }^{5}$ | 1,087 | 97 | 1,184 |
|  | Hanford Reach + Priest Pool | 121,361 | 7,181 | 128,542 |
|  | Priest Pool Return | 11,376 | 744 | 12,120 |
|  | Hanford Reach Escapement | 109,984 | 6,437 | 116,421 |

${ }^{1}$ McNaryDam fish counts: August 9 - October 31
${ }^{2}$ Wanapum Dam fish counts, August 14 through November 5
${ }^{3}$ Priest Rapids Dam fish counts, August 18 through November 5. Grant PUD continued counts through Nov 15 but McNary counts ended on Oct 31. Allowed 5 days to account for difference in passage timing
${ }^{4}$ Fallback/Reascension Adjustment estimate ( $13.5 \%$ ) based on 154 run of the river PIT tagged fish from the BOAFF and the lower Columbia River test fishery observed at Priest Rapids Dam and Priest Rapids Hatchery PIT tag arrays
${ }^{5}$ Ice Harbor counts ended on Oct 31
${ }^{6}$ Prosser counts, August 16 through November 5
2016 Priest Rapids Pool Escapement

| Count Source | 2016 |  |  |
| :---: | :---: | :---: | :---: |
|  | Adult | Jack | Total |
| Wanapum Adult Passage | 24,764 | 2,096 | 26,860 |
| Wanapum Dam Fallback Adjustment | Unknown | Unknown | Unknown |
| Priest Rapids Fallback Adjustment ${ }^{2}$ | 5,549 | 432 | 5,981 |
| Wanapum Tribal Fishery Above PRD | 35 |  | 35 |
| OLAFT | 444 | 0 | 444 |
| Priest Rapids Pool Sport Fishery | 250 | 61 | 311 |
| Total | 47,587 | 0 | 47,587 |
| Priest Rapids Adult Passage ${ }^{3}$ | 41,013 | 3,191 | 44,204 |
| Priest Rapids Dam Pool Escapement | 9,971 | 3,191 | 13,162 |

${ }^{2}$ Fallback/Reascension Adjustment estimate (13.5\%) based on 154 run of the river PIT tagged fish from the BO AFF and the lower Columbia River test fishery observed at Priest Rapids Dam and Priest Rapids Hatchery PIT tag arrays.
${ }^{3}$ Priest Rapids passage for fall Chinook based on counts from August 18 through November 15.

## Appendix H Carcass Drift Assessment

A common objective of hatchery monitoring and evaluation programs is to identify the spawning distribution of both hatchery and natural origin fish. Initially, we believed that the proportion of hatchery origin spawners ( pHOS ) could be calculated for each of the five reaches. However, previous carcass bias assessments within the Hanford Reach suggest a substantial amount of downstream carcass drift into lower reaches (Richards and Pearsons, 2013). Hence, it is uncertain that the carcass recovery locations directly represent spawner distributions in some locations.

In order to gain a better understanding of natural post-spawn carcass drift, we tried two different approaches for tagging carcasses. During 2014, we used a long pole to floy-tag 993 carcasses in place without moving them (Richards and Pearsons, 2015). Tagging occurred from October 26 through November 23. This prevented the collection of accurate size and gender data as many carcasses tagged were underwater. We anticipated that some carcasses would move downstream as river flows fluctuated. Recovery efforts occurred from November 1 through December 19. Recovery rates ranged from 31-37\% for donor Sections 1, 2, and 4 (Table 1).

Table H. 1 Numbers of operculum-tagged Chinook salmon carcasses released and recovered by donor section within the Hanford Reach, Return Year 2014

|  | Donor Section 1 | Donor Section 2 | Donor Section 3 | Donor Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Ta | 486 | 107 | 225 | 176 | 994 |
| Fish Recovered by Recipient Section | 143 |  |  |  | 146 |
|  | 1 | 32 |  |  | 34 |
|  | 3 | 1 | 35 |  | 39 |
|  | 4 | 0 | 4 | 60 | 68 |
|  | 1 | 1 | 0 | 4 | 6 |
| $\mathrm{P}^{\wedge}$ Recovered for each Donor Section | 0.319 | 0.327 | 0.173 | 0.364 | 0.295 |
| Proportion Recovered by Section | 0.942 |  |  |  |  |
|  | 0.006 | 0.943 |  |  |  |
|  | 0.019 | 0.029 | 0.897 |  |  |
|  | 0.026 | 0.000 | 0.103 | 0.938 |  |
|  | 0.006 | 0.029 | 0.000 | 0.063 |  |
| Proportion Recovered by Section into recipient Section | 1.000 |  |  |  |  |
|  | 0.007 | 0.993 |  |  |  |
|  | 0.020 | 0.030 | 0.949 |  |  |
|  | 0.024 | 0.000 | 0.096 | 0.880 |  |
|  | 0.066 | 0.293 | 0.000 | 0.641 |  |

Donor Section 3 had the lowest recovery rate at $17 \%$. We found that many tagged carcasses did not move from the tag sites; hence the results suggest that carcass drift was occurring at very low rates. We now believe that large portions of carcasses remain in their initial location of deposition.

During 2015 and 2016, we adjusted our approach in attempt to mimic post-spawn fish dying near redd locations and subsequently drifting downstream. Each year, we opercula-tagged roughly 1,000 intact carcasses, collected size and gender data and then redistributed them in the proximity of specific spawning areas within Sections 1-4 (Figure ). Tagging occurred primarily in November for both years. Depths at release were visually estimated to range from 1 to 7
meters. River flow $\mathrm{m} / \mathrm{s}$ at release was not measured. No fish were released in eddies or slack water. Released carcasses were generally observed sinking quickly to the bottom and then slowly drifting downstream. Recovery efforts occurred from during November and early December. During 2015, crews recovered 39 (3.9\%) tagged carcasses (Table 2). The recovery rate was notably lower for fish released in Section 4 compared to the other sections. Although the numbers recovered were low, results show that large proportion of tagged fish recovered were found downstream of their adjacent donor section. During 2016, crews recovered 47 ( $4.8 \%$ ) tagged carcasses (Table 3). The recovery rate was notably lower for fish released in Section 2 compared to the other sections. Similar to the results of 2015, large proportions of fish tagged and recovered during 2016 were found downstream of their adjacent donor section.


Figure 1 Opercula Tagged Male fall Chinook in the Hanford Reach, 2015 Carcass Drift assessment.

Table H. 2 Numbers of operculum-tagged Chinook salmon carcasses released and recovered by donor section within the Hanford Reach, Return Year 2015

|  | Donor Section 1 | Donor Section 2 | Donor Section 3 | Donor Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Ta | 231 | 62 | 343 | 362 | 998 |
| Fish Recovered by Recipient Section | 4 |  |  |  | 4 |
|  | 0 | 1 |  |  | 1 |
|  | 6 | 3 | 4 |  | 13 |
|  | 2 | 0 | 13 | 4 | 19 |
|  | 0 | 0 | 1 | 1 | 2 |
| $\mathrm{P}^{\wedge}$ Recovered for each Donor Section | 0.052 | 0.065 | 0.052 | 0.014 | 0.039 |
| Proportion Recovered by Section | 0.333 |  |  |  |  |
|  | 0.000 | 0.250 |  |  |  |
|  | 0.500 | 0.750 | 0.222 |  |  |
|  | 0.167 | 0.000 | 0.722 | 0.800 |  |
|  | 0.000 | 0.000 | 0.056 | 0.200 |  |
| Proportion Recovered by Section into Recipient Section | 1.000 |  |  |  |  |
|  | 0.000 | 1.000 |  |  |  |
|  | 0.340 | 0.509 | 0.151 |  |  |
|  | 0.099 | 0.000 | 0.428 | 0.474 |  |
|  | 0.000 | 0.000 | 0.217 | 0.783 |  |

Table H. 3 Numbers of operculum-tagged Chinook salmon carcasses released and recovered by donor section within the Hanford Reach, Return Year 2016

|  | Donor Section 1 | Donor Section 2 | Donor Section 3 | Donor Section 4 | Totals |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Fish Tagged by Donor Section | 263 | 138 | 332 | 254 | 987 |
| Fish Recovered by Recipient Section | 3 |  |  |  | 3 |
|  | 0 | 0 |  |  | 0 |
|  | 10 | 0 | 4 |  | 14 |
|  | 7 | 1 | 12 | 5 | 25 |
|  | 0 | 0 | 1 | 4 | 5 |
| $\mathrm{P}^{\wedge}$ Recovered for each Donor Section | 0.076 | 0.007 | 0.051 | 0.035 | 0.048 |
| 1 | 0.150 |  |  |  |  |
| Proportion $\quad 2$ | 0.000 | 0.000 |  |  |  |
| Recovered by $\qquad$ | 0.500 | 0.000 | 0.235 |  |  |
| Section | 0.350 | 1.000 | 0.706 | 0.556 |  |
| 5 | 0.000 | 0.000 | 0.059 | 0.444 |  |
| 1 | 1.000 |  |  |  |  |
| Proportion $2$ | 0.000 | 0.000 |  |  |  |
| Recovered by $\quad 3$ | 0.680 | 0.000 | 1.000 |  |  |
| recipient Section $\qquad$ | 0.134 | 0.442 | 0.560 | 1.000 |  |
| - 5 | 0.000 | 0.000 | 0.117 | 1.000 |  |

## Appendix I

## Carcass bias assessment results

Carcass surveys of Chinook Salmon are conducted each fall to characterize fish spawning in the Hanford Reach. However, it is possible that carcasses collected during surveys do not represent the spawning population. There could be carcass collection bias against smaller/younger fish or males in the stream surveys (Zhou 2002; Murdoch et al. 2010; Richards and Pearsons, 2013). If true, this bias may compromise estimates associated with age and gender compositions by origin as well as escapement estimates of hatchery and natural origin fish. We began a pilot project to evaluate potential size and sex recovery bias in 2011. This work has occurred annually with the exception of 2014 when measurements were not taken on the mark sample so a bias estimate could not be estimated.

The methods for collecting, sampling, and releasing tagged carcass associated with this evaluation have varied slightly between years. In general, roughly 1,000 carcasses were collected for demographic data and tagged with numbered plastic tags. Depending on the year, the tagged fish were either systematically released either near shore or mid river or over known active redd locations.

The release strategy for years 2011-2013 included releasing tagged carcass either near shore or mid-channel near the point of initial recovery. Carcasses released near shore had higher proportions of recaptures compared to fish released mid channel. It was not uncommon for carcasses released near shore to be recovered the following day in the same vicinity of their release. In 2015 and 2016, we released tagged carcasses over active redd locations to better match the natural disposition of post spawn carcasses. After release into the river, the carcasses generally sunk quickly and gradually moved downstream along the bottom in a similar manner to that of post-spawn fish.
The annual recovery rates of tagged carcasses decreased annually from a high of $17.2 \%$ in 2011 to a low of $3.8 \%$ in 2015 (Tables 1-6). The annual recovery rates may be influenced by the release method and by reduced chances of recovering tagged carcasses during large spawning escapements of fall Chinook salmon to the Hanford Reach.

In general, the level of carcass recovery bias was low and varied between years; suggesting that carcass samples collected may be reflective of the spawning population. The accumulation of additional data following the methods used in years 2015 and 2016 will benefit the analysis of possible size related carcass recovery bias.

Table I. 1 Summary of mark recapture of post-spawn fall Chinook Salmon in the Hanford Reach, 2011. Post orbital to hypural plate length (POHL) calculated from linear regression equation for fork length versus known POHL

|  |  |  | Release Locations |  |  |  |  |  | Total Released |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Bank |  |  | Mid Channel |  |  |  |  |
| Released \# |  |  | 500 |  |  | 493 |  |  | 993 |  |
| Recaptured \# |  |  | 110 |  |  | 61 |  |  | 171 |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.220 |  |  | 0.124 |  |  | 0.172 |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 26 | 0.026 | 66 | 0.066 | 172 | 0.173 | 137 | 0.138 | 401 | 0.404 |
| Female | 0 | 0.000 | 14 | 0.014 | 331 | 0.333 | 247 | 0.249 | 592 | 0.596 |
| Total | 26 | 0.026 | 80 | 0.081 | 503 | 0.507 | 384 | 0.387 | 993 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 3 | 0.018 | 10 | 0.058 | 34 | 0.199 | 26 | 0.152 | 73 | 0.427 |
| Female | 0 | 0.000 | 2 | 0.012 | 53 | 0.310 | 43 | 0.251 | 98 | 0.573 |
| Total | 3 | 0.018 | 12 | 0.070 | 87 | 0.509 | 69 | 0.404 | 171 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | 0.0 |  |  | . 008 |  |  |  |  |  | . 23 |
| Female | 0.0 |  |  | . 002 |  |  |  |  |  | 23 |
| Total | 0.0 |  |  | 010 |  |  |  |  |  | 00 |

Table I. 2 Summary of mark recapture of post-spawn fall Chinook Salmon in the Hanford Reach, 2012. Post orbital to hypural plate length (POHL) calculated from linear regression equation for fork length versus known POHL

|  |  |  | Release Locations |  |  |  |  |  | Total Released |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Bank |  |  | Mid Channel |  |  |  |  |
| Released \# |  |  | 489 |  |  | 500 |  |  | 989 |  |
| Recaptured \# |  |  | 103 |  |  | 34 |  |  | 137 |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.211 |  |  | 0.068 |  |  | 0.139 |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 49 | 0.050 | 172 | 0.174 | 157 | 0.159 | 142 | 0.144 | 520 | 0.526 |
| Female | 0 | 0.000 | 31 | 0.031 | 192 | 0.194 | 246 | 0.249 | 469 | 0.474 |
| Total | 49 | 0.050 | 203 | 0.205 | 349 | 0.353 | 388 | 0.392 | 989 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 10 | 0.073 | 25 | 0.182 | 20 | 0.146 | 19 | 0.139 | 74 | 0.540 |
| Female | 0 | 0.000 | 5 | 0.036 | 22 | 0.161 | 36 | 0.263 | 63 | 0.460 |
| Total | 10 | 0.073 | 30 | 0.219 | 42 | 0.307 | 55 | 0.401 | 137 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | -0.0 |  |  |  |  |  | 0.0 |  |  | . 014 |
| Female | 0.0 |  |  |  |  |  | -0.01 |  |  | . 14 |
| Total | -0.0 |  |  |  |  | 46 |  |  |  | 00 |

Table I.3. Summary of mark recapture of post-spawn fall Chinook Salmon in the Hanford Reach, 2013. Post orbital to hypural plate length (POHL) calculated from linear regression equation for fork length versus known POHL

|  |  |  | Release Locations |  |  |  |  |  | Total Released |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Bank |  |  | Mid Channel |  |  |  |  |
| Released \# |  |  | 552 |  |  | 521 |  |  | 1,073 |  |
| Recaptured \# |  |  | 69 |  |  | 45 |  |  | 114 |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.125 |  |  | 0.086 |  |  | 0.106 |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | $47-58 \mathrm{~cm}$ |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 206 | 0.192 | 332 | 0.309 | 183 | 0.170 | 60 | 0.056 | 781 | 0.727 |
| Female | 1 | 0.001 | 55 | 0.051 | 184 | 0.171 | 55 | 0.051 | 295 | 0.274 |
| Total | 206 | 0.192 | 387 | 0.360 | 367 | 0.341 | 115 | 0.107 | 1,075 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 16 | 0.140 | 42 | 0.368 | 19 | 0.167 | 8 | 0.070 | 85 | 0.746 |
| Female | 0 | 0.000 | 5 | 0.044 | 19 | 0.167 | 5 | 0.044 | 29 | 0.254 |
| Total | 16 | 0.140 | 47 | 0.412 | 38 | 0.333 | 13 | 0.114 | 114 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | 0.0 |  | -0.0 |  | 0.0 |  | -0.0 |  |  |  |
| Female | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |  |  |
| Total | 0.0 |  | -0.0 |  | 0.0 |  | -0.0 |  |  |  |

Table I. 4 Summary of mark recapture of post-spawn fall Chinook Salmon in the Hanford Reach, 2015. Post orbital to hypural plate length (POHL)

| Total Release in Mid-Channel Redd Locations, RY2015 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Released \# |  |  | 997 |  |  |  |  |  |  |  |
| Recaptured \# |  |  | 38 |  |  |  |  |  |  |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.038 |  |  |  |  |  |  |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | $47-58 \mathrm{~cm}$ |  | 59.69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 39 | 0.039 | 128 | 0.128 | 183 | 0.184 | 172 | 0.173 | 522 | 0.524 |
| Female | 1 | 0.001 | 37 | 0.037 | 287 | 0.288 | 151 | 0.151 | 476 | 0.477 |
| Total | 39 | 0.039 | 165 | 0.165 | 470 | 0.471 | 323 | 0.324 | 997 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 0 | 0.000 | 6 | 0.158 | 9 | 0.237 | 8 | 0.211 | 23 | 0.605 |
| Female | 0 | 0.000 | 1 | 0.026 | 7 | 0.184 | 7 | 0.184 | 15 | 0.395 |
| Total | 0 | 0.000 | 7 | 0.184 | 16 | 0.421 | 15 | 0.395 | 38 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male | 0.0 |  |  |  |  |  |  |  |  |  |
| Female | 0.0 |  |  |  |  |  |  |  |  |  |
| Total | 0.0 |  |  |  |  |  |  |  |  |  |

Table I. 5 Summary of mark recapture of post-spawn fall Chinook Salmon in the Hanford Reach, 2016. Post orbital to hypural plate length (POHL).

| Total Release in Mid-Channel Redd Locations, RY2016 |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Released \# |  |  | 987 |  |  |  |  |  |  |  |
| Recaptured \# |  |  | 46 |  |  |  |  |  |  |  |
| Recapture $\mathrm{P}^{\wedge}$ |  |  | 0.047 |  |  |  |  |  |  |  |
| Mark Release Fall Chinook Salmon |  |  |  |  |  |  |  |  |  |  |
| POHL | $<47 \mathrm{~cm}$ |  | 47-58 cm |  | 59-69 cm |  | $>69 \mathrm{~cm}$ |  | Total |  |
| Gender | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ | \# | $\mathbf{P}^{\wedge}$ |
| Male | 43 | 0.044 | 171 | 0.173 | 181 | 0.183 | 119 | 0.121 | 514 | 0.521 |
| Female | 0 | 0.000 | 35 | 0.035 | 334 | 0.338 | 104 | 0.105 | 473 | 0.479 |
| Total | 43 | 0.044 | 206 | 0.209 | 515 | 0.522 | 223 | 0.226 | 987 | 1.000 |
| Recaptures |  |  |  |  |  |  |  |  |  |  |
| Male | 4 | 0.087 | 7 | 0.152 | 11 | 0.239 | 4 | 0.087 | 26 | 0.565 |
| Female | 0 | 0.000 | 1 | 0.022 | 15 | 0.326 | 4 | 0.087 | 20 | 0.435 |
| Total | 4 | 0.087 | 8 | 0.174 | 26 | 0.565 | 8 | 0.174 | 46 | 1.000 |
| Bias |  |  |  |  |  |  |  |  |  |  |
| Male |  | . 43 | 0.0 |  | -0.0 |  |  | . 34 |  | . 044 |
| Female |  | 00 | 0.0 |  | 0.0 |  |  | . 18 |  | . 044 |
| Total |  | . 43 |  |  | -0.0 |  |  | . 52 |  | 000 |

Table I.6 Mark and recapture bias of post-spawn fall Chinoook Salmon in the Hanford Reach by size group (post orbital to hypural plate - POHL), Return Years 2011-13 and 2014-15. Bias = $P^{\wedge}$ Released - $P^{\wedge}$ Recovered.

| Return |  |  | Post Orbital to Hypural Plate Length Size Groups |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | \# Tagged | \# Recovered | $<\mathbf{4 7} \mathbf{~ c m}$ | $\mathbf{4 7 - 5 8} \mathbf{~ c m}$ | $\mathbf{5 9 - 6 9} \mathbf{c m}$ | $>\mathbf{6 9} \mathbf{c m}$ |
| $2011^{\text {a }}$ | 993 | 171 | 0.009 | 0.010 | -0.002 | -0.017 |
| $2012^{\mathrm{a}}$ | 989 | 137 | -0.023 | -0.014 | 0.046 | -0.009 |
| $2013^{\mathrm{a}}$ | 1073 | 114 | 0.051 | -0.052 | 0.008 | -0.007 |
| $2015^{\text {b }}$ | 997 | 38 | 0.039 | -0.019 | 0.05 | -0.071 |
| $2016^{\text {b }}$ | 987 | 46 | -0.043 | 0.035 | -0.043 | 0.052 |
| Mean | $\mathbf{1 0 0 8}$ | $\mathbf{1 0 1}$ | $\mathbf{0 . 0 0 7}$ | $\mathbf{- 0 . 0 0 8}$ | $\mathbf{0 . 0 1 2}$ | $\mathbf{- 0 . 0 1 0}$ |

[^5]
## Appendix J <br> Demographic comparisons for double index tag groups released from Priest Rapids Hatchery, Brood Years 2009-2011.

Double Index Tag (DIT) groups of fall Chinook salmon have been released annually from Priest Rapids Hatchery (PRH) starting with the progeny of the 2009 brood. Adipose clipped fish from these DIT groups have been recovered in various mark selective fisheries (MSF) occurring in marine, ocean, and freshwater zones designated by the Washington Department of Fish and Wildlife (WDFW). The Regional Mark Processing Center database was queried to identify mark selective fisheries occurring since 2010 that included recoveries of PRH DIT groups (Table 1). Detailed descriptions of these fisheries are available at websites maintained by the RMPC, Oregon Department of Fish and Game, and WDFW. The level of contribution to these fisheries, some of which are summer Chinook salmon fisheries, is beyond the scope of this document.

Survival estimates for DIT groups from release and recovery at PRH was calculated by dividing the total DIT recoveries at PRH for each brood year (ages $1-6$ ) by the corresponding number of juveniles marked for each DIT group. Comparisons between DIT groups within a brood year strongly suggest there is no difference in survival (Table 2). Similar comparisons for gender composition, age at maturity, as well as, size at age strongly suggest there is no difference between the DIT groups recovered at PRH for a given brood year (Tables 3, 4, and 5).

Table J.1. Regional Mark Processing Center location names of mark selective fisheries showing recoveries of Priest Rapids Hatchery origin codedwire tagged adipose clipped fish from brood years 2009 - 2011.

| Location Name |  |  |
| :--- | :--- | :--- |
| 1A (BUOY10 - BRIDGE) | COL R OR SPORT SEC 6 | COL R WA SPORT SEC 2 |
| 1B (BRIDGE - BEAVER) | COL R OR SPORT SEC 7 | COL R WA SPORT SEC 5 |
| ASTORIA SPORT 2 | COL R OR SPORT SEC 8 | COL R WA SPORT SEC 8 |
| BONNEVILLE POOL UPPER | COL R OR SPT SEC 10 | COL R WA SPORT SEC 9 |
| BROOKINGS SPORT 6 | COL R PRIEST-WANAPUM | COL R WN SPORT SEC 1 |
| COL R OR SPORT SEC 1 | COL R ROCK I-ROCKY R | COLUMBIA R AT DESCHUTES |
| COL R OR SPORT SEC 2 | COL R ROCKY R-WELLS | COOS BAY SPORT 5 |
| COL R OR SPORT SEC 3 | COL R WA SEC 4 | COWLITZ R 26.0002 |
| COL R OR SPORT SEC 4 | COL R WA SEC 6 | EDIZ HOOK |
| COL R OR SPORT SEC 5 | COL R WA SEC 7 | GARIBALDI SPORT 3 |
| WILLAPA HARBOR | WINCHESTER B SPORT 5 | JOHN DAY POOL LOWER |
| NEWPORT SPORT 4 | MARINE AREA 3 | MARINE AREA 1 |
| PORT ANGELES -OUTER | MARINE AREA 4 | MARINE AREA 2 |
| SEKIU | SIMILKAMEEN R 490325 | WELLS DAM- CHIEF JOE |

Table J.2. Survival Comparisons between DIT Groups by brood year.

| Brood <br> Year | Mark plus <br> CWT |  | Age 2 |  |  |  |  |  | Age 3 | Age 4 | Age 5 | Age 6 | Total |
| :---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ad-Clipped | 0.0004 | 0.0014 | 0.0006 | 0.0003 | 0.0000 | 0.0027 |  |  |  |  |  |  |
|  | No Mark | 0.0004 | 0.0014 | 0.0007 | 0.0002 | 0.0000 | 0.0027 |  |  |  |  |  |  |
| 2010 | Ad-Clipped | 0.0009 | 0.0033 | 0.0052 | 0.0006 | 0.0000 | 0.0100 |  |  |  |  |  |  |
|  | No Mark | 0.0009 | 0.0035 | 0.0050 | 0.0006 | 0.0000 | 0.0100 |  |  |  |  |  |  |
| 2011 | Ad-Clipped | 0.0003 | 0.0021 | 0.0024 | 0.0004 | 0.0000 | 0.0051 |  |  |  |  |  |  |
|  | No Mark | 0.0003 | 0.0023 | 0.0026 | 0.0003 | 0.0000 | 0.0055 |  |  |  |  |  |  |
| Mean | Ad-Clipped | 0.0005 | 0.0023 | 0.0027 | 0.0004 | 0.0000 | 0.0059 |  |  |  |  |  |  |
|  | No Mark | 0.0005 | 0.0024 | 0.0028 | 0.0004 | 0.0000 | 0.0061 |  |  |  |  |  |  |

Table J.3. Gender Composition of DIT Groups by brood year.

|  | Males |  | Females |  |
| :---: | :---: | :---: | :---: | :---: |
| Brood Year | Ad-CWT | CWT Only | Ad-CWT | CWT Only |
| 2009 | 0.720 | 0.720 | 0.280 | 0.280 |
| 2010 | 0.540 | 0.550 | 0.460 | 0.450 |
| 2011 | 0.491 | 0.521 | 0.473 | 0.527 |
| Mean | $\mathbf{0 . 5 8 4}$ | $\mathbf{0 . 5 9 7}$ | $\mathbf{0 . 4 0 4}$ | $\mathbf{0 . 4 1 9}$ |

Table J.4. Age Composition of DIT Groups by brood year.

| Brood <br> Year | DIT Group | Age Composition (Genders Combined) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ad-CWT | 1,648 | 0.137 | 0.520 | 0.244 | 0.099 | 0.000 |
|  | CWT Only | 2,787 | 0.145 | 0.526 | 0.242 | 0.088 | 0.000 |
| 2010 | Ad-CWT | 6,017 | 0.086 | 0.334 | 0.522 | 0.057 | 0.001 |
|  | CWT Only | 11,087 | 0.089 | 0.346 | 0.504 | 0.060 | 0.001 |
| 2011 | Ad-CWT | 6,329 | 0.054 | 0.407 | 0.470 | 0.070 | 0.000 |
|  | CWT Only | 3,016 | 0.057 | 0.413 | 0.474 | 0.056 | 0.000 |
| Mean | Ad-Clipped | $\mathbf{4 , 6 6 5}$ | $\mathbf{0 . 0 9 2}$ | $\mathbf{0 . 4 2 0}$ | $\mathbf{0 . 4 1 2}$ | $\mathbf{0 . 0 7 5}$ | $\mathbf{0 . 0 0 0}$ |
|  | No Mark | $\mathbf{5 , 6 3 0}$ | $\mathbf{0 . 0 9 7}$ | $\mathbf{0 . 4 2 8}$ | $\mathbf{0 . 4 0 7}$ | $\mathbf{0 . 0 6 8}$ | $\mathbf{0 . 0 0 0}$ |

Table J. 5 Size at age for DIT Groups by brood year.

| Brood Year | DIT Group | Fall Chinook fork length (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | Age-2 |  |  | Age-3 |  |  | Age-4 |  |  | Age-5 |  |  | Age-6 |  |  |
|  |  | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD | N | Mean | SD |
| 2009 | AD-CWT | 226 | 49 | 4 | 857 | 67 | 5 | 402 | 78 | 5 | 163 | 85 | 5 | 0 |  |  |
|  | CWT Only | 404 | 48 | 4 | 1,465 | 66 | 5 | 674 | 77 | 6 | 244 | 84 | 6 | 0 |  |  |
| 2010 | AD-CWT | 519 | 48 | 4 | 2,011 | 68 | 4 | 3,138 | 77 | 5 | 340 | 81 | 5 | 9 | 89 | 5 |
|  | CWT Only | 985 | 48 | 4 | 3,840 | 68 | 5 | 5,585 | 77 | 5 | 663 | 82 | 5 | 14 | 81 | 6 |
| 2011 | AD-CWT | 162 | 47 | 4 | 1,227 | 66 | 5 | 1,417 | 76 | 5 | 210 | 82 | 6 | 0 |  |  |
|  | CWT Only | 188 | 47 | 4 | 1,369 | 66 | 5 | 1,571 | 77 | 5 | 185 | 82 | 6 | 0 |  |  |
| Mean | AD-CWT | 302 | 48 | 4 | 1,365 | 67 | 5 | 1,652 | 77 | 5 | 238 | 83 | 5 | 9 | 89 | 5 |
|  | CWT Only | 526 | 48 | 4 | 2,225 | 67 | 5 | 2,610 | 77 | 5 | 364 | 83 | 6 | 14 | 81 | 6 |

## Appendix K <br> Alternative pNOB and PNI Estimates

An alternative pNOB was developed to account for the genetic influence on pNOB resulting from the PRH spawning protocol of spawning one male with one, two, or four females. It is intended to represent actual gene flow to the progeny instead of strictly the origin and number of parents. However, it should be noted that although PNI was intended to index gene flow, the alternative method of estimating pNOB as described below has not been used elsewhere and is currently undergoing review. The PNI calculation for the alternative pNOB method is PNI = Alt pNOB/(Alt pNOB + pHOS)
The alternative pNOB is calculated by assigning scores to the estimated matings of males and females based on origin during the spawning of the PRH broodstock.

The hatchery x hatchery matings $=0.0$ points,
Hatchery x natural matings $=0.5$ points, and
Natural x natural matings $=1.0$ points.
The scores of all of the matings were averaged to generate the overall alternative pNOB. For example, the alternative pNOB calculation for the mating of one natural origin male x two hatchery origin females is $(0.5+0.5) / 2$ females $)=0.5$, whereas the conventional pNOB calculation for this mating equals ( 1 natural $/(1$ natural +2 hatchery) $=0.33$.

The origin assignments of fish spawned were based on a combination of otolith marks, adipose clips, and CWTs (CWT0, as done for the conventional pNOB calculation previously discussed. The fish from the OLAFT and ABC were spawned with either fish from those collections or adipose intact broodstock fish from the PRH volunteer trap to try to reduce the number of matings with hatchery origin fish; hence, improving the pNOB for the Grant PUD program.

Similar to that done for estimates of pNOB by program, alternative pNOB and PNI estimates are given for the PRH facility as a whole and specific to the Grant PUD production associated with each brood year. The pHOS used for these estimates are given in Table 55.
The conventional and alternative pNOB values for Grant PUD production spawned at PRH and Grant PUD associated pHOS are presented in Table 1. Both methods of calculating PNI associated with the Grant PUD production provide PNI values in excess of the stated PNI target of 0.67 for most years.

Table K. 1 Conventional and alternative calculations of pNOB and PNI associated with the production specific to Grant County PUD, Return Years 2012 2016

| Conventional pNOB = pNOB/(NOB + HOB) |  |  |  |
| :---: | :---: | :---: | :---: |
|  | GCPUD Broodstock Combined | GCPUD pHOS ${ }^{\mathbf{1}}$ | PNI |
| 2012 | 0.182 | 0.068 | 0.729 |
| 2013 | 0.225 | 0.151 | 0.598 |
| 2014 | 0.343 | 0.039 | 0.898 |
| 2015 | 0.313 | 0.057 | 0.846 |
| 2016 | 0.260 | 0.049 | 0.841 |
| Return Year | Alternative pNOB = Total Score / Total Matings |  |  |
| 2012 | GCPUD Broodstock | GCPUD pHOS $\mathbf{1}^{\mathbf{1}}$ | PNI |
| 2013 | 0.197 | 0.068 | 0.744 |
| 2014 | 0.284 | 0.151 | 0.653 |
| 2015 | 0.423 | 0.039 | 0.916 |
| 2016 | 0.434 | 0.057 | 0.884 |

[^6]
## Appendix L

## Explanation of methods for calculating adult-to-adult expansions based on coded-wire tag recoveries at Priest Rapids Hatchery

Expanding adult coded wire tag (CWT) recoveries of either PRH or RSH origin fish by the corresponding brood's juvenile CWT rates has historically resulted in an under estimate of adult returns to locations within the Hanford Reach for each brood. Over the last fifteen years juvenile code-wire tag rates ranged from roughly $3 \%$ to $25 \%$ for PRH and roughly $6 \%$ for RSH. For many years, WDFW fish management staff has used adult-to-adult CWT expansions for the PRH origin returns to PRH for run-reconstruction associated with their annual fall Chinook salmon forecast. We used similar methods to expand PRH and RSH origin adult CWT recoveries in the vicinity of Hanford Reach to calculate PNI. An example of the calculations for the adult-to-adult expansion for the 2010 brood during return year 2014 is provided below. We make the assumption that the total number of PRH origin returns to PRH can be determined by removing other hatchery fish from the return: this is done by expanding the few other hatchery CWT recoveries by their corresponding juvenile CWT rates. Other hatchery CWT groups often have tag rates exceeding $50 \%$; therefore, we assume juvenile tag rate expansions are representative for these groups. In addition, we make the assumption that very few natural origin fish return to PRH.

| Adult-to-Adult Expansion BY2010 | $=\quad$ |
| :--- | :--- |
|  | $\underline{\text { Total }_{\text {BY2010 }} \text { CWT Recoveries at PRH }^{\text {Cot }} \text { PRH Origin Returns to PRH }}$ |

$$
\text { Adult-to-Adult Expansion BY2010 } \quad=\quad \underline{8,719}=0.211
$$

We then use the Adult-to-Adult Expansion by2010 to expand all recoveries of $\mathrm{PRH}_{\text {BY2010 }}$ in the Hanford Reach stream survey for return year 2014. This method is duplicated for each brood present in the given return year for both PRH and RSH to determine the total number of PRH and RSH origin fish in the escapement. The estimated number of PRH origin fish in the RY2014 Hanford Reach escapement based on the adult-to-adult expansion is higher than the number calculated using the conventional juvenile tag rate (Table 1).

Table L. $1 \quad$ The number of PRH origin fish in the RY 2014 Hanford Reach escapement calculated form Adult-to-Adult Expansions versus Juvenile Tag Rates.

| BY | CWT Recovered | Adult-to- <br> Adult Exp | Expanded <br> CWT | Survey Sample <br> Rate | Total PRH origin in <br> Escapement |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 5 | 0.216 | 23 | 0.1063 | 218 |
| 2010 | 139 | 0.211 | 659 | 0.1063 | 6,197 |
| 2011 | 18 | 0.127 | 142 | 0.1063 | 1,333 |
| 2012 | 5 | 0.160 | 31 | 0.019 | 1,645 |
| Adult-to-Adult Exp estimate for PRH origin fish in the Hanford Reach Escapement |  |  |  |  |  |


[^0]:    ${ }^{1}$ There were 157 males and 311 females taken directly from the trap and spawned. These fish are not included in the total fish ponded.

[^1]:    ${ }^{1}$ Origin determined from "in-sample" otoliths, adipose clips and/or coded-wire tags.
    ${ }^{\text {a }}$ One age-6 female assigned to natural origin based on the absence of marks or tags. The 2006 brood year was not otolith marked.

[^2]:    ${ }^{1}$ It is assumed for this analysis that all fish not possessing an otolith mark, ad-clipped or hatchery origin coded-wire tag were natural origin.

[^3]:    ${ }^{1}$ Escapement includes adults and jacks

[^4]:    ${ }^{1}$ Origin based on the presence of otoliths marks, hatchery coded-wire tags, and adipose clips present in the sub-sample.
    ${ }^{2} \mathrm{~N}$ equals the number fish included in the demographic sample for a specific brood year. Sample rates varied between return
    years; therefore the age composition is based on pooled sample data expanded for total returns by year.
    ${ }^{\text {a }}$ Does not include age- 6 returns

[^5]:    ${ }^{a}$ Marked fish were released near shore or in mid Channel in roughly equal proportions. Lengths were calculated from linear
    regression equation for fork length versus known POHL
    ${ }^{\mathrm{b}}$ Marked fish were released over the top of known redd locations.

[^6]:    ${ }^{1}$ The proportion of the pHOS specific to the GCPUD mitigation smolt releases from PRH

